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8.

River Glen:  
River Channel Assessment - Annex B

An Analysis of Invertebrate Records for The River Glen, 1976 - 1991

September 1992



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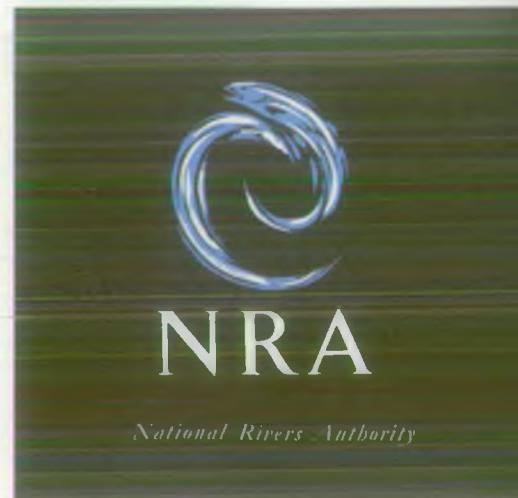
ANGLIAN REGION

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OI/447/4/A



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River Channel Assessment - Annex B**

**An Analysis of Invertebrate Records for The River Glen, 1976 - 1991**

**September 1992**

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**ANNEX B**

**AN ANALYSIS OF  
INVERTEBRATE RECORDS FOR THE RIVER GLEN, 1976-1991**

**Undertaken for the National Rivers Authority, Anglian Region**

**By**

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**September 1992**

## TABLE OF CONTENTS

|  |           |
|--|-----------|
| <b>1 SUMMARY .....</b>   | <b>1</b>  |
| 1.1 The data set.....  | 1         |
| 1.2 Reach and site level differences in invertebrate fauna.....          | 1         |
| 1.3 Biological quality .....   | 1         |
| 1.4 Temporal changes .....   | 1         |
| 1.5 Within-site temporal changes.....                                    | 2         |
| 1.6 Seasonal changes.....  | 2         |
| 1.7 Changes in invertebrate fauna with changing biological quality ..... | 2         |
| 1.8 Recommendations .....  | 2         |
| <b>2 INTRODUCTION .....</b>  | <b>3</b>  |
| <b>3 THE DATA SET .....</b>  | <b>4</b>  |
| 3.1 Sites .....  | 4         |
| 3.2 Dates .....  | 4         |
| 3.3 Taxonomic lists .....  | 4         |
| <b>4 METHODS .....</b>   | <b>8</b>  |
| 4.1 Sampling.....  | 8         |
| 4.2 Standardisation of taxa lists .....                                  | 8         |
| 4.3 Biotic indices.....  | 8         |
| 4.4 Sample classification using TWINSPAN.....                            | 8         |
| 4.5 Sample ordination using CANOCO.....                                  | 9         |
| <b>5 ANALYSIS AND DISCUSSION .....</b>                                   | <b>10</b> |
| 5.1 Taxonomic diversity.....   | 10        |
| 5.2 Biological quality .....   | 10        |
| 5.3 Ordination of all samples based upon the family data .....           | 10        |
| 5.4 Classification of all samples based upon the family data.....        | 14        |
| 5.4.1 The first major division.....                                      | 14        |
| 5.4.2 Subdivision of the main and East Glen sites.....                   | 14        |
| 5.4.3 Subdivision of the West Glen and Tham sites.....                   | 18        |
| 5.4.4 Definition of reach communities .....                              | 18        |
| 5.5 Long-term changes in biological quality .....                        | 21        |
| 5.5.1 Changes in scores at the catchment scale.....                      | 21        |
| 5.5.2 Changes in scores at the site and reach scale.....                 | 21        |
| 5.6 Site case studies.....   | 24        |
| 5.6.1 West Glen, upstream of Essendine .....                             | 24        |
| 5.6.2 West Glen at Banthorpe Lodge.....                                  | 26        |
| 5.6.3 River Glen at Kates Bridge .....                                   | 28        |
| 5.6.4 River Glen at Surfleet Seas End.....                               | 30        |
| 5.6.5 East Glen at Braceborough.....                                     | 32        |
| 5.7 Seasonality .....  | 34        |
| 5.8 Performance of biotic indices .....                                  | 38        |
| 5.8.1 Variation within LQI classes.....                                  | 38        |
| 5.8.2 Classification and Ordination of samples within LQI classes .....  | 38        |
| 5.8.3 Implications of within LQI class differences.....                  | 47        |
| <b>6 CONCLUSIONS .....</b>   | <b>50</b> |
| 6.1 Biological records .....   | 50        |
| 6.2 Site and reach level differences in Invertebrate communities.....    | 50        |
| 6.3 Biological scores.....   | 51        |
| 6.4 Long-term changes.....   | 51        |
| 6.5 Seasonal differences.....  | 52        |
| <b>7 RECOMMENDATIONS.....</b>  | <b>53</b> |
| 7.1 Preliminary reach assessment .....                                   | 53        |
| 7.2 Selection of routine monitoring sites .....                          | 53        |
| 7.3 Biological monitoring - frequency .....                              | 53        |
| 7.4 Biological monitoring - field procedure .....                        | 55        |
| 7.4.1 Invertebrate sampling.....   | 55        |
| 7.4.2 Invertebrate identification .....                                  | 55        |
| 7.5 Associated routine monitoring .....                                  | 55        |
| <b>8 REFERENCES.....</b>   | <b>56</b> |
| <b>APPENDICES</b>  |           |
| APPENDIX 1 TAXA LISTS FOR ALL SAMPLES                                    |           |
| APPENDIX 2 BIOTIC SCORES FOR ALL SAMPLES                                 |           |
| APPENDIX 3 TWINSPAN PREFERENTIAL TAXA                                    |           |

## 1 SUMMARY

### 1.1 The data set

The data set comprises invertebrate records from 44 sites on the main River Glen and its tributaries. These can be grouped into a series of physically defined reaches. Sites and reaches are shown in Figure B1. Site codes are explained in Table B1.

Records for each site vary in frequency and distribution throughout the 1976-91 period. Taxonomic level of identification varies between records, therefore the data used in the analyses have been reduced to presence / absence family level.

### 1.2 Reach and site level differences in invertebrate fauna

TWINSPAN classification (Figures B5 and B6) and ordination using correspondence analysis (Figure B4) revealed distinct reach and site level differences in fauna which can be related to a variety of water quality, hydrological, hydraulic and other environmental variables. Figure B7 summarises the classification of sites (emboldened site codes indicate the majority of samples from this site falling in this group; plain text indicates just two or more samples). "Key taxa" are those showing strong preference for the group indicated. Classification and ordination first separated the Tham and middle West Glen sites from the rest, on the basis of their rich fauna including many taxa associated with moderate to fast flows and stable substrates. Subsequently the East Glen sites, Lower Glen/Bourne Eau sites and a group of "intermediate" sites were distinguished; the first having a very impoverished fauna (related to intermittent flow and poor water quality), the second having a fauna typical of slow-flowing, weedy habitats, and the third having an intermediate fauna, sharing taxa found in the other two. Communities associated with three major reach-types: "Upland", "lowland" and "intermediate" were identified and described are in Table B3.

### 1.3 Biological quality

Biological quality varied between sites in a way consistent with the divisions described above. The highest scoring sites were around the lower Tham and the West Glen below the Tham confluence, reflecting the diverse fauna of these sites and particularly the predominance of high-scoring insect taxa characteristic of these high-quality habitats. High scores were also obtained from the lower Glen sites around the Bourne Eau confluence - water and habitat quality is also high, but due to physical habitat differences the actual taxa present are very different from the Tham / West Glen. The interrelated factors: small size, intermittent flow and poor water quality are assumed to be responsible for the poor scores in the East and upper West Glens.

### 1.4 Temporal changes

The overall trend in biological scores during the period 1976-1991 was upwards, the majority of sites showing a steep rise in all scores from 1976 to the mid to late 1980's, followed by a slight decline in 1991. Patterns of change did vary between sites, as illustrated in Figure B10 which shows sites showing significant trends in scores. Notably the lower main Glen sites displayed constant improvement in scores. "intermediate sites" showed a slight decline in 1990-91, and some East Glen sites steadily declined.

### 1.5 Within-site temporal changes

Detailed analysis of a number of sites (section 5.6) whilst confirming the temporal trends mentioned above revealed different faunal changes associated with temporal changes in biological quality. responses to detrimental changes in habitat quality are complex, sometimes involving a reversion to a poorer fauna similar to that many years earlier, while in other cases simply a reduced version of the fauna of the previous year.

### 1.6 Seasonal changes

Evidence was present for seasonal differences in the frequency of individual taxa and biotic scores, particularly ASPT. The greater seasonality in the life-histories of higher-scoring insect taxa may account for the changes in ASPT. Variability in the frequency of sampling in each month prevent detailed investigation of seasonal changes in this data set.

### 1.7 Changes in invertebrate fauna with changing biological quality

Diversity within the higher LQI classes is high (section 5.8), with reach and site scale differences as clear within LOI classes as within the whole data set. The clear implication is that site and reach specific differences other than water quality are vital in determining the fauna. This was confirmed by examining changes in fauna associated with changes from LQI class B and C to class A-A++. The "improved" fauna associated with a change from the lower to higher class was almost invariably predictable and dependent on the initial fauna, which in turn was related to specific reaches. Table B4 summarises the dominant within-class reach groups and the taxa associated with changes from the lower to higher class.

### 1.8 Recommendations

Recommendations are made towards an improved system of biological monitoring, involving an initial detailed survey of a river system in order to determine reach-communities, followed by coordinated invertebrate, macrophyte and hydrological monitoring. This strategy aims at enabling the magnitude, nature and causes of habitat changes to be determined and related to short and long term changes in invertebrate communities.

## **2 INTRODUCTION**

The River Glen, incorporating the Rivers East and West Glen and a lower section below their confluence downstream to the confluence with the River Welland, supports a wide variety of invertebrate habitat types from fast flowing, gravel-bed reaches to slow flowing, fenland drains.

Monitoring of the invertebrate communities, both routine and in response to pollution events, has been regularly practiced on the Glen since 1976. During this time the frequency of sampling and level of invertebrate identification has varied considerably, but leading up to the 1990-91 situation where sampling at most sites was carried out at least twice a year and a substantial proportion of the fauna is identified to species level.

Despite its inconsistencies, the data set for the 1976-91 period contains a substantial amount of information, comprising some 250 records from 44 sites distributed throughout the River Glen system.

This report aims to assess:

- the nature of the invertebrate communities in different parts of the catchment;
- long-term changes in biological quality at sites with long records;
- the performance of indices of biological quality in both monitoring within-site change and comparing sites within the river system.

### 3 THE DATA SET

#### 3.1 Sites

The distribution of sites on the East, West and main River Glen is shown in Figure B1. Sediment analysis of the West Glen and main Glen allow the river to be divided into a number of physically distinct reaches: The upper West Glen (reach A); the West Glen around Essendine (reach B); the West Glen around Banthorpe Lodge (reach C); the main Glen from the confluence of the East and West Glens to the confluence with Bourne Eau (reach D); the East Glen (E - lack of physical information prevents this from being further subdivided); the lower main Glen downstream of the confluence with Bourne Eau (G); Bourne Eau (B); and the River Tham (T).

The codes given to sites are listed in Table B1. Where a particular sample is referred to later in the report, the site code is followed by the month and year. In the one instance where two samples from the same site were taken in the same month, the full date follows the site code.

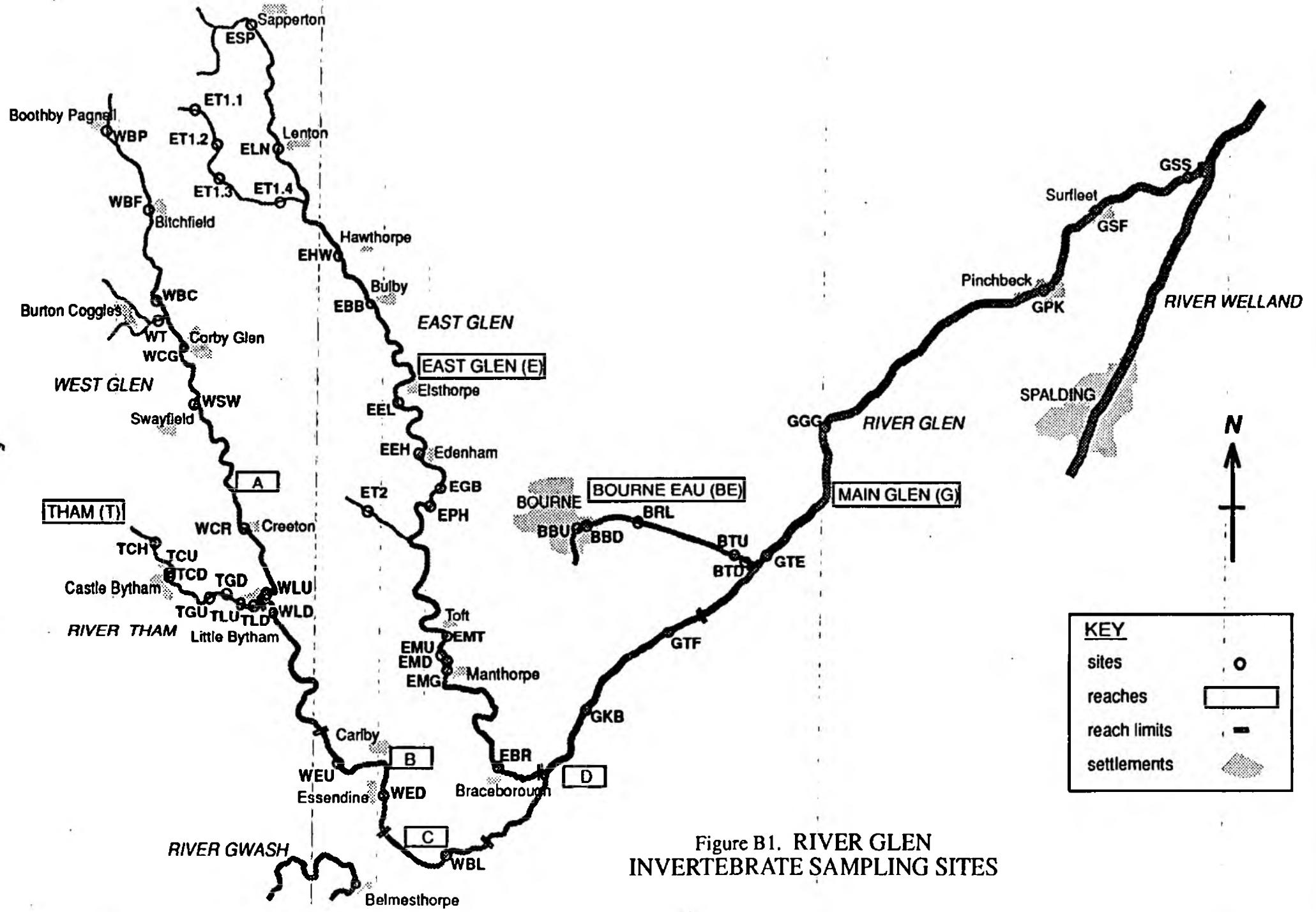
#### 3.2 Dates

Records exist for most years in the 1976-91 period, with the exception of a gap from 1980 to 1981. The year 1988 has a particularly good site coverage due to a special survey being carried out (Extence, 1989), the results of which have been added to the NRA routine sampling data. 1990 has good seasonal coverage of the routine monitoring sites, and it is understood that current practice is to sample in spring, summer and autumn. Table B2 shows the years in which each site was sampled.

Although a few sites have regular records for the entire period (eg. WLD, GSS), the majority have a patchy record, some sampled no more recently than the early 1980's (eg. GGG), and some only in recent years (eg. ELN).

#### 3.3 Taxonomic lists

Each record consists of a list of taxa found at the site, occasionally with an indication of abundance. Taxa recorded, and the level of identification, are undoubtedly influenced by the format of the recording sheet. In general the level of identification expected has tended towards the species level in later years; the earliest recording sheets are mainly to family level, ie. sufficient to allow calculation of the standard biological indices.



**Table B1. River Glen site codes and locations**

| SITE CODE | N.G.R.    | LOCATION  |
|-----------|-----------|---|
| ELN       | TF 022302 | E. Glen at Lenton   |
| EHW       | TF 038272 | E. Glen at Hawthorpe Rd.  |
| EBB       | TF 047260 | E. Glen at Bulby - between fd. and br.                                |
| EEL       | TF 054233 | E. Glen at Elsthorpe  |
| EEH       | TF 061219 | E. Glen at Edenham d/s A151 rd.br.                                    |
| EGB       | TF 067209 | E. Glen at Gravel Bridge, d/s rd.br.                                  |
| EPH       | TF 064205 | E. Glen at Pasture Hill Fm. d/s br.                                   |
| EMT       | TF 068169 | E. Glen at Manthorpe Toft, u/s STW                                    |
| EMU       | TF 066164 | E. Glen at Manthorpe, u/s STW outfall                                 |
| EMD       | TF 067163 | E. Glen at Manthorpe, d/s STW outfall                                 |
| EMG       | TF 068160 | E. Glen at Manthorpe G.S., d/s weir                                   |
| EBR       | TF 082135 | E. Glen at Braceborough, u/s & d/s rd. br.                            |
| ET11      | SK 998313 | Trib.of E. Glen   |
| ET14      | TF 022288 | Trib.of E. Glen   |
| WBP       | SK 973307 | W. Glen at Boothby Pagnell  |
| WBF       | SK 985286 | W. Glen at Bitchfield u/s br., d/s pipe                               |
| WBC       | SK 987261 | W. Glen, B1176, Burton Coggles G.S.                                   |
| WCG       | SK 995248 | W. Glen at Corby Glen u/s rd. br.                                     |
| WSW       | SK 998233 | W. Glen at Swayfield Rd., d/s Corby Glen STW                          |
| WCR       | TF 011199 | W. Glen at Creton   |
| WLU       | TF 019177 | W. Glen at Little Bytham, u/s confl.,d/s fd. & ft. br.                |
| WLD       | TF 019177 | W. Glen at Little Bytham, d/s confl.,d/s rd.br.                       |
| WEU       | TF 038137 | W. Glen at Essendine u/s & d/s Carby-Essendine rd.br.                 |
| WED       | TF 050127 | W. Glen at Essendine  |
| WBL       | TF 068112 | Glen at Banthorpe Lodge, d/s br.                                      |
| GKB       | TF 107148 | Glen at Kates Bridge G.S., d/s weir & br.                             |
| GTF       | TF 129170 | Glen at Thurlby Fen, end of Fen rd.                                   |
| GTE       | TF 156190 | Glen at Tongue end, d/s confl. with Bourne Eau                        |
| GGG       | TF 173225 | Glen at Guthram Gowt  |
| GPK       | TF 235260 | Glen at Pinchbeck, u/s br. by sluice                                  |
| GSF       | TF 251281 | Glen at Surfleet, A16   |
| GSS       | TF 277291 | Glen at Surfleet Seas End, d/s Blue Gowt Drain                        |
| TCH       | SK 986196 | Tham at Cottage Hill, u/s bridge, d/s track                           |
| TCU       | SK 992187 | Tham at Castle Bytham, d/s ft.br. to motte & bailey, u/s village pond |
| TCD       | SK 993183 | Tham at Castle Bytham, d/s fd. in village, d/s pond                   |
| TGU       | TF 003181 | Tham at Glebe Fm., u/s STW  |
| TGD       | TF 008181 | Tham, d/s STW   |
| TLU       | TF 012178 | Tham at Little Bytham   |
| TLD       | TF 015179 | Tham at Little Bytham   |
| BBU       | TF 102198 | Bourne Eau at Bourne, u/s confl. Car Dyke; d/s rd. br.                |
| BBD       | TF 105198 | Bourne Eau at Bourne, d/s rd.br., sluice +weir                        |
| BRL       | TF 121199 | Bourne Eau, d/s railway br., d/s pylons                               |
| BTU       | TF 148190 | Bourne Eau at Tongue End, end of farm track                           |
| BTD       | TF 153188 | Bourne Eau at Tongue End, u/s Glen confl.                             |

|                               |                      |                    |
|-------------------------------|----------------------|--------------------|
| abbreviations: u/s = upstream | rd.br. = road bridge | conf. = confluence |
| d/s = downstream              | ft.br. = foot bridge |                    |
| br. = bridge                  | fd. = ford           |                    |

Table B2. Dates of samples from River Glen sites

| SITE  | 1976 | 1977 | 1978 | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 |   |
|-------|------|------|------|------|------|------|------|------|------|------|------|------|------|------|------|------|---|
| ELN   |      |      |      |      |      |      |      |      |      |      |      |      |      | 1    | 2    | 2    | 1 |
| EHW   |      |      |      |      |      |      |      |      |      |      |      |      |      | 1    |      |      |   |
| EBB   | 1    | 1    | 1    |      |      |      | 1    | 2    |      |      |      |      | 1    |      |      |      |   |
| EEL   |      |      |      |      |      |      |      |      |      |      |      |      | 1    |      |      |      |   |
| EEH   | 1    |      | 1    | 1    |      |      |      | 1    | 1    | 1    |      | 1    | 1    | 2    | 3    | 1    |   |
| EGB   |      |      |      |      |      |      |      | 1    |      |      |      |      |      | 1    | 1    |      |   |
| EPH   | 1    |      |      |      |      |      | 1    |      |      |      |      |      | 1    |      |      |      |   |
| EMT   |      |      |      |      |      |      |      |      |      |      |      |      | 1    |      |      |      |   |
| EMU   |      |      |      |      |      |      |      |      |      |      |      |      | 2    |      |      |      |   |
| EMD   |      |      |      |      |      |      |      |      |      |      |      |      | 2    |      |      |      |   |
| EMG   | 1    | 1    |      |      |      |      |      | 1    | 1    |      |      |      | 2    |      |      |      |   |
| EBR   |      |      | 1    | 1    |      |      |      | 2    | 1    | 1    | 1    | 2    | 1    | 2    | 3    | 1    |   |
| ET1.1 |      |      |      |      |      |      |      | 1    |      |      |      |      |      |      |      |      |   |
| ET1.4 |      |      |      |      |      |      |      | 1    |      |      |      |      | 1    |      |      |      |   |
| WBP   |      |      |      |      |      |      |      |      |      |      |      |      | 1    |      |      |      |   |
| WBF   |      | 1    | 2    |      |      |      |      | 1    |      |      |      |      | 1    |      |      |      |   |
| WBC   |      |      |      |      |      |      |      |      |      |      |      |      | 1    | 2    | 3    | 1    |   |
| WCG   |      | 1    | 2    |      |      |      |      | 1    | 1    |      |      |      | 1    | 1    |      |      |   |
| WSW   |      |      |      |      |      |      |      |      |      |      |      |      | 1    | 1    |      |      |   |
| WCR   |      |      |      |      |      |      |      |      |      |      |      |      | 1    |      |      |      |   |
| WLU   |      |      |      |      |      |      |      |      |      | 1    |      |      |      |      |      |      |   |
| WLD   | 1    | 1    |      | 2    |      |      |      | 1    | 1    | 2    | 1    | 1    | 2    | 2    | 3    | 1    |   |
| WEU   | 1    | 1    | 1    | 1    |      |      |      | 1    | 2    |      |      |      |      |      |      |      |   |
| WED   |      |      |      |      |      |      |      |      |      |      |      |      | 1    |      |      |      |   |
| WBL   | 1    |      | 1    | 2    |      |      |      | 2    | 1    | 1    | 1    | 1    | 1    | 2    | 2    | 1    |   |
| GKB   | 2    | 1    | 1    |      |      |      |      |      |      |      |      |      |      | 1    | 3    | 1    |   |
| GTF   | 1    | 1    | 1    | 1    |      |      |      | 1    |      | 1    | 1    | 1    |      |      |      |      |   |
| GTE   |      |      | 1    | 2    |      |      |      | 1    | 1    |      | 1    | 1    | 1    | 2    | 3    | 1    |   |
| GGG   | 2    | 1    | 1    | 1    |      |      |      | 1    | 1    |      |      |      |      |      |      |      |   |
| GPK   | 2    | 1    | 1    | 2    |      |      |      | 1    | 1    |      |      |      |      |      |      |      |   |
| GSF   |      |      |      |      |      |      |      |      |      |      |      |      |      | 2    | 3    | 1    |   |
| GSS   | 2    | 1    | 1    | 1    |      |      |      | 1    | 1    | 1    | 1    | 1    |      | 2    | 3    | 1    |   |
| TCH   |      |      |      |      |      |      |      | 1    | 1    |      |      |      | 1    |      |      |      |   |
| TCU   |      |      |      |      |      |      |      |      | 1    |      |      |      | 1    |      |      |      |   |
| TCD   |      |      |      |      |      |      |      | 1    | 1    |      |      |      | 1    |      |      |      |   |
| TGU   |      |      |      |      |      |      |      | 1    | 1    | 1    |      |      |      |      | 1    | 1    |   |
| TGD   |      |      |      |      |      |      |      |      |      | 1    |      |      |      |      |      |      |   |
| TLU   |      | 1    | 1    |      |      |      |      | 1    |      |      |      |      |      |      |      |      |   |
| TLD   | 1    |      |      |      |      |      |      | 1    |      | 1    | 1    |      | 1    | 1    | 2    | 3    | 1 |
| BBU   | 1    |      | 1    | 1    |      |      |      |      |      |      |      |      |      |      |      |      |   |
| BBD   |      | 1    |      | 1    |      |      |      | 1    | 1    |      |      |      |      |      |      |      |   |
| BRL   |      |      | 1    |      |      |      |      | 1    | 2    |      |      |      |      |      |      |      |   |
| BTU   |      |      |      |      |      |      |      |      |      |      |      |      |      | 2    |      |      |   |
| BTD   |      |      |      |      |      |      |      |      |      |      |      |      |      | 2    | 1    |      |   |

## 4 METHODS

### 4.1 Sampling

All the samples were collected using a standard 1.5-minute kick-and-sweep technique with a hand net. All micro-habitats within a site were sampled in order to maximise the range of taxa recorded, except where in very deep sites sampling was restricted to the margins. Samples were sorted "live", either in the field or laboratory, and taxa identified and recorded on a standard recording form.

### 4.2 Standardisation of taxa lists

Information from the recording forms was entered in full onto a Microsoft EXCEL spreadsheet, using separate spreadsheets for different formats of recording sheets. These were then edited together to form a composite file on which every taxon name used was included. Due to differences between recording sheets the same taxa may have been referred to under different names, e.g. *Baetis vernus* could have been recorded variously under *Baetidae*, *Other Baetidae*, *Baetis* sp., *Baetis* indet., *Baetis vernus/tenax* or *Baetis vernus*. For the purposes of comparing sites and dates it was therefore necessary to group such records to the appropriate taxonomic level: in this example, *Baetidae*. A "grouped" file was therefore produced for all sites. In most instances the groupings were to family level, and a further file was constructed which recorded all taxa to family level or higher.

An indication of abundance class was given on a minority of records, and for the purposes of standardisation ignored: presence / absence data were used in all the analyses.

### 4.3 Biotic indices

Using the standard formulae (Extence and Ferguson, 1989) the following biotic indices were calculated for each record, using the family data file:

Biological Monitoring Working Party Score (BMWP)  
number of BMWP scoring taxa  
Average Score Per Taxon (ASPT)  
Lincoln Quality Index (LQI)

Calculation of the LQI required information on the type of habitat sampled: Riffle, run or pool. Such information was provided on most record sheets; where it was absent the most likely habitat type was assumed.

### 4.4 Sample classification using TWINSPAN

Two-Way Indicator Species Analysis (Hill, 1979) was used to classify samples on the basis of the family level data into groups from different regions, sites, dates and seasons sharing similar invertebrate communities.

TWINSPAN uses a method of "refined ordination" to classify sites. A hierarchical pattern of divisions is produced, and with each division "preferential" taxa identified which are significantly more frequent in one sub-group or the other. Subsequently an "indicator ordination" is performed which identifies taxa on the basis of the presence or absence of which alone the same classification could be produced. For this reason the "indicator" taxa chosen are frequently the rarer taxa associated with each group, rather than more common taxa with large differences in frequency, even though these may be more ecologically significant. For this reason the discussion focuses on the preferential taxa of the main TWINSPAN ordination,

although indicator taxa are shown in the dendograms. In some instances "key" taxa are referred to - these are the most strongly preferential taxa whose presence or absence is believed to be particularly significant in group-separation.

It should be noted that TWINSPAN places greater weight on similarities in taxa present than on the absence of taxa, with the result that samples with very reduced communities tend to be grouped with others of which their faunas are subsets, at least in the earlier divisions. For this reason the TWINSPAN classifications should be viewed in conjunction with the correspondence analysis ordinations which are more effective at identifying sites with an absence of commoner taxa.

TWINSPAN sub-divides groups of samples repeatedly until the requested level of divisions has been reached. The lowest level of division which still represents ecologically meaningful end-groups, ie. there is an ecological basis for the differences in taxa between groups rather than chance sampling effects, has to be subjectively assessed. In making this assessment reference was made to the preferential taxa associated with the divisions, and to the correspondence analysis (Section 4.5) which is extremely useful in determining the actual amount of faunal variation between end-groups.

#### 4.5 Sample ordination using CANOCO

Ordination was performed using the correspondence analysis option of the CANOCO package (Ter Braak, 1988). Correspondence analysis allows samples to be grouped according to similarities in their taxonomic composition. As such it is a useful tool both on its own for grouping samples and for confirming the groupings produced by complementary analyses such as TWINSPAN. Correspondence analysis was used in preference to any form of detrended correspondence analysis (eg. DECORANA). In this particular (Glen) data set where major differences in groups of samples were being investigated, rather than small between-sample changes in relation to linear environmental variables, the small benefits of detrending in eliminating some mathematical errors were considered to be outweighed by the problems that exist with the detrending process which invalidate the interpretation of the distance between points in the ordination diagram.

## 5 ANALYSIS AND DISCUSSION

### 5.1 Taxonomic diversity

From the combined list of taxa, in which the grouping of taxa under a family or higher group name was kept to the minimum necessary to standardise across the 15 year period, a total of 112 types were recorded. Lists for all the samples are given in Appendix 1.

### 5.2 Biological quality

BMWP scores, ASPTs and numbers of scoring taxa for all samples are given in Appendix 2.

Mean site scores for BMWP, no. scoring taxa and ASPT are show in Figure B2. One standard deviation error bars give an indication of the range of scores at each site over the 15 year period, although these have little true significance due to variable sample size. It is apparent that the lower West Glen and the River Tham have the most diverse and sensitive communities as indicated by all three scores. High BMWP scores are a result of both the number and "quality" of taxa. Rather surprisingly the other reaches have very similar score ranges despite the diverse nature of the habitats, with the smaller E. Glen riffle sites scoring similarly to the slow flowing lower Glen and Bourne Eau reaches, possibly due to water quality problems. Bourne Eau is notable in the steady increase in scores in progressively more downstream sites.

When reorganised and plotted against distance from the confluence with the River Welland there is a predictable relationship between scores and distance (Figure B3), with the highest scores occurring in the middle regions. The top of the tributaries have reduced communities due to their intermittent nature, and the worsening effect this has on organic pollution problems, whilst the lowest sites are deep, slow flowing and have rather specialised faunas. The middle reaches show the greatest diversity, with more predictable, moderate, flows and better water quality. LQI shows improved scores in the lower reaches due to the compensation for "restricted riffles".

### 5.3 Ordination of all samples based upon the family data

The ordination diagram constructed from the results of a correspondence analysis of the family data is shown in Figure B4. Axes one and two of the ordination accounted for the majority of the variation in the data and are therefore the only two used. Two points are omitted from the diagram: two samples from the West Glen at Corby Glen which possessed very few taxa and plotted very distantly from the rest of the points. The remainder of the sites plotted within a relatively small range on axis one of the ordination, indicating the overall similarity of their faunas. However, certain between-reach differences can be seen.

There are two noticeable trends: Along axis two, the River Tham and reach A of the (upper) West Glen overlap considerably towards the top of the diagram, with the other reaches plotting lower down. The lower West Glen (reaches B, C and D) plot slightly closer to the middle than the other sites but with no clear separation. Secondly, along axis one, the sites separated from the Tham and reach A on axis two (which overlap on axis one) are further separated into the main Glen, Bourne Eau and East Glen in a left to right direction. These trends are summarised in the inset of Figure B4. The two trends seem logical: The West Glen and Tham, being physically close and probably of similar habitat type, share a similar fauna. The lower Glen and Bourne Eau separate from the other sites being of a very different physical nature, but also possess differences from each other. The East Glen is surprisingly different from the West Glen - without knowledge of the physical nature of this reach it is difficult to explain, but pollution and/or drought may be factors. Reaches B, C and D which fall in the centre of the ordination probably possess taxa common to all the other reaches.

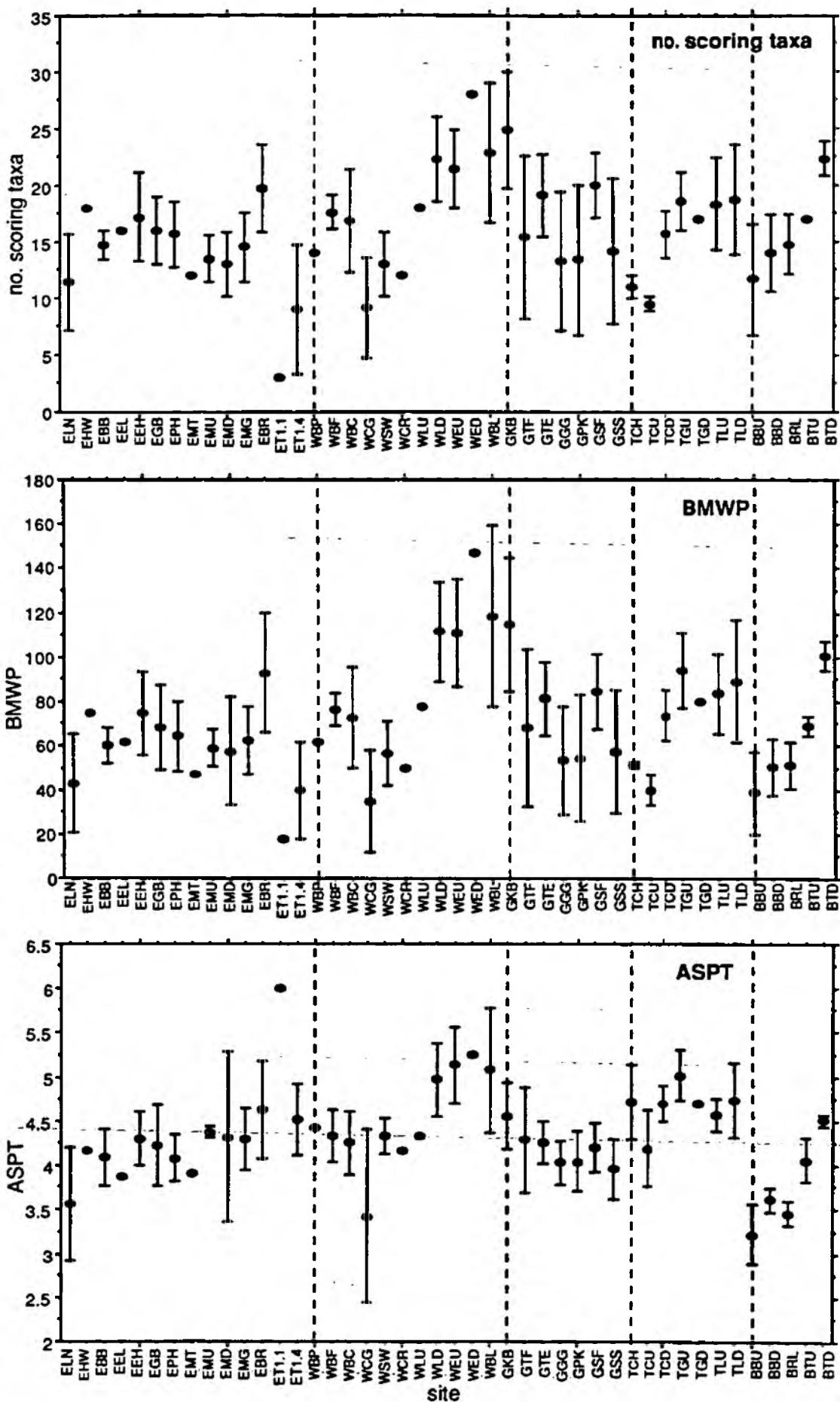


Figure B2. Mean number of scoring taxa, BMWP and ASPT for each site, with 1 S.D. error bars.

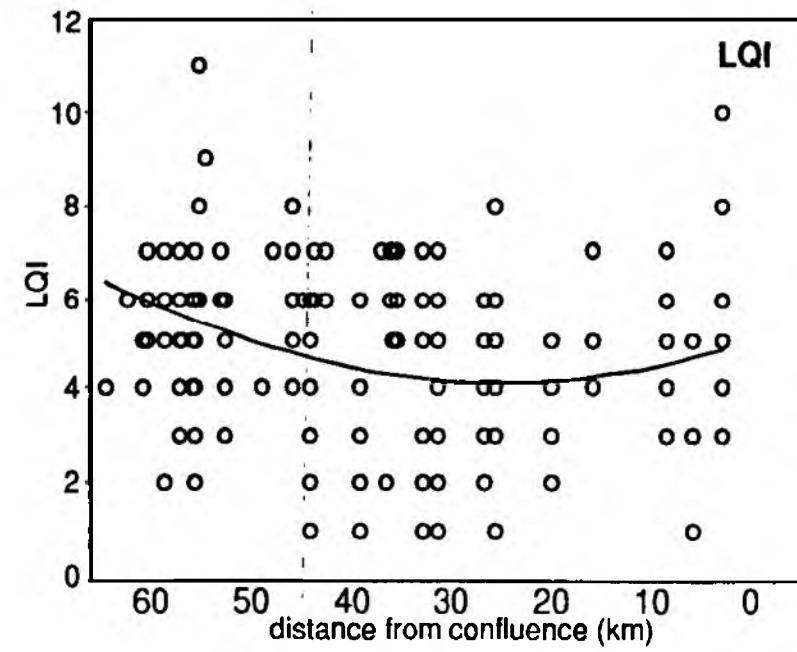
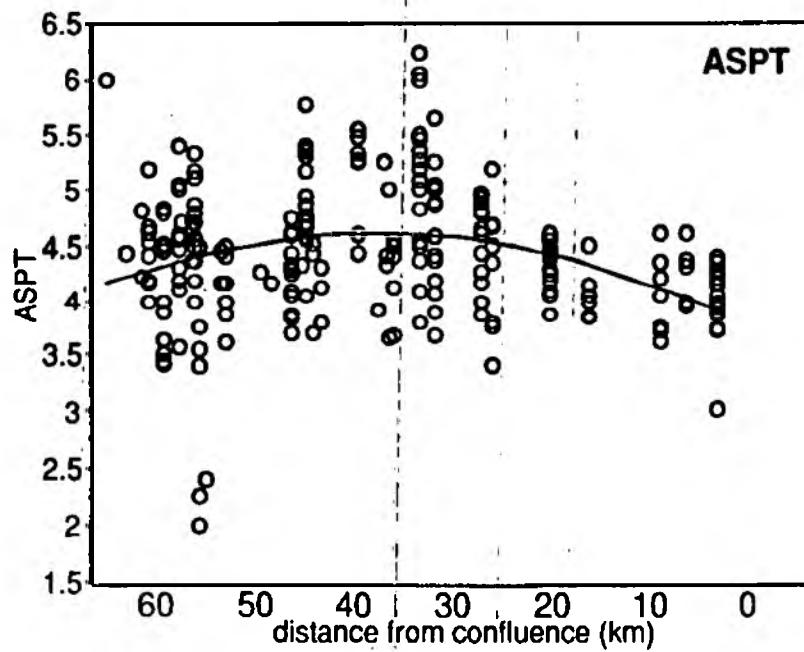
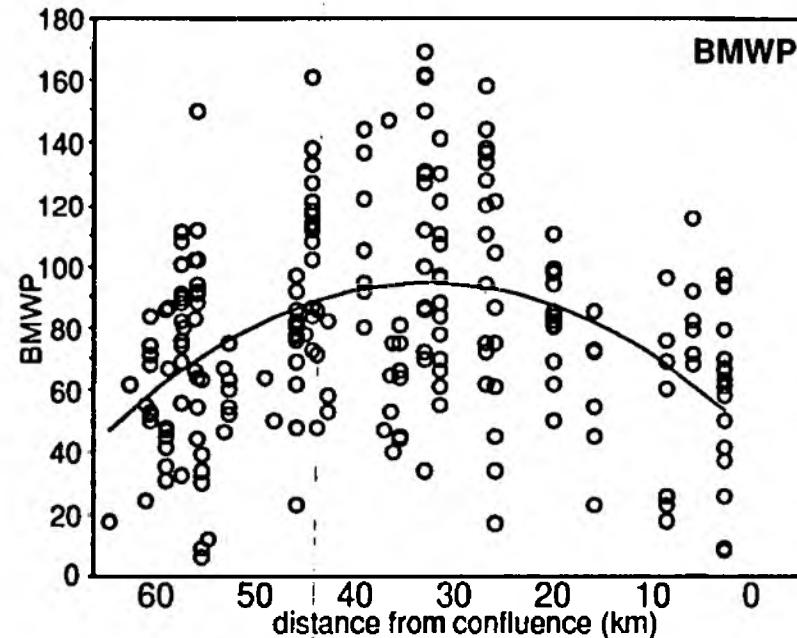
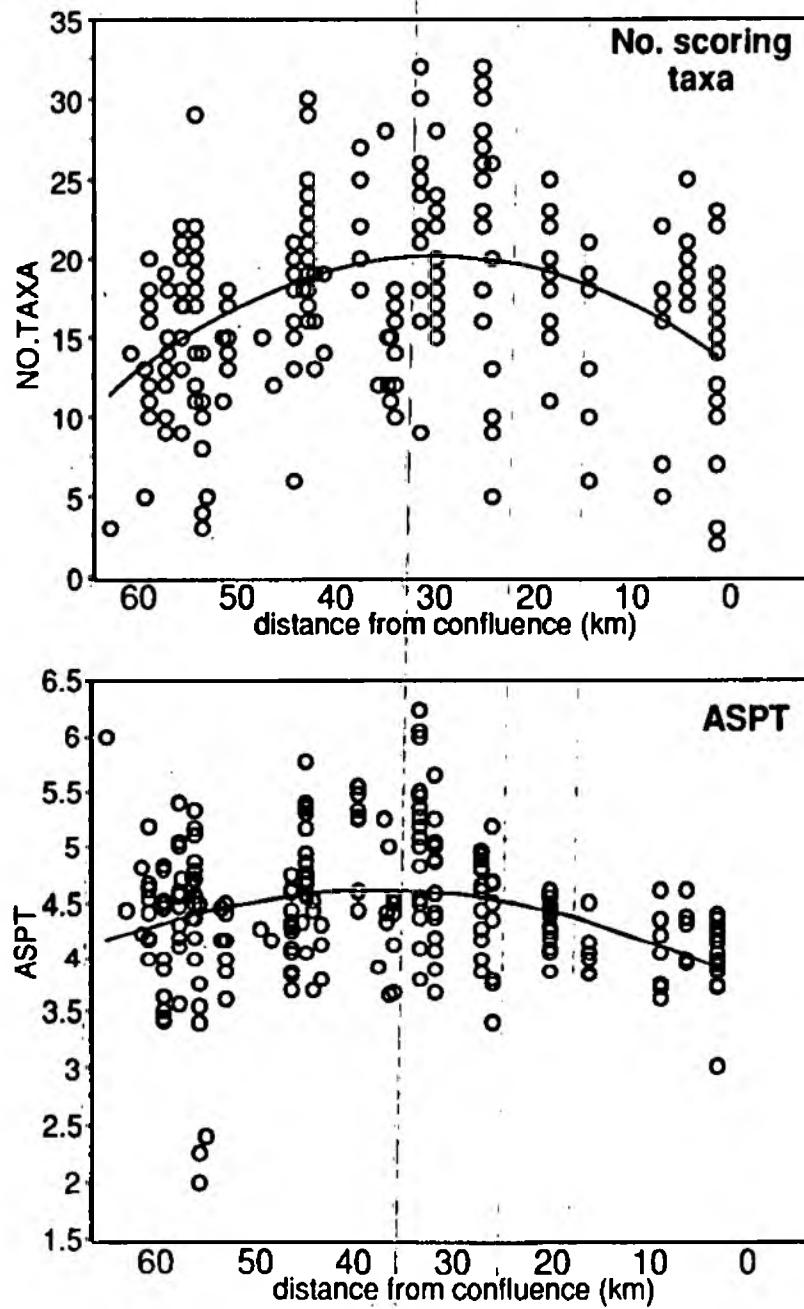
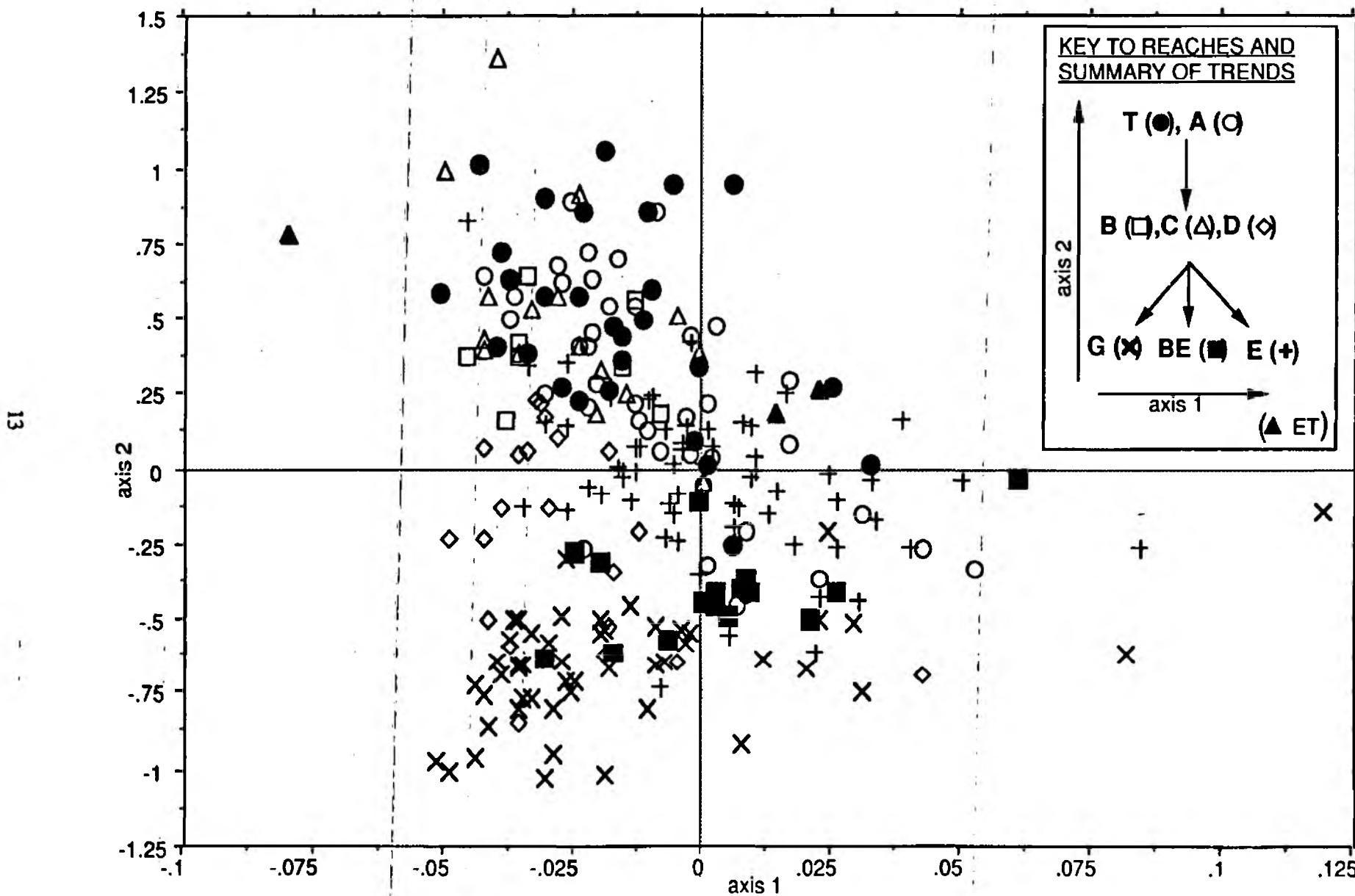


Figure B3. Number of scoring taxa, BMWP, ASPT and LQI for all samples, against distance from source



**Figure B4.** Correspondence analysis of River Glen family data

## 5.4 Classification of all samples based upon the family data

A classification of the whole data set was made using TWINSPAN. Figure B5 shows the dendrogram produced from the output, with taxa derived from the indicator ordination shown at each division. End group numbering is for convenient reference and is not the same as the TWINSPAN division numbering. Figure B6 shows the diagonal samples-and-taxa matrix with groups of sites numbered at the bottom using the same notation as in the dendrogram (Figure B5). Preferential taxa for each division are listed in Appendix 3.1. A summary of the sites, physical and biotic characteristics and key taxa associated with the eight significant divisions of the classification is illustrated in Figure B7 (emboldened site codes indicate the majority of samples from this site falling in this group; plain text indicates just two or more samples). "Key taxa" are those showing strong preference for the group indicated.

### **5.4.1 The first major division**

The first major division separates (with some exceptions) the main river Glen samples, Bourne Eau and the bulk of the East Glen samples from the West Glen and Tham samples. Some sites have samples in both groups, notably GKB and EBR. These two are physically on the borderline between the two groups and so this seems logical. Taxa associated with this division suggest that the separation is based on physical rather than water quality differences: The taxa found significantly more frequently in the "East and lower Glen" group are the Ostracoda, Cladocera, Coenagniidae and Valvatidae; while those preferring the "West Glen and Tham" group include a number of Ephemeroptera: Ephemeridae, Ephemeralidae, and Leptophlebiidae; several Trichoptera: Rhyacophilidae, Polycentropodidae, Hydropsychidae, Hydroptilidae, Limnephilidae, and Goeridae; and also the Elmidae, Tipulidae, Aculyidae, Simuliidae and Dendrocoelidae. In Figure B6 these taxa are arrayed on the top left (East and lower Glen group) and bottom right (West Glen and Tham group). It is apparent from this matrix that by far the majority of taxa occur in both groups, ie. the obvious site distinctions are based on more subtle differences in frequencies of rather than the presence or absence of large groups of taxa. This might well be expected as all sites are within one catchment with many physical similarities. Also, recolonisation of any site is likely to be from others in the same catchment, which may significantly influence faunal composition of that site.

### **5.4.2 Subdivision of the main and East Glen sites: groups 1-16**

The second TWINSPAN division separates end-groups 1-8 from 9-16, within the East and lower Glen group. Groups 1-8 encompass most of the lower main Glen samples and all but two of the Bourne Eau samples. Groups 9-16 contain most of the East Glen samples but also a mixture of others from the main and West Glen and Tham. Only one family - the Hydropsychidae - is given as preferential for the groups 1-8, while those taxa strongly preferential for the groups 9-16 include the Piscicolidae, Cladocera, Coenagniidae, Hydroptilidae, Valvatidae and Physidae. The preferential taxa probably reflect the physical difference between the smaller, faster-flowing nature of the sites in groups 9-16 as opposed to the slower-flowing lowland main Glen sites.

The subdivision of groups 1-8 into groups 1-4 and 5-8 does not appear to be site-based but identifies a group samples (in end groups 1-4) which possess a number of taxa including the Gerridae, Notonectidae and Unionidae at relatively high frequency but the Cladocera, Copepoda and Hydroptilidae less frequently. The subsequent division identifies groups 1 and 2 as being less diverse than groups 3 and 4, lacking a number of taxa including a number of Coleopteran families. Groups 5-8 are separated fairly clearly into groups 5-6 and 7-8 based upon the frequency of taxa which again suggest differences in flow conditions and/or possibly macrophyte development: groups 5 and 6, which mainly consist of samples from the lower main Glen sites, have higher frequencies of Coenagniidae, Caenidae and Corixidae; whereas groups 7 and 8, which include samples from Bourne Eau and the two sites close to the confluence of the East and main Glen - EBR and GKB - possess greater frequencies of taxa including Polycentropodidae and Simuliidae.

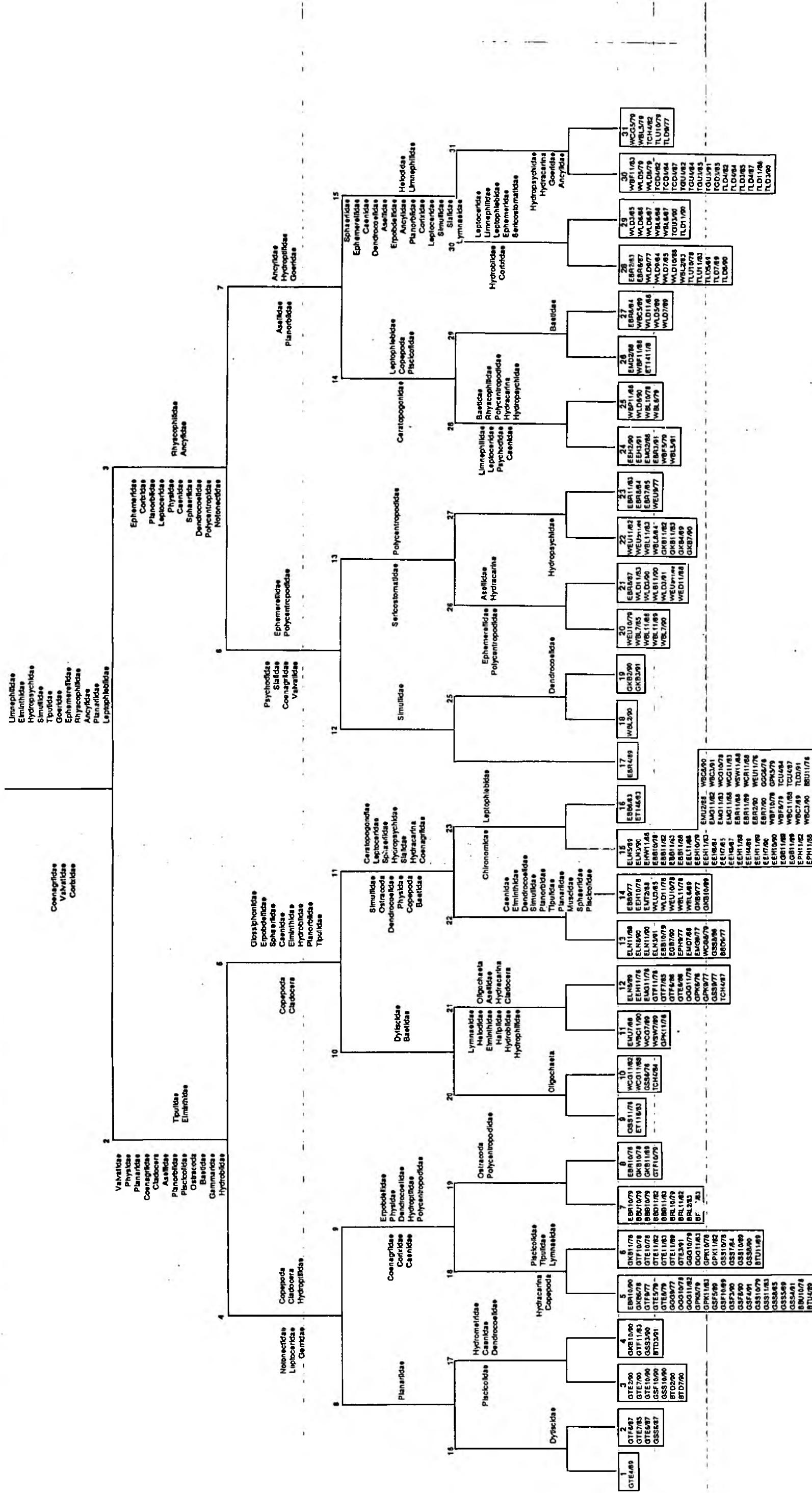
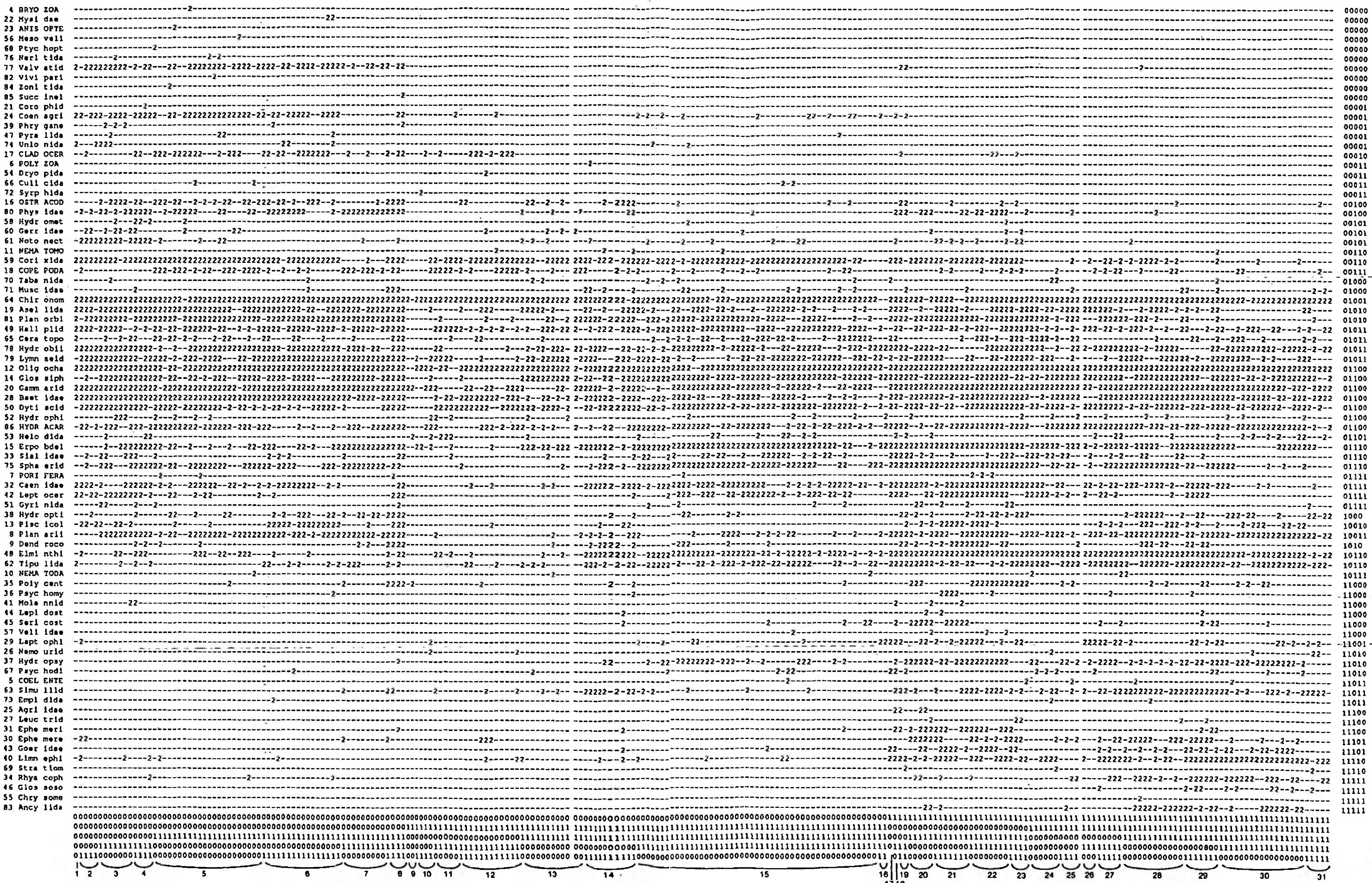


Figure B5. TWINSPAN groupings from an analysis of family data for all samples. End group numbers are referred to in the text; division numbers relate to those given in appendix 3.1.

Figure B6. TWINSPAN samples and taxa matrix. End groups numbers are as in figure 5.



| TWINSPAN<br>division   | 8  | 9  | 10  | 11   | 12   | 13   | 14   | 15  |
|------------------------|--|--|---|--|--|--|--|---|
| TWINSPAN<br>end groups | 1-4  | 5-8  | 9-12  | 13-16  | 17-19  | 20-23  | 24-27  | 28-31   |
| sites                  | BTD<br>GTE<br>GTF<br>GTE   | GTF<br>GTE<br>GGG<br>GPK<br>GSS<br>GSF<br>BBU<br>BBD<br>BRL<br>BTU<br>GKB<br>EBR | TCH<br>GTF<br>GSS<br>GPK<br>WCG                         | ELN<br>EBB<br>EEH<br>EGB<br>EPH<br>EMG<br>EBR<br>WBF<br>WBC<br>WCG<br>TCU<br>GKB<br>WEU<br>WBL   | GKB  | EBR<br>WEU<br>WBL<br>GKB<br>WLD  | WBF<br>EBR<br>EEH<br>WLD<br>WBL  | WLD<br>TCD<br>TGU<br>TLD<br>TLU<br>EBR<br>WBL                         |
| group description      | lowland, slow-flowing, abundant macrophytes, good water quality, rich fauna  | lowland, slow-flowing, macrophytes, reasonable water quality and fauna           | intermittent, or poor water quality, very reduced fauna | poor habitat, poor fauna   | good, diverse habitat, rich fauna                  | moderate flow, some macrophytes moderate water quality, intermediate fauna         | moderate flow, clean substrates, rich fauna  | moderate/fast flow, stable, clean substrates, macrophytes, rich fauna |
| key taxa               | Leptoceridae, Hydrometridae, Gerridae, Notonectidae, Unionidae               | Cladocera, Copepoda, Hydroptilidae   | Cladocera, Copepoda                                     | Glossiphoniidae, Erpobdellidae, Aeselidae, Sialidae, Caenidae, Elmimithidae, Hydrobiidae, Planorbidae  | Coenagrionidae, Valvatidae, Helodidae, Psychodidae | Piscicolidae, Copepoda, Ephemeralidae, Polycentropodidae, Psychomyiidae, Gyrinidae | Dendrocoelidae, Piscicolidae, Rhyacophilidae, Hydroptilidae, Ancyliidae, Goeridae, Glossosomatidae |   |
|                        | Piscicolidae, Cladocera, Coenagrionidae, Hydroptilidae, Valvatidae, Physidae |  |   | Ephemerellidae, Notonectidae, Physidae, Sericostomatidae, Gyrinidae  | Rhyacophilidae, Ancyliidae                         |  |  |   |
|                        | Ostracoda, Cladocera, Coenagrionidae, Valvatidae                             |  |   | Leptophlebiidae, Ephemeralidae, Ephemerellidae, Rhyacophilidae, Hydropsychidae, Limnephilidae, Goeridae, Ancyliidae, Simuliidae, Polycentropodidae |  |  |  |   |

Figure B7. Summary of site groupings based upon TWINSPAN classification of samples. Key taxa are those showing a strong preference for the group indicated.

Groups 9-16 are subdivided into 9-12 and 13-16 on the basis of a range of taxa absent or at very low frequency in the former group. This is particularly apparent in the samples - and - taxa matrix (fig.6), where many of the common taxa located in the middle of the diagram are visibly less frequent in these groups. In particular the Hydropsychidae and Leptoceridae are absent from this group altogether. Samples comprising groups 9-12 are drawn from a wide range of sites, and appear to represent occasions when the fauna at those sites has been severely impacted (drought or pollution?). Many of these samples are from the late 70's period. Further separation of these groups (9-12) appears to reflect the severity of the faunal reduction, with group 9 the worst affected. Groups 13-16 subdivide into groups 13-14: a range of samples from the upper East Glen sites, the lower East Glen sites below Manthorpe STW and a few "odd" sites from the West and main Glen; and groups 15-16: The larger East Glen sites, a significant number of upper West Glen sites and a few from the Tham and lower Glen. All of these samples appear to be from "poor quality" sites or from dates when "better" sites were under stress (eg. WEU, WLD and WBL in 1976). The taxa separating groups 13-14 from 15-16 include the Simuliidae and Physidae which are more frequent in the former group and the Coenagrionidae, Sialidae, Hydropsychidae, Leptoceridae and Sphaeriidae in the latter. It could be that the former group is the more severely impacted than the latter.

#### 5.4.3 Subdivision of the West Glen and Tham sites: groups 17-31

Within this set of groups the first subdivision into groups 17-23 and 24-31 separates the majority of the lower West Glen and "intermediate" sites (EBR and GKB) from the upper West Glen and Tham sites. The taxa associated with the division reflect this: the Rhyacophilidae and Aencylidae which prefer faster flow and more stable substrates show a strong preference for the West Glen and Tham sites, whereas taxa such as the Ephemeroidea, Polycentropodidae, Gyrinidae, Corixidae, Notonectidae and Sericostomatidae which tolerate or prefer slower flowing conditions are more frequent in the lower West and East Glens and Kates Bridge.

Within groups 17-23 the ecological basis of further subdivisions are difficult to interpret, these being based on substitutions of taxa particularly from the Trichoptera, Coleoptera and Diptera.

Groups 24-31 subdivide into groups 24-27 and 28-31 mainly on the basis of a collection of more "sensitive" taxa which are more frequent in the latter group. Sites dominating this group are from the Tham and middle West Glen, and are those which scored highest using the various biotic indices. Taxa preferring this group include the Piscicolidae, Rhyacophilidae, Hydroptilidae, Goeridae, Glossosomatidae and Aencylidae. Taxa preferring the other group (24-27) include the Asellidae, Sialidae, Leptoceridae and Planorbidae. Further subdivisions within groups 24-27 and 28-31 are again very difficult to interpret, being based mainly on the substitution of a few taxa at each division.

#### 5.4.4 Definition of reach communities

The TWINSPAN classification produced 31 end groups between which both meaningful between-site and temporal (between-sample, within site) differences and some chance between-sample differences could be discerned (Sections 5.4.2 and 5.4.3). At the higher level of division eight major groupings based mainly upon reach-scale differences were apparent. From these it is possible to define three major reach-types:

- "lowland" - divisions 8 and 9 (end groups 1-8) contain most of the lowland-type sites, division 8 containing the most faunistically rich samples.
- "upland" - division 15 (end groups 28-31) contain the upland-type sites.
- "intermediate" - divisions 12, 13 and 14 (end groups 17-27) contain sites possessing taxa in common with both the "upland" and "lowland" sites, the three divisions representing slightly different communities, all faunistically rich.

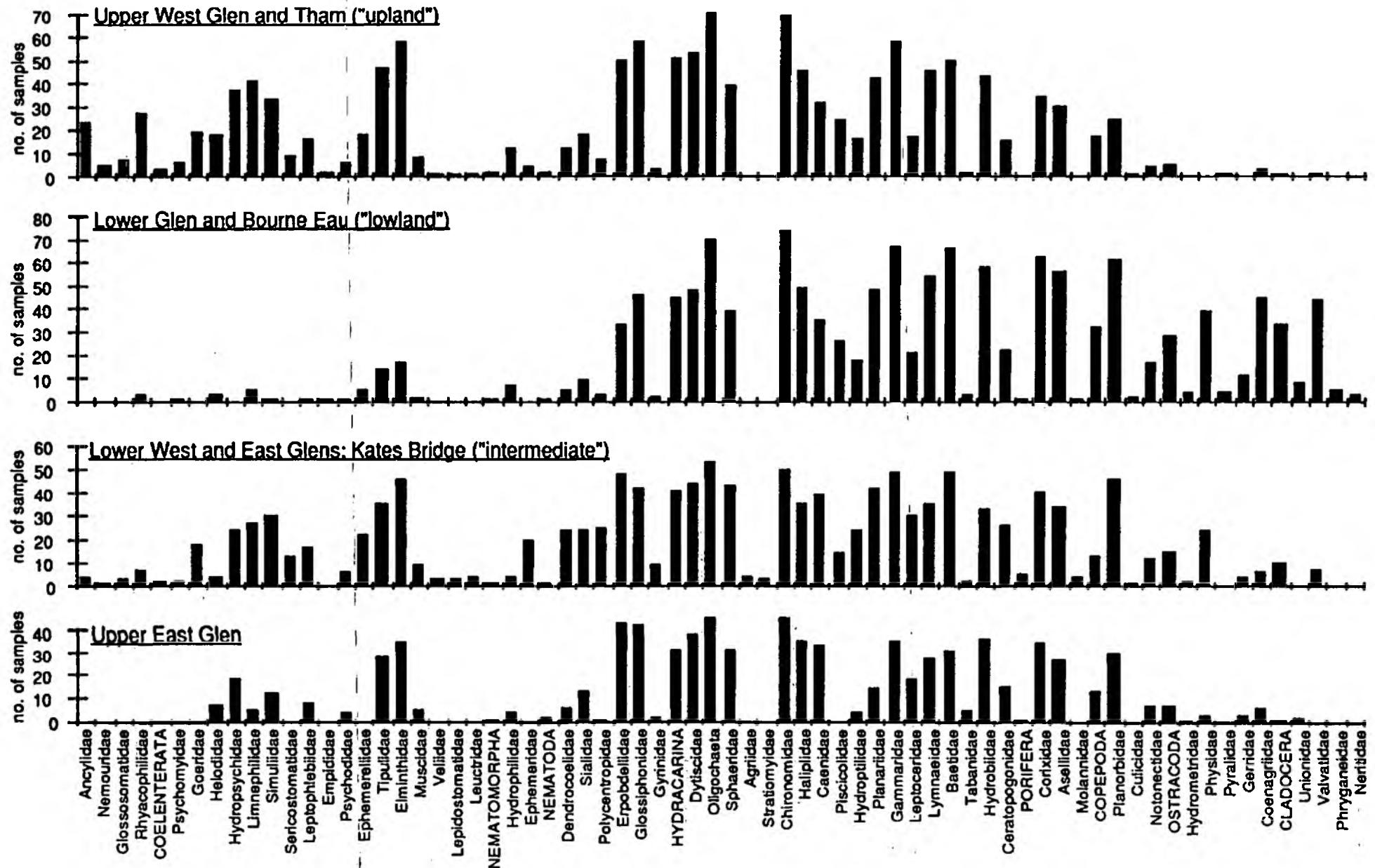


Figure B8. Frequencies of taxa within reach-types.

| "UPLAND"         | "INTERMEDIATE"   | "LOWLAND"       |
|------------------|------------------|-----------------|
| Chironomidae     | Chironomidae     | Chironomidae    |
| Oligochaeta      | Oligochaeta      | Oligochaeta     |
| Gammaridae       | Gammaridae       | Gammaridae      |
| Baetidae         | Baetidae         | Baetidae        |
| Corixidae        | Corixidae        | Corixidae       |
| Planorbidae      | Planorbidae      | Planorbidae     |
| Hydrobiidae      | Hydrobiidae      | Hydrobiidae     |
| Asellidae        | Asellidae        | Asellidae       |
| Lymnaeidae       | Lymnaeidae       | Lymnaeidae      |
| Haliplidae       | Haliplidae       | Haliplidae      |
| Dytiscidae       | Dytiscidae       | Dytiscidae      |
| Planariidae      | Planariidae      | Planariidae     |
| Glossiphonidae   | Glossiphonidae   | Glossiphonidae  |
| HYDRACARINA      | HYDRACARINA      | HYDRACARINA     |
| Coenagriliidae   | Coenagriliidae   | Coenagriliidae  |
| Sphaeridae       | Sphaeridae       | Sphaeridae      |
| Caenidae         | Caenidae         | Caenidae        |
| Erpobdellidae    | Erpobdellidae    | Erpobdellidae   |
| COPEPODA         | COPEPODA         | COPEPODA        |
| OSTRACODA        | OSTRACODA        | OSTRACODA       |
| Piscicolidae     | Piscicolidae     | Piscicolidae    |
| Ceratopogonidae  | Ceratopogonidae  | Ceratopogonidae |
| Leptoceridae     | Leptoceridae     | Leptoceridae    |
| Hydroptilidae    | Hydroptilidae    | Hydroptilidae   |
| Elminthidae      | Elminthidae      | Elminthidae     |
| Notonectidae     | Notonectidae     | Notonectidae    |
| Tipulidae        | Tipulidae        | Tipulidae       |
| Sialidae         | Sialidae         | Sialidae        |
| Hydrophilidae    | Hydrophilidae    | Hydrophilidae   |
| Limnephilidae    | Limnephilidae    | Limnephilidae   |
| Ephemerellidae   | Ephemerellidae   | Ephemerellidae  |
| Dendrocoelidae   | Dendrocoelidae   | Dendrocoelidae  |
| Rhyacophilidae   | Rhyacophilidae   | Rhyacophilidae  |
| Helodidae        | Helodidae        | Helodidae       |
| Polycentropidae  | Polycentropidae  | Polycentropidae |
| Tabanidae        | Tabanidae        | Tabanidae       |
| Muscidae         | Muscidae         | Muscidae        |
| Gyrinidae        | Gyrinidae        | Gyrinidae       |
| Hydropsychidae   | Hydropsychidae   |                 |
| Simuliidae       | Simuliidae       |                 |
| Ancylidae        | Ancylidae        |                 |
| Goeridae         | Goeridae         |                 |
| Leptophlebiidae  | Leptophlebiidae  |                 |
| Sericostomatidae | Sericostomatidae |                 |
| Glossosomatidae  | Glossosomatidae  |                 |
| Psychodidae      | Psychodidae      |                 |
| Psychomyiidae    | Psychomyiidae    |                 |
| Ephemeridae      | Ephemeridae      |                 |
| COELENTERATA     | COELENTERATA     |                 |
|                  | Physidae         | Physidae        |
|                  | CLADOCERA        | CLADOCERA       |
|                  | Valvatidae       | Valvatidae      |
|                  | Gerridae         | Gerridae        |
|                  | Hydrometridae    | Hydrometridae   |
| Nemouridae       |                  | Unionidae       |
| NEMATODA         |                  | Phryganeidae    |
| NEMATOMORPHA     |                  | Pyralidae       |
| Empididae        |                  | Neritidae       |
|                  |                  | Culicidae       |

Table B3. Taxa lists for the three reach-types.  
Taxa occurring in less than 1% of samples in the reach omitted.

Divisions 10 and 11 (end groups 9-16), which include the East Glen sites, represent samples with impoverished faunas rather than any distinct reach community.

Figure B8 displays histograms showing the frequency of occurrence of taxa within each reach-type. The taxa have been ordered according to their proportional occurrence in the two major reach-types: Taxa occurring more frequently in the "upland" sites are ranged to the left; those more frequent in the "lowland" sites ranged to the right. It can be seen that although there are a large number of ubiquitous taxa, the majority of taxa do show a preference for one or other reach-type. The "intermediate" sites possess taxa of both community-types, plus two taxa unique to these sites (Agriidae and Stratiomyidae - at very low frequency so probably chance occurrences). Interestingly, the upper East Glen sites still record taxa from both reach-types as well as the ubiquitous ones. This suggests that the poorer faunas are due to a reduced habitat diversity at individual sites rather than a general degradation at all sites.

On the basis of the TWINSPAN divisions alone it is possible to predict the potential "best" community which each of the three major reach-types could support: For the "lowland" sites the combination of taxa associated with division 8 represents the most diverse community of this reach-type. The "upland" sites potentially possess those taxa recorded in the division 15 samples. The "intermediate" sites which fall into divisions 12-14 may possess the larger range of taxa defined by these divisions, although sites within this reach type are more variable, some tending to the "upland" and some to the "lowland" end of this intermediate spectrum. Table BB3 lists the taxa occurring in the divisions associated with the major reach types, which could potentially be found in any site within that reach. Taxa occurring in less than 1% of samples have been omitted as these may be chance rarities. It is notable that the "specialist" taxa (restricted to one or two reach-types) are not necessarily those having the highest BMWP scores, ie. specialism and sensitivity are not synonymous.

## **5.5 Long-term changes in biological quality**

### **5.5.1 Changes in scores at the catchment scale.**

For the three indices: BMWP, ASPT and number of scoring taxa per sample, significant changes were found over the 15 year period. When all the sample scores for each year were plotted against year of sampling (Figure B9.) it was found that all three scores increased between 1976 and the late 1980's, but then declined in 1989-91. It appears that the number of invertebrate taxa have increased and then declined again, and also that the taxa which have increased and then decreased in frequency are those most sensitive to organic pollution as measured by these indices. It must be remembered, however, that the range of sites surveyed has changed during this period, and it is quite possible that the earliest samples tended to be taken in response to reported pollution incidents, whereas in later years more routine sampling of "unpolluted" sites was taking place. Such a change in sampling strategy may produce the observed increase in biological quality.

### **5.5.2 Changes in scores at the site and reach scale**

A similar analysis of the records for individual sites where at least six samples had been taken over the 1976-91 period confirmed the overall trends found in the whole data set, but also revealed sites with different temporal changes in scores. Figure B10 shows those sites where a significant trend in score over time was seen when scores were regressed against year of sampling, and indicates the type of change - from steady increase, through initial increase followed by a slight decrease; slight increase followed by major decrease, to steady decrease in score. (Sites where the records only covered part of the 1976-91 period are shown overlapping the groups to which they may belong.)

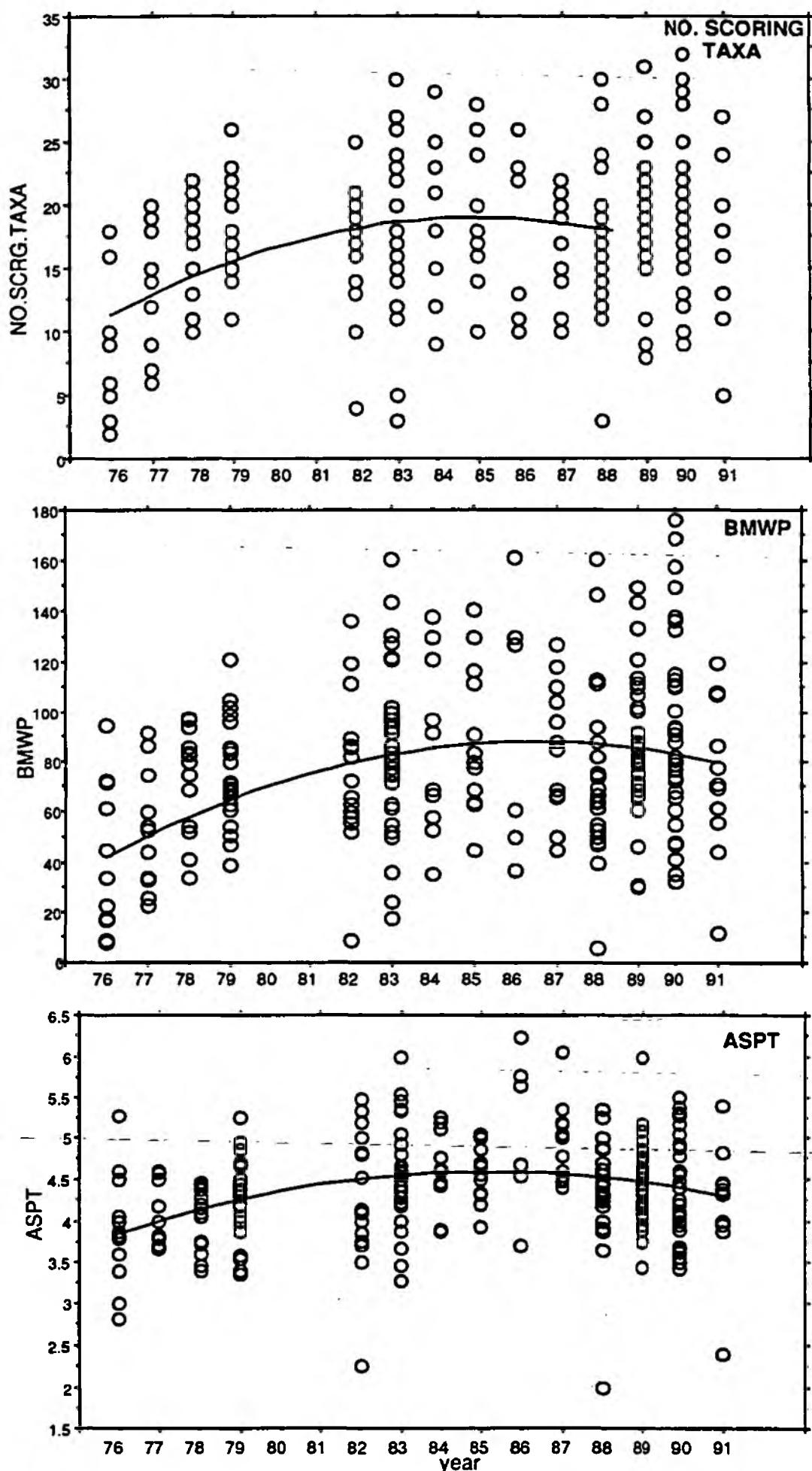
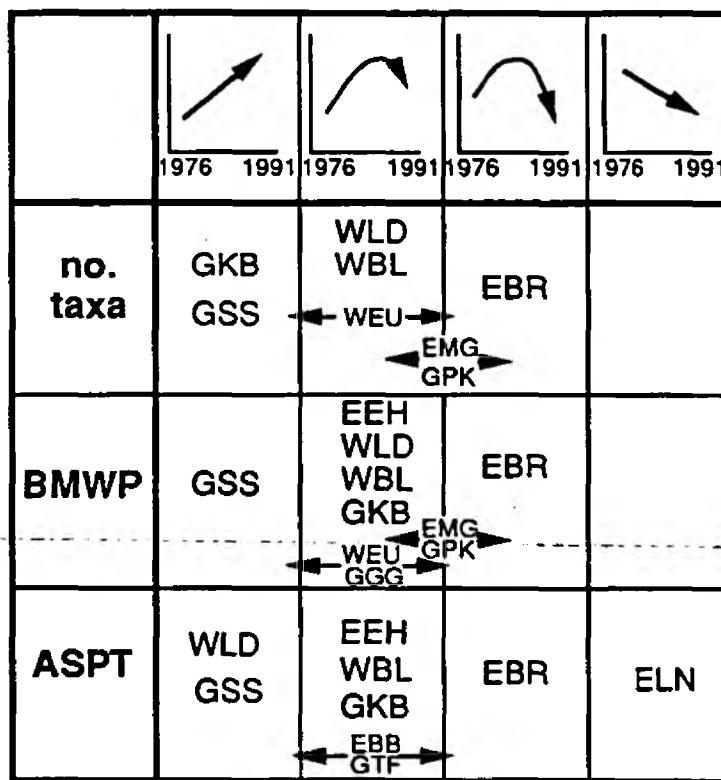


Figure B9. Regressions of number of scoring taxa, BMWP and ASPT against year of sample for all River Glen invertebrate samples.



**Figure B10.** Patterns of 1976-1991 changes in number of scoring taxa, BMW P and ASPT for sites where the trend was significant.

The majority of sites followed the second trend, as seen in the total data set, but a few - the lower main river sites and one lower West Glen site - showed steady increases in score over the whole period. In contrast, a number of the East Glen sites showed pronounced decreases in scores. This strong regional pattern suggests an increase in quality of the main river despite major decreases in quality (due to pollution or drought?) in the East Glen.

Not all sites showed the same trend in all three scores, suggesting very different community changes. For instance, the site GSS showed a steady increase in BMWP as a result of increases in both the number and "quality" of taxa (ASPT). Sites EBR and WBL displayed similar changes but in the reverse direction, with recent decreases in BMWP, number and Average Score Per Taxon. GKB, however, showed a rise then fall in BMWP despite continually rising numbers of taxa, due to the ASPT declining. This site may in recent years be experiencing a decrease in water quality, but due to physical (perhaps hydrological?) changes is able to support a greater diversity of the more pollution-tolerant taxa. Site ELN may be responding in a similar way, as the only significant change in score was a steady decrease in ASPT.

## 5.6 Site case studies

### **5.6.1 West Glen, upstream of Essendine**

Figure B11 displays the results of TWINSPAN classification, and ordination using PP correspondence analysis of the sample data for the site WEU. TWINSPAN indicator taxa, which could be used alone to produce similar groupings, are shown on the dendrogram and ordination diagram. The dates of samples and their LQI scores are indicated. Preferential taxa associated with each TWINSPAN division are listed in Appendix 3.2.

The records for WEU span the period 1976 to 1983. During this time the BMWP scores and number of scoring taxa increased significantly, but without a corresponding increase in ASPT, suggesting that the additional taxa in later years are not significantly more pollution-sensitive. Environmental changes other than an improvement in water quality may have taken place to account for the increased diversity of taxa.

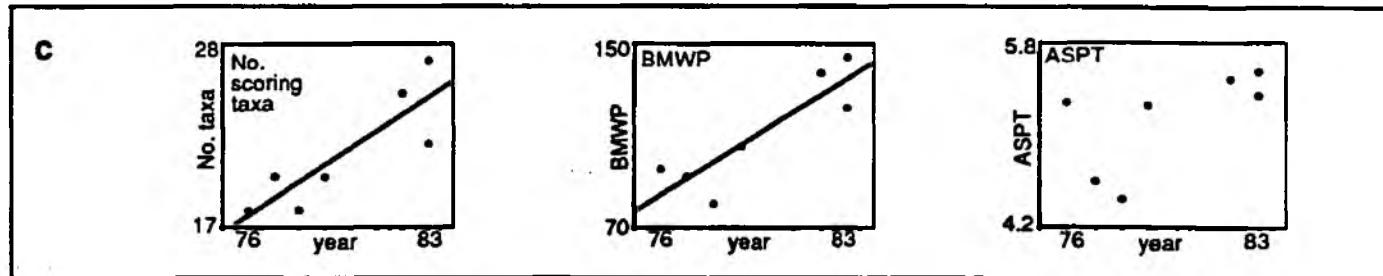
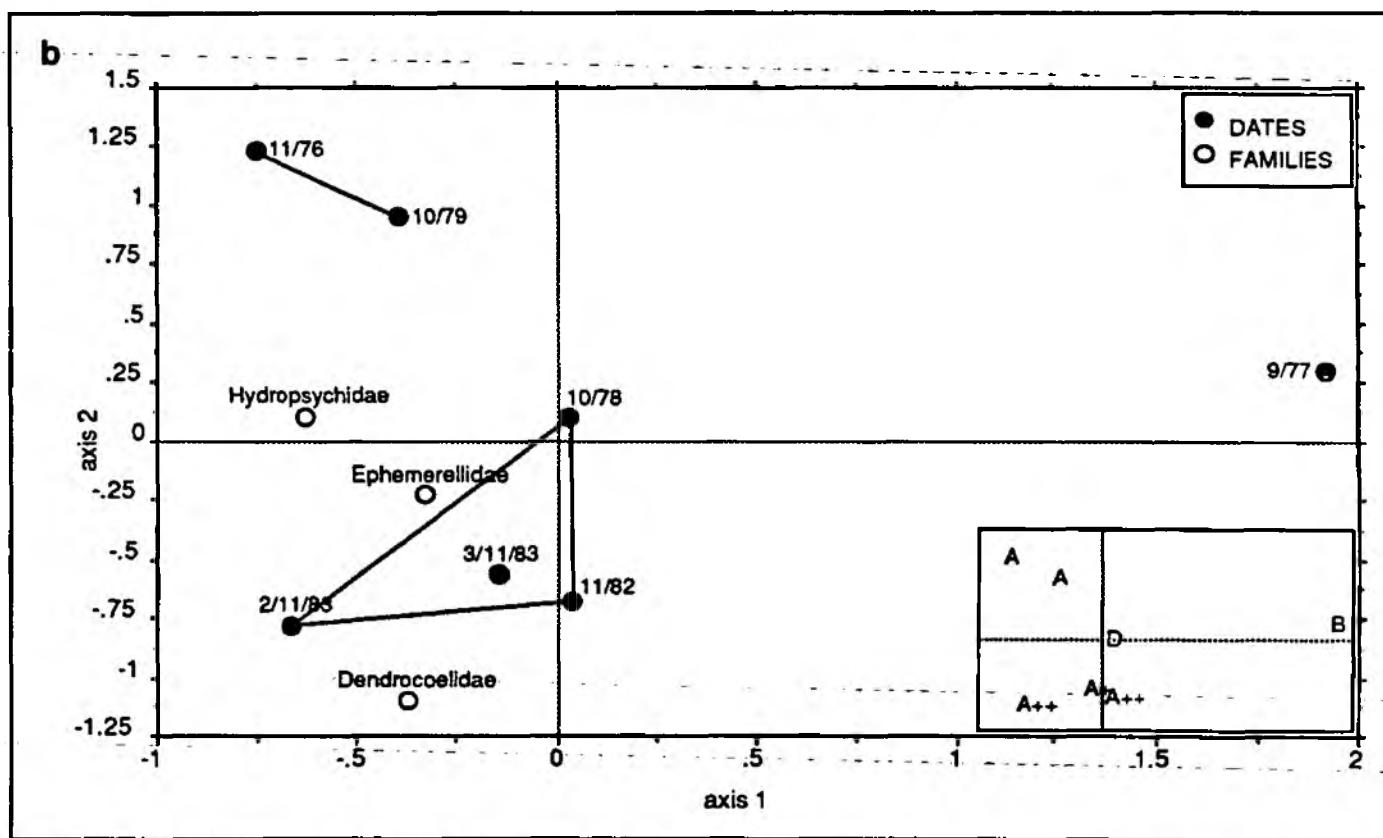
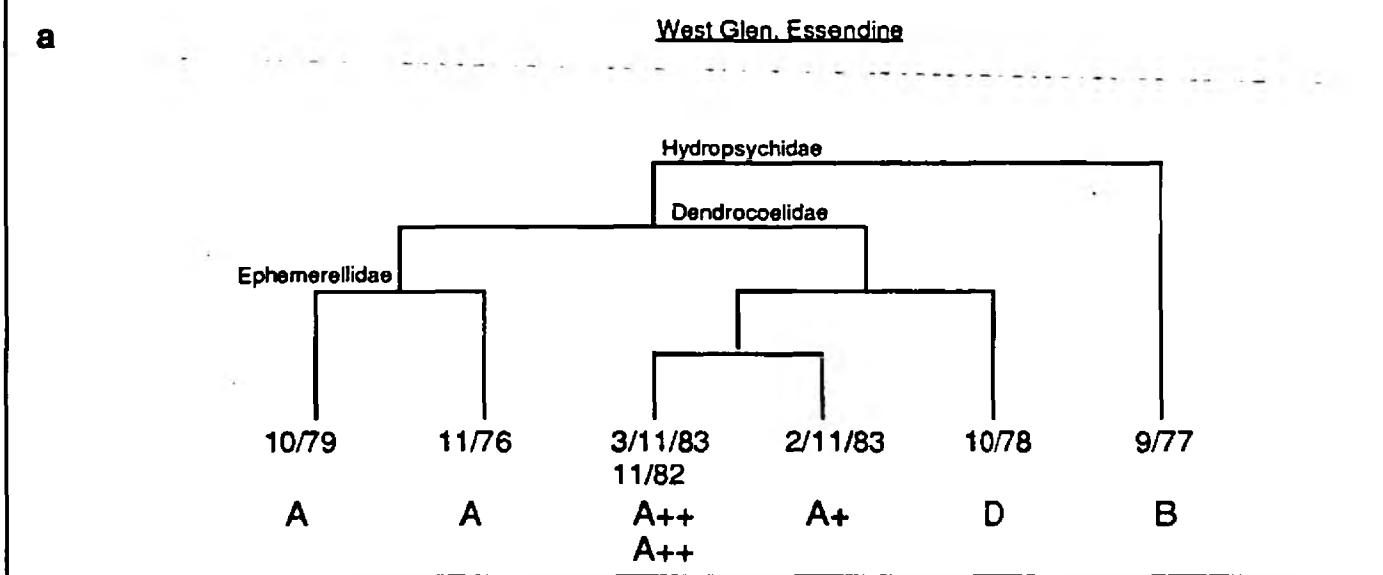
Both classification and ordination of the data identify three distinct groups of samples:

- 1: 9/77
- 2: 10/79; 11/76
- 3: 2/11/83; 3/11/83; 11/82; 1078

Sample 9/77 is most clearly separated from the other samples by both analyses. It is very probable that seasonal factors are responsible for this separation, as taxa absent from this September sample but frequent in the others include the mayflies Leptophlebiidae, Ephemerellidae and Ephemeridae as well as a number of other insect taxa that would be absent or extremely small in this month.

Samples in group 2 - 10/79 and 11/76 possess four taxa absent from group 3: Hydroptilidae, Sialidae, Gyrinidae and Copepoda. The latter three of these in particular can tolerate, or even require, standing water conditions. This together with the fact that taxa absent from group 2 but present in samples later than 1979 include the Simuliidae and Rhyacophilidae - taxa requiring flowing water - suggest that the hydrology of the site may have changed over this period.

Sample 3/11/83 was taken in addition to the sample on 2/11/83 as dredging was in progress on the first occasion. A number of taxa are present in one sample but not the other, but with no clear pattern attributable to the disturbance. It is notable that despite recent dredging these samples score higher than previous ones.



**Figure B11.** a) TWINSPAN classification, b) ordination and c) changes in biotic scores based on family data for the West Glen at Essendine

### 5.6.2 West Glen at Banthorpe Lodge

Summary diagrams of the analyses are shown in Figure B12, and TWINSPAN preferential taxa listed in Appendix 3.2.

Sixteen samples spanning the period 1976-91 were collected from this site. Biological quality varied considerably during this period, with a clear pattern of increasing quality until 1991, when a marked deterioration occurred.

On the basis of the TWINSPAN analysis four main groups can be distinguished:

- 1: 7/90, 2/83, 2/90, 11/88, 6/89, 11/89;
- 2: 7/85, 6/86, 6/87, 8/84, 11/83;
- 3: 8/79, 5/79, 5/91, 10/78;
- 4: 11/76.

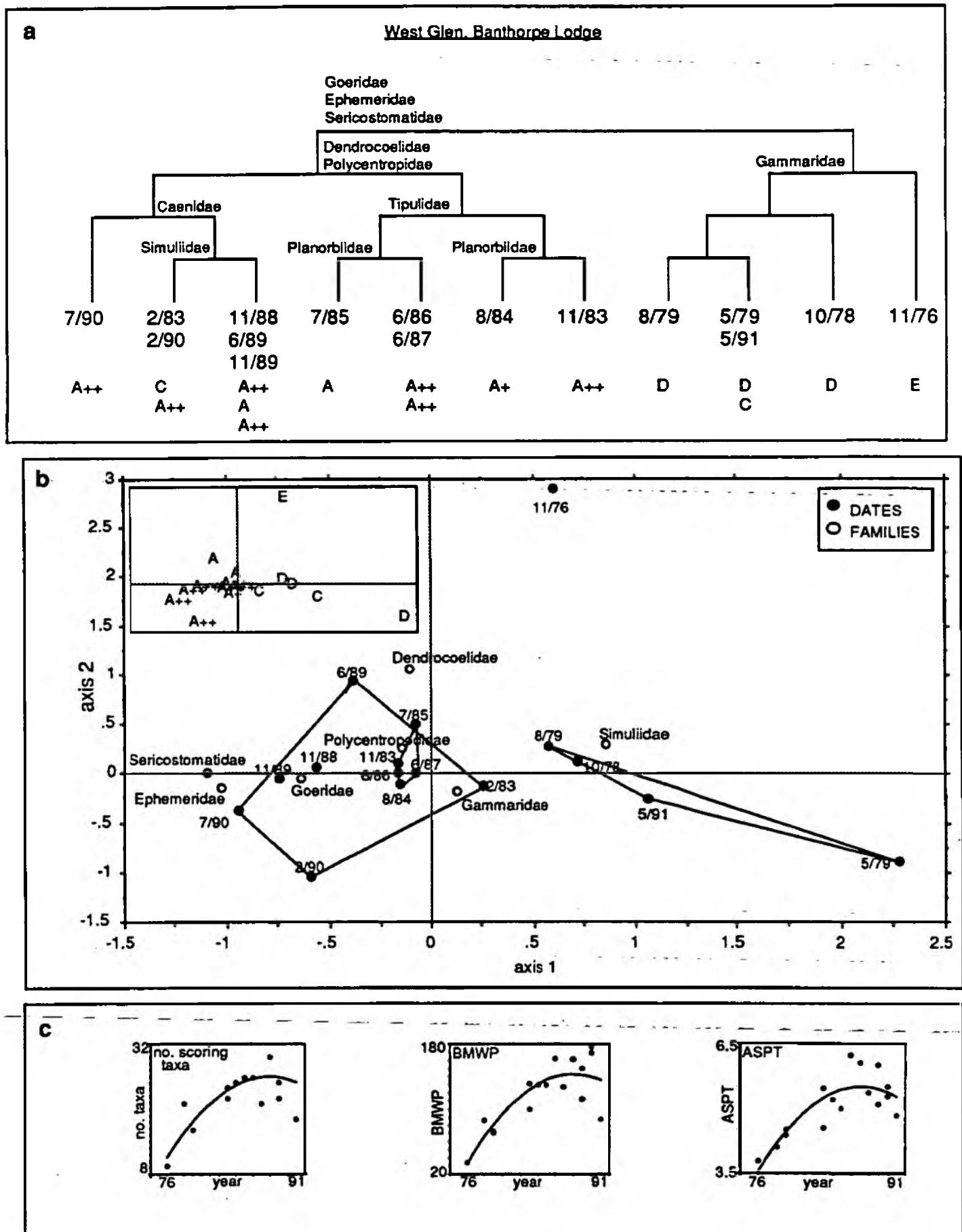
The first division separates the low-scoring, pre-1980 samples plus the 1991 sample from the rest. The only taxa showing a preference for this group are three families of Diptera: Ceratopogonidae, Stratiomyidae and Muscidae, and Copepoda. All are tolerant of moderate pollution and stagnant conditions.

Within the low scoring groups 3 and 4, the group 4 sample, the earliest in the recorded period, has a markedly reduced fauna and clearly separates in the classification and ordination. Preferential taxa for this sample include the Dytiscidae, Copepoda, Ostracoda, Lymnaeidae and Dendrocoelidae. These taxa, more notable as tolerant of standing water conditions rather than extreme pollution, suggest that the 1976 sample has an unusual and reduced community due to drought conditions rather than pollution, although the two may very well be acting together.

It is notable that the poor quality 1991 sample is associated with the pre-1980 samples of similar low score, both in the classification and the ordination, showing that not only has the fauna changed, but that the new community is very similar to that 12 years earlier. Presumably the taxa gained or replacing other taxa during the 1980's are the ones first lost as conditions deteriorate. It would also suggest that the physical change in habitat is towards similar conditions to those in the late '70's, which could suggest that there has been a recent deterioration in the habitat (due to drought) which has otherwise remained fairly constant.

The invertebrate communities of samples comprising TWINSPAN groups 1 and 2 are characterised by a wide variety of relatively pollution-intolerant taxa, especially Ephemeroptera and Trichoptera. The Ephemeropteran families showing preference for this group are those associated with clean, fine substrates and slow to moderate flow velocities (Caenidae, Ephemeridae and Leptophlebiidae). The presence of the Trichopteran families Hydropsychidae and Polycentropodidae, and the molluscs Aculyidae suggest more stable but again un-silted substrates. The site is probably a clean gravel and sand riffle or run under normal flow conditions, becoming silted or ponded under very low flows or organic pollution. Other taxa associated with weedy margins, such as the Corixidae and Agriidae are also associated with this group.

The samples separated by TWINSPAN into groups 1 and 2 appear to be largely overlapping in the ordination, implying that there are relatively insignificant differences in the communities of the two groups. According to the preferential taxa offered by TWINSPAN the species characterising group 2 include several associated with weedy conditions - several molluscs (Lymnaeidae, Planorbidae, Hydrobiidae) Corixidae, Agriidae and Asellidae.



**Figure B12.** a) TWINSPAN classification, b) ordination and c) changes in biotic scores based on family data for the West Glen at Banthorpe Lodge

### 5.6.3 River Glen at Kates Bridge

Summary diagrams of the analyses are shown in Figure B13, and TWINSPAN preferential taxa listed in Appendix 3.2.

The changes in the invertebrate community at Kates Bridge between 1976 and 1991 are more complex than those shown for sites WEU and WBL (above). Biological quality as measured by BMWP and LQI has risen between 1976 and mid-1990 but then fallen in late-1990-91. The ASPT showed greatest change between the late 1970's and early 1980's with a quite dramatic increase. ASPT then fell in 1991. BMWP and number of scoring taxa showed greater proportional increases between the early 1980's and late 1980's/1990 than did ASPT, suggesting some other habitat improvement than water quality may have been involved to account for the increasing diversity.

The four major TWINSPAN groups contain the following samples:

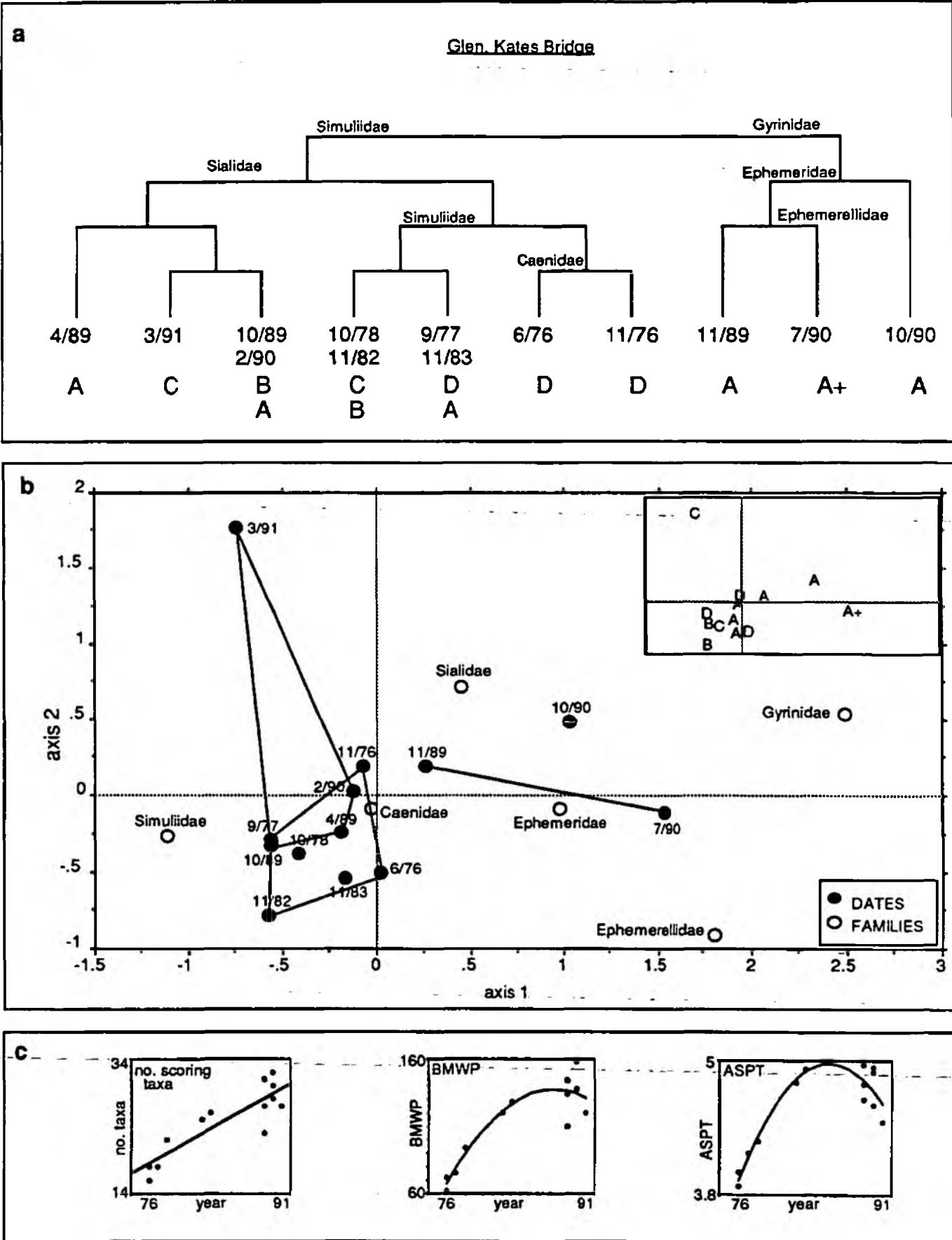
- 1: 4/89, 3/91, 10/89, 2/90;
- 2: 10/78, 11/82, 9/77, 11/83, 6/76, 11/76;
- 3: 11/89, 7/90;
- 4: 10/90.

The first division separates groups 3 and 4 from 1 and 2 on the basis of a large number of taxa confined to, or more frequent in, the former two groups. These include a surprisingly large number of Hemipterans: Notonectidae, Gerridae, Hydrometridae and Veliidae. These indicate slow-flowing conditions (or at least slow-flowing margins), probably with macrophytes. Being relatively mobile these animals may indicate habitats where there has been drought and subsequent recolonisation, but the presence of other taxa such as Ephemeridae, Ephemerellidae and Coenagrionidae with one or two year life-cycles makes this less likely.

Within the high-scoring groups 3 and 4, group 4 (10/90) is separated from group 3 on the presence of a number of taxa associated with more lentic conditions: Cladocera, Coenagrionidae and a number of Diptera, and the absence of several taxa more associated with flowing water and/or clean substrates eg. Caenidae, Sphaeriidae. However, this distinction is not very clear with group 3 also containing several taxa which suggest slow-flowing, weedy conditions.

There is very little meaningful difference between TWINSPAN groups 1 and 2. Only the Sialidae are present in all the group 1 but none of the group 2 samples, and no taxon is present in all group 2 but no group 1 samples. A number of taxa are more frequent in one or other group, but without any obvious pattern related to habitat preference or . The ordination also shows these two groups as largely overlapping, but separated from groups 3 and 4. Only sample - 3/91 - is pulled out in the ordination as significantly different from all the others.

Reference to the ordination diagram reveals a major difference from the situation at Banthorpe Lodge: Whereas at WBL the 1991 sample showed a community with much in common with the 1970's situation, at GKB a different fauna is present in the most recent sample. With falling biological quality the community has not reverted to the earlier type but to a very different assemblage. Sample 3/91 possesses three taxa absent from all other samples: Stratiomyidae, Psychodidae and Helodidae; while lacking two taxa present in all others: Baetidae and Corixidae. It also shares with the 1990 samples several taxa absent from or infrequent in the low-scoring, pre-1980 samples eg. Sialidae, Caenidae and Leptoceridae; while lacking Copepoda which were common pre-1980 but absent in 1990. The combination of some unique features, plus similarities with temporally close samples rather than samples of similar score, leads to the 3/91 sample plotting separately in the ordination. The fact that TWINSPAN failed to separate this sample whereas the ordination did may be a due to the weight given by the latter method to unique species rather than general similarities. The similarity of the 3/91 sample to the other later samples and contrast with the early samples suggests that a permanent change in habitat has occurred at Kates Bridge.



**Figure B13.** a) TWINSPLAN classification, b) ordination and c) changes in biotic score based on family data for the River Glen at Kates Bridge.

#### 5.6.4 River Glen at Surfleet Seas End

Summary diagrams of the analyses are shown in Figure B14, and TWINSPAN preferential taxa listed in Appendix 3.2.

This site, the furthest downstream in the system and close to the confluence with the River Welland, has 16 records of invertebrate samples spanning the 1976-91 period. During this time a biological quality as measured by BMWP, number of taxa, ASPT and LQI, has shown a general increase from very poor quality in 1976 to excellent (LQI A) in late 1990, with the 1991 sample being lower (LQI C - good quality) but not as significantly lower as in the previous examples. The range of biological quality from LQI H to A is notably lower than in the more upstream examples (LQI E to A++), as might be expected given the physical nature of this site.

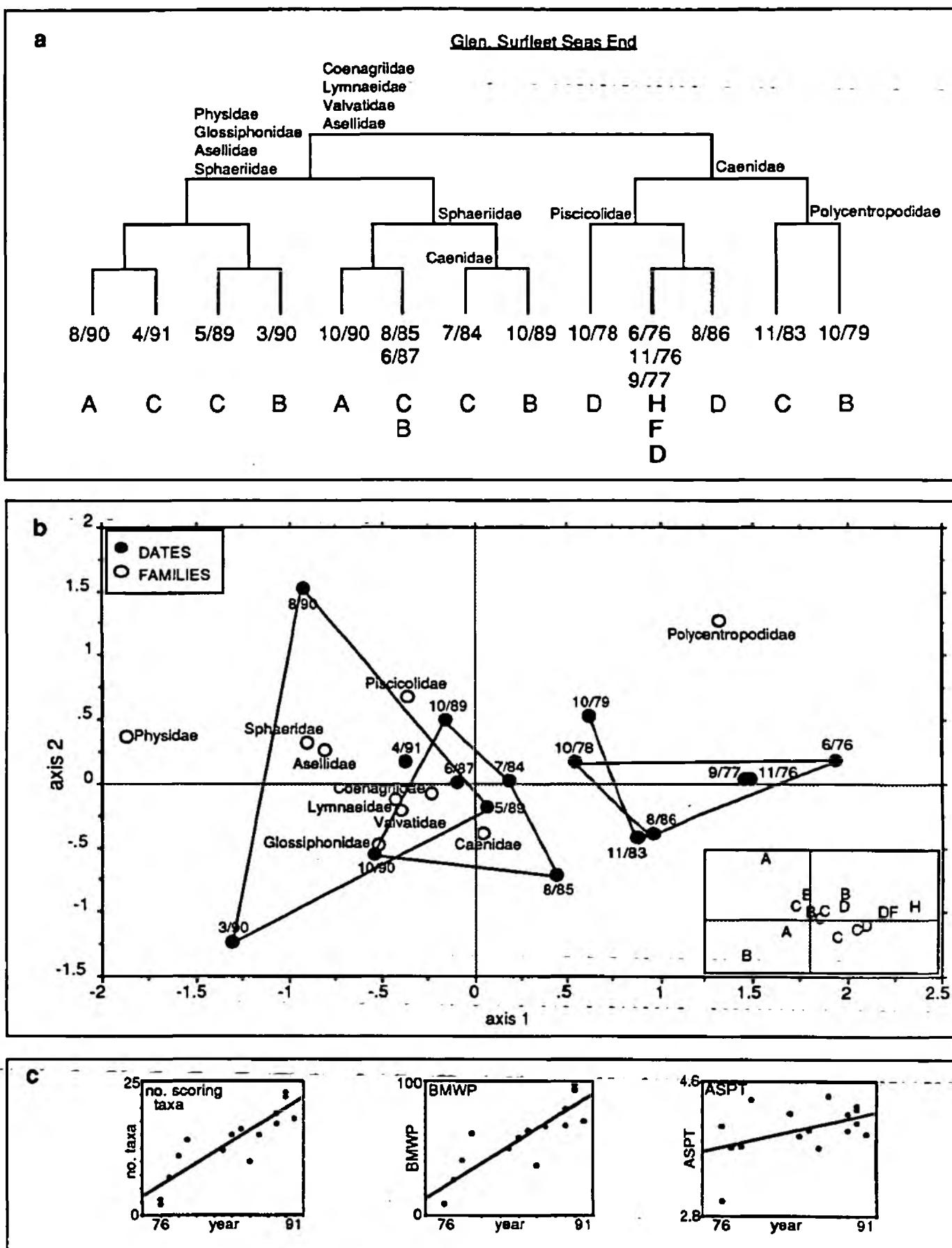
The TWINSPAN classification produces the following four major groupings:

- 1: 8/90, 4/91, 5/89, 3/90;
- 2: 10/90, 8/85, 6/87, 7,84, 10/89;
- 3: 10/78, 6/76, 11/76, 9/77, 8/86;
- 4: 11/83, 10/79.

The first major division (groups 1 and 2 from 3 and 4) is generally supported by the ordination, although this shows a broad trend along axis 1 rather than a distinct split into two groups. This first division appears to combine differences in score and year of sampling: Groups 3 and 4 are generally the earliest and lowest scoring samples, with the exceptions of 10/79 which fits in with the temporal grouping but has a higher score (B) than other samples in the group; and 8/86, which is a more recent sample but shares a low score (D) with the other group members. Two explanations could account for this interdependency of date and score: First, the range of taxa which could possibly exist at the site, in terms of their sensitivity to pollution or other habitat degradation, could be limited by water and habitat quality, and this determines the biological score observed. The actual community present, however, as well as being limited to the range so described, must be a modification of the community previously living at that site, with gains or losses of a few taxa in response to habitat changes. Secondly, changes in one aspect of habitat quality at the site, such as water quality, could produce the observed changes in biological score, but other features of the habitat, e.g. hydrology, management, which may have changed more subtly over time, determine the actual community within the limits set by (in this example) water quality. Either of these explanations would account for samples sharing more in common with other samples from a similar time period than those of similar score from a different time period.

The community differences associated with the first TWINSPAN division are largely based on a wide variety of taxa which occur in groups 1 and 2 (containing the later and higher scoring samples) which are absent from the other two groups, including many mollusc families: Unionidae, Sphaeriidae, Valvatidae, Lymnaeidae and Physidae; several beetles: Haliplidae, Dytiscidae and Elmidae; and two relatively pollution-sensitive Trichopteran families: Leptoceridae and Rhyacophilidae. There is a notable lack of Ephemeroptera - only Baetidae and Caenidae are recorded, these being present in all the groups of samples. The only taxa found more frequently in groups 3 and 4 are the Cladocera and Copepoda, suggesting fairly stagnant conditions at the time of these samples. An oddity occurring in samples 6/87 and 8/90 is the family Mysidae, suggesting some saline input at this lowland site.

The division between groups 1 and 2 is based upon a number of taxa, including the Rhyacophilidae and Physidae, which are restricted to or more frequent in group 1. Only a few taxa are limited to group 2, including the Cladocera and Gerridae. Various habitat changes such as weed cutting or dredging, could affect the distribution of the many of the taxa characterising these groups, perhaps more so than water quality.



**Figure B14.** a) TWINSPLAN classification, b) ordination and changes in biotic score based on family data for the River Glen at Surfleet Seas End.

The ordination separates samples 8/90 and 3/90 far more clearly than the classification. Sample 3/90 possesses three taxa absent from all other: Dendrocoelidae, Helodidae and Corophidae; and lacks three taxa common in the majority of other higher-scoring samples: Piscicolidae, Hydracarina and Ostracoda. Sample 8/90 again has a few unique taxa: Psychomyidae, Pyralidae; a few relatively rare ones: Unionidae, Mysidae; and lacks the beetles Elmidae which occur in the other high scoring samples. Thus the ordination again appears more sensitive to unique taxa than the classification.

Within the TWINSPAN groups 3 and 4 there is considerable variation in the numbers of taxa within samples rather than type. Samples 6/76 and 11/76 are have severely restricted fauna, containing only Oligochaetes, Gammaridae, Chironomidae, Hydracarina and Cladocera. Saline intrusion is the probable cause, due to drought. Because they share all of the taxa present in the 1976 samples TWINSPAN groups these and the 1977 sample together. However, this later sample has a considerably improved fauna with the addition of Dytiscidae, Baetidae, Planorbiidae, Corixidae and Copepoda. The other two samples in group 3 possess taxa including Haliplidae and Hydrobiidae, in addition to those in the other samples of the group.

Group 4 samples share most of the taxa present in group 3, but in addition have a number of others including two families of Trichoptera (Hydropsycheidae and Polycentropodidae) - an Order completely absent from group 3. It is the presence of taxa such as the Cladocera, Gerridae and Copepoda in these group 4 samples in common with group 3 which places them on this side of the first classification division rather than with the other samples of similar score.

It seems probable that there has been an overall gradual improvement in habitat at this site, subsequent to (but not related to) a dramatic improvement following the drought of 1976 which resulted in a period of saline intrusion.

#### 5.6.5 East Glen at Braceborough

Summary diagrams of the analyses are shown in Figure B15, and TWINSPAN preferential taxa listed in Appendix 3.2.

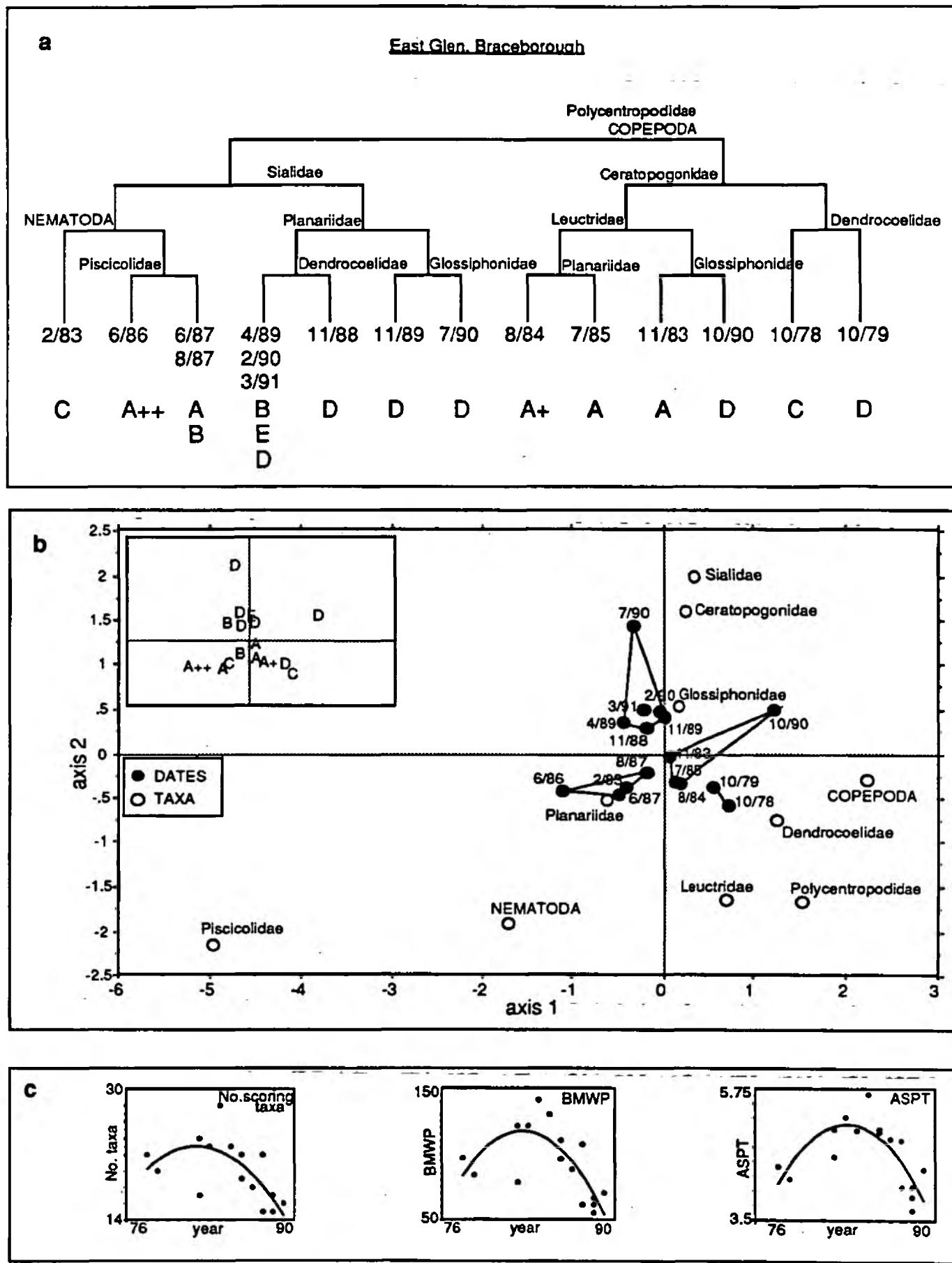
The River East Glen at Braceborough differs from the previous examples in that biological quality peaked in 1986 and declined subsequently, whereas in the majority of sites the peak was in 1989-90. Sixteen biological records exist for Braceborough, over the period 1978-91.

The TWINSPAN classification produces the following main groups:

- 1: 2/83, 6/86, 6/87, 8/87;
- 2: 4/89, 2/90, 3/91, 11/88, 11/89, 7/90;
- 3: 8/84, 7/85, 11/83, 10/90;
- 4: 10/78, 10/79.

These groups coincide with differences in both biological quality of samples and years of sampling. The first TWINSPAN division separates the later samples (groups 1 and 2) from the earlier ones (groups 3 and 4), with two exceptions (see later). Within each of these two major groups the second split appears to divide samples of high (groups 1 and 3) from low (groups 2 and 4) biological quality.

The first TWINSPAN division shows the "early" samples (groups 3 and 4) to have a number of taxa absent or less frequent in the "late" group, including relatively "sensitive" families such as the Leuctridae, Ephemeralidae, Hydroptilidae and Polycentropodidae; the molluscs Physidae and Valvatidae; a number of Coleopteran families: Haliplidae, Hydrophilidae and Gyrinidae; and the micro-crustaceans Copepoda, Cladocera and Ostracoda. The "late" group has relatively few preferential taxa: Limnephilidae, Hydropsychidae, Leptoceridae, Sericostomatidae and Psychodidae. Bearing in mind that this first split seems to be based on



**Figure B15.** a) TWINSPAN classification, b) ordination and c) changes in biotic scores based on family data for the East Glen at Braceborough.

temporal rather than water quality differences a possible explanation for the community changes seen might be that there has been a reduction in macrophyte cover or marginal habitat in more recent years.

Within the "late" group, the high-scoring group 1 separates from group 2 on the basis of a number of Ephemeroptera: Leptophlebiidae, Ephemerellidae, Ephemeridae and Caenidae; and Trichoptera: Hydropsychidae, Hydroptilidae, Limnephilidae and Glossosomatidae. Similarly, within the "early" group, high-scoring group 3 separates from group 4 because of the preference of several Ephemeroptera and Trichoptera: Leptophlebiidae, Ephemeridae, Caenidae, Limnephilidae and Goeridae; but also other insect orders: Odonata (Coenagrionidae), Plecoptera (Leuctridae) and a large number of Coleoptera and Hemiptera: Gyrinidae, Hydrophilidae, Corixidae, Notonectidae and Gerridae. These reinforce the idea that good marginal habitat may be sustaining these rich communities.

Two samples do not appear to fit in with the early/late, high-scoring/low-scoring pattern outlined above. Sample 2/83 has a low score for group 1. It lacks many taxa present in the other members of this group, but has more in common with it than with group 4. Sample 10/90 also has a low score compared to the rest of its group, but most of those taxa it has in common with other members of that group. However, these are not all the most common taxa in the group, which may account for the ordination separating this sample more clearly than the classification. In other respects the ordination confirms the TWINSPAN classification.

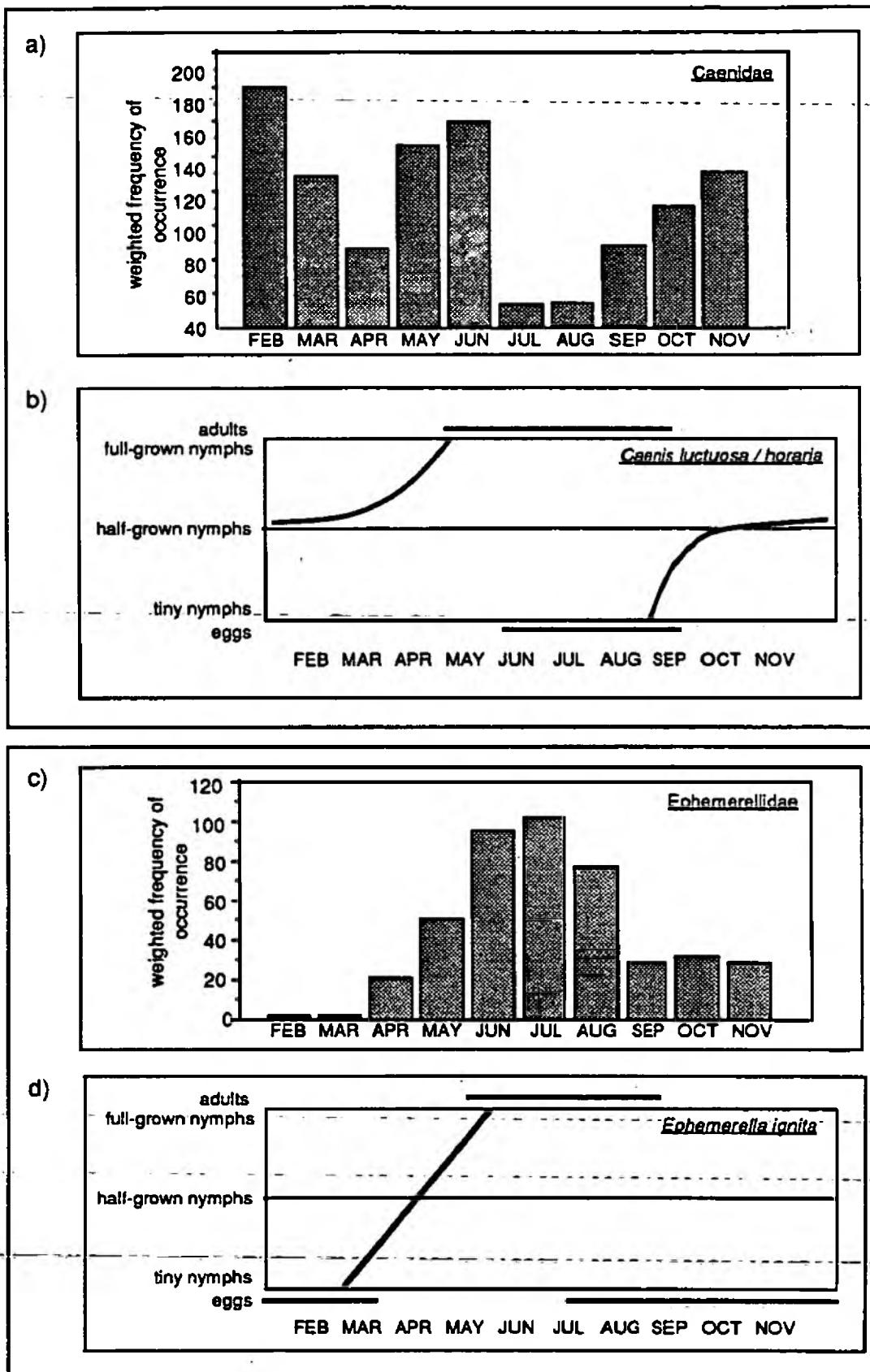
Disregarding the two "odd" samples that could be due to short-term events, the pattern suggests that there has been a permanent habitat change, possibly linked to improved marginal habitat, superimposed upon which are short term changes due to a range of stresses such as pollution or drought.

### 5.7 Seasonality

The life history patterns of many invertebrates, particularly univoltine insect species, create seasonal peaks and troughs in abundance of many taxa. It might therefore be expected that this element of seasonality in individual taxon abundance may lead to changes in numbers of taxa collected at a site and biotic scores depending on the time of year, particularly where a large proportion of the fauna share similar life histories. Insect dominated sites may show fewer taxa in early autumn, for instance, between emergence of one generation and hatching of the next.

Comparison of samples collected during different months using classification or ordination of the whole data set proved unsuitable for examining seasonal differences due to the overwhelming influence of site-specific variation. Applying similar methods to individual sites was also inappropriate as no single site had sufficient number of records.

Investigating the seasonal distribution of individual taxa was problematical due to the uneven distribution of samples between months. In an attempt to overcome this "weighted frequencies" were calculated for the commoner taxa by first separating the samples into the three major reach types ("upland", "intermediate" and "lowland" - see Section 5.4.4 ), then dividing the number of occurrences of each taxon within each month by the proportion of samples taken in each month within the reach type. The results for the three reach types were then totalled to give a single "weighted frequency" for each taxon for each month. These were then plotted against months to reveal any seasonal differences. It should be noted that the above procedure is not statistically reliable and was used only to look for general patterns. For several taxa seasonal patterns were apparent. Figure B16 shows the distribution of the Caenidae and Ephemerellidae and relates these to the known life-history curves for the common species in these families (Elliott et al 1988). The species chosen as representative of the families (*Ephemerella ignita* and *Caenis luctuosa* and *C. horaria*) have been found to be common in the River Glen (Milan, pers. comm.). For the Caenidae, a sharp decrease in frequency of collection occurs in July, corresponding to the emergence of the adults, followed by an increase during the period when



**Figure 16.** Frequency of occurrence of the families Ephemerellidae and Caenidae in each month. Frequencies are weighted to compensate for unequal sampling intensity within each month and reach type (units are arbitrary). Observed weighted frequencies: a) Caenidae and c) Ephemerellidae; temporal distribution of life-stages: b) *Caenis luctuosa* and *C. horaria* and d) *Ephemerella ignita* (adapted from Elliott *et al* 1988).

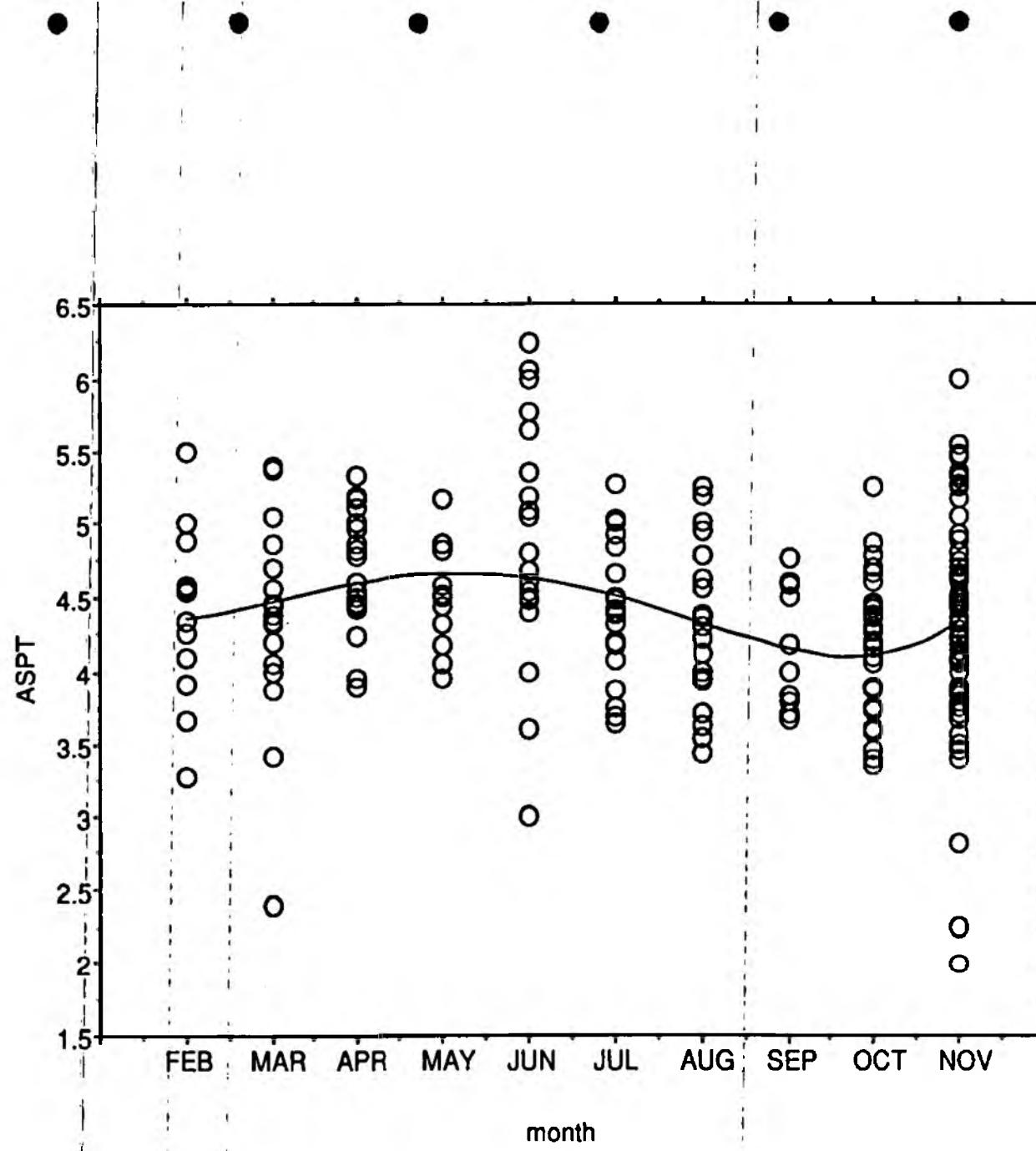
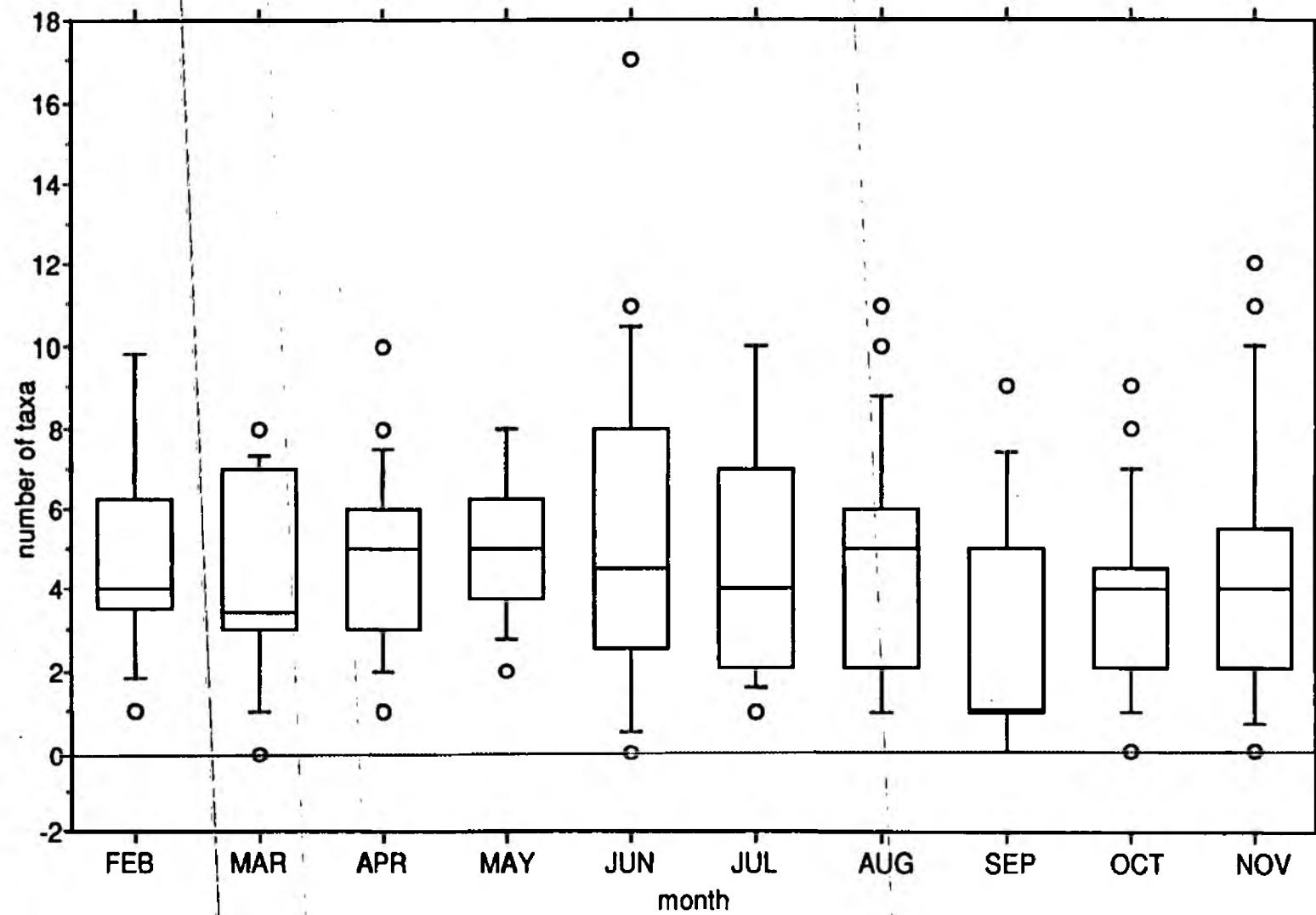


Figure B17. Regression of ASPT against month, all samples



**Figure B18.** Numbers of Ephemeroptera, Plecoptera and Trichoptera taxa collected each month. Box plots indicate means, one standard deviation error and 95% confidence limits.

the next generation of nymphs would be hatching from eggs. In the Ephemerellidae the decrease in frequency of collection (which occurs during August-September) is later than that predicted from the known life-cycle - possibly due to a regional difference in the time of emergence.

No significant seasonal differences were found in the total number of taxa or BMWP. However, a significant relationship (NB Again the problem of the uneven distribution of samples between months makes this analysis unreliable) was found between ASPT and month of collection (Figure B17). Lowest scores were found in autumn, as might be predicted as high scoring taxa are predominantly insects with aquatic larvae and nymphs, the majority of which emerge in summer. Other explanations may be possible, linked to this season being the one of lowest flows.

Based on the assumption that the taxa contributing to the seasonality of ASPT scores would be predominantly the Ephemeroptera, Trichoptera and Plecoptera, numbers of these taxa alone were plotted against month (Figure B18). Contrary to expectations no significant seasonality was found in numbers of these taxa recorded, although the seasonal trends coincided with those of ASPT (Figure B17).

## 5.8 Performance of biotic indices

The use of biotic scores to evaluate the diversity and quality of invertebrate fauna at a site potentially masks differences in communities due to the grouping of taxa with quite varied behaviours, life histories and other ecological traits into score groups using the sole criterion of sensitivity to organic pollution. Classification and ordination of samples based on individual taxa can be used to look at variability within score groups, in order to identify how much between-sample variation is being ignored by consideration of biotic scores alone.

### **5.8.1 Variation within LQI classes**

Figure B19 shows the correspondence analysis ordination of the family data from all sites with the area of the plot covered by samples from each LQI class shaded. This demonstrates that the area covered by samples within the A classes (A, A+ and A++) is greater, ie. the variety of communities is greater, than that of the lower classes. Thus grouping sites as "A" class disguises a greater range of faunas than E or F classes. Samples in the lower score classes have a greater proportion of taxa in common than those in the highest classes.

### **5.8.2 Classification and Ordination of samples within LQI classes**

#### **LQI A, A+ and A++ class samples**

Classification of the samples which scored A, A+ or A++ on the LQI scale using TWINSPAN revealed that the within class diversity could be accounted for largely by between site differences (Figures B20 and B21). A very distinct split was apparent between the lower main Glen and Bourne Eau sites on the left, and the Tham and West Glen on the right. The few East Glen samples achieving A class status split between the two main groups, with the more upstream sites on the left and downstream ones (EBR) on the right. This division mirrors that observed for the whole data set. Preferential taxa associated with this division in the A class samples (listed in Appendix 3.3) are the same as for the whole set, with the exceptions that the families Hydrophilidae, Gerridae and Physidae are also preferential for the left-hand group, and Sericostomatidae for the right-hand group in the A class samples, whereas the

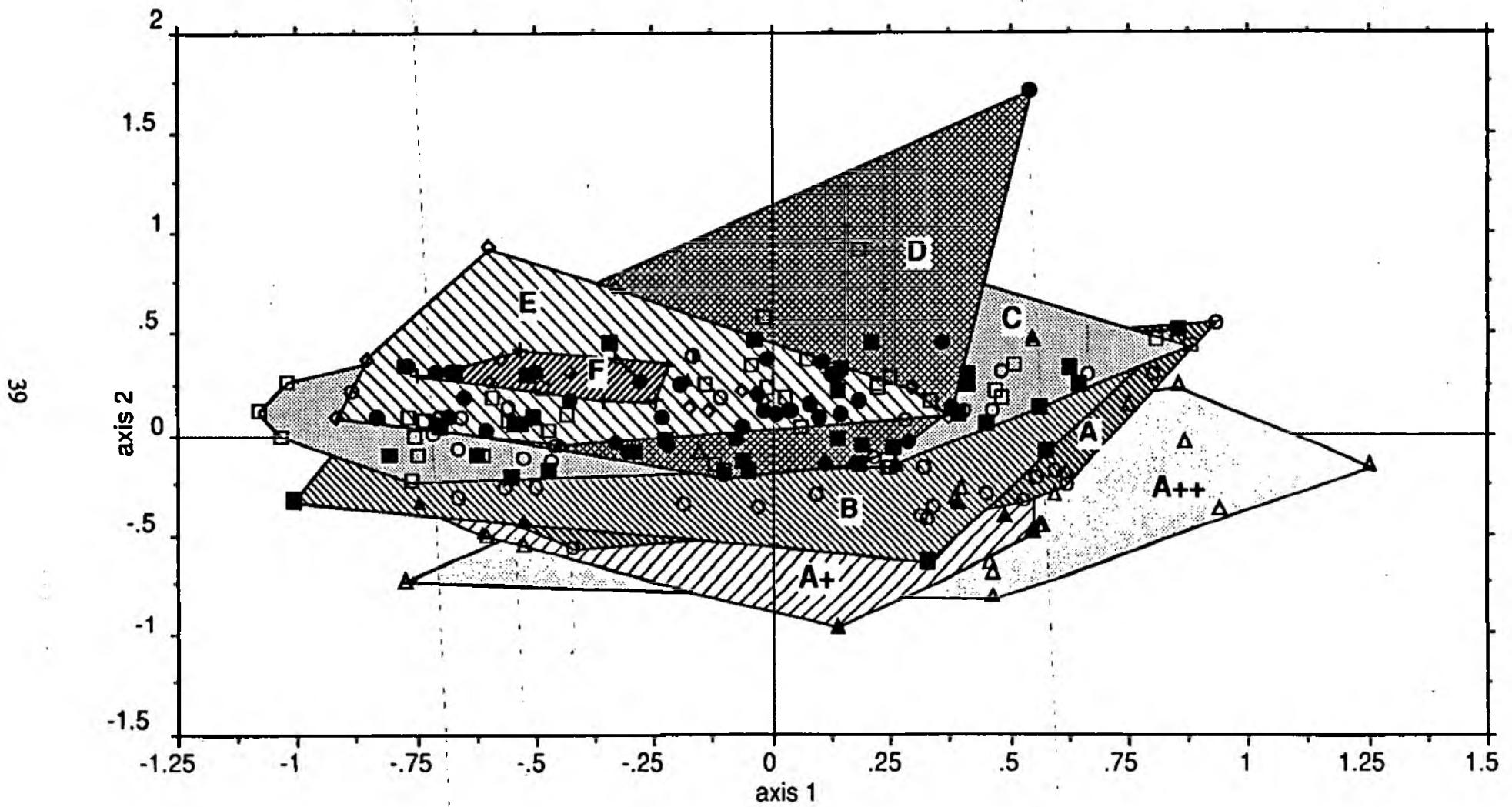
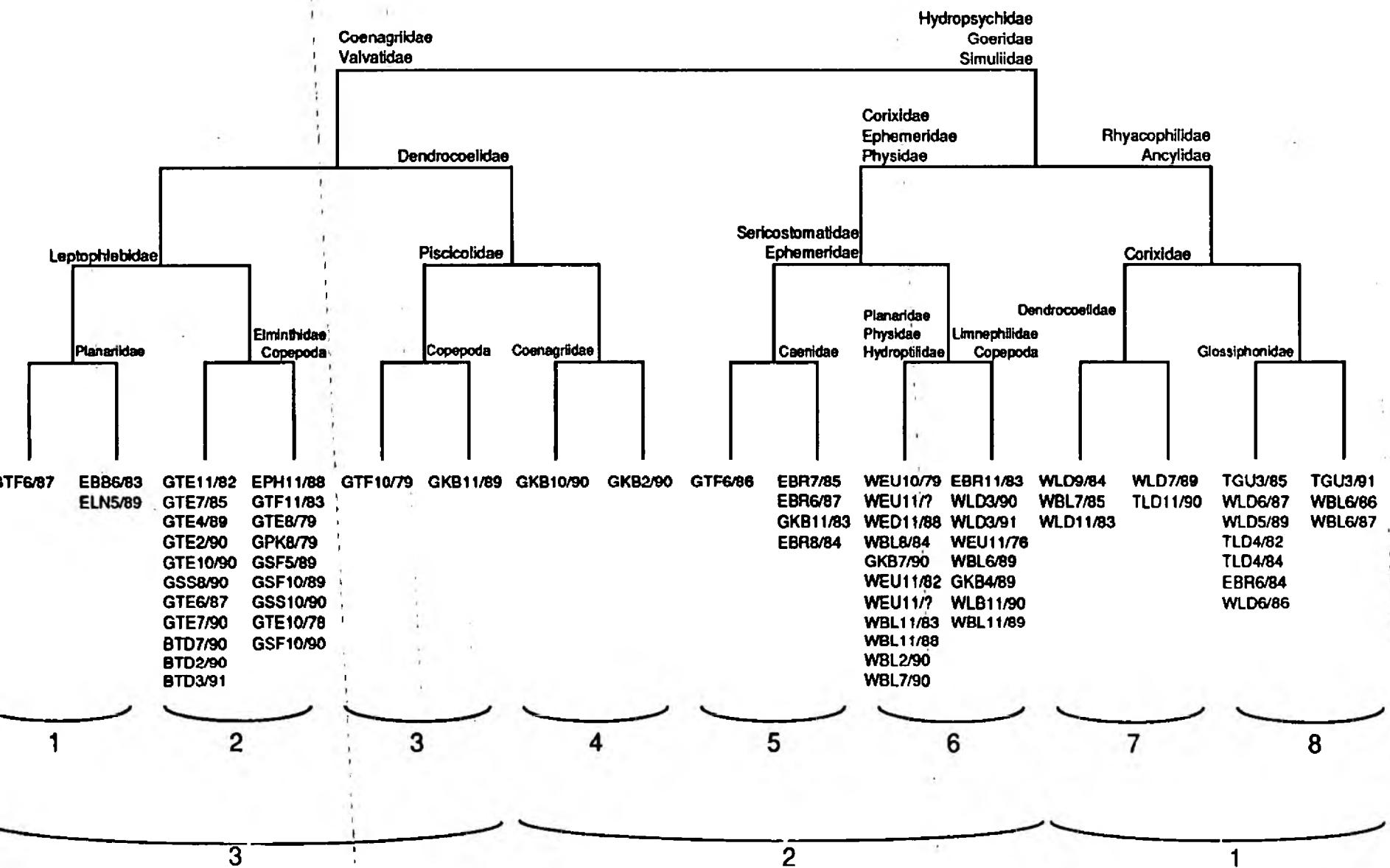


Figure B19. Ordination based on a correspondence analysis of family data for all samples; LQI classes are highlighted.



**Figure B20.** TWINSPAN classification of LQI class A, A+ and A++ samples. Groups 1-8 are referred to in figures 20 and 21, and groups 1-3 in figure 26.



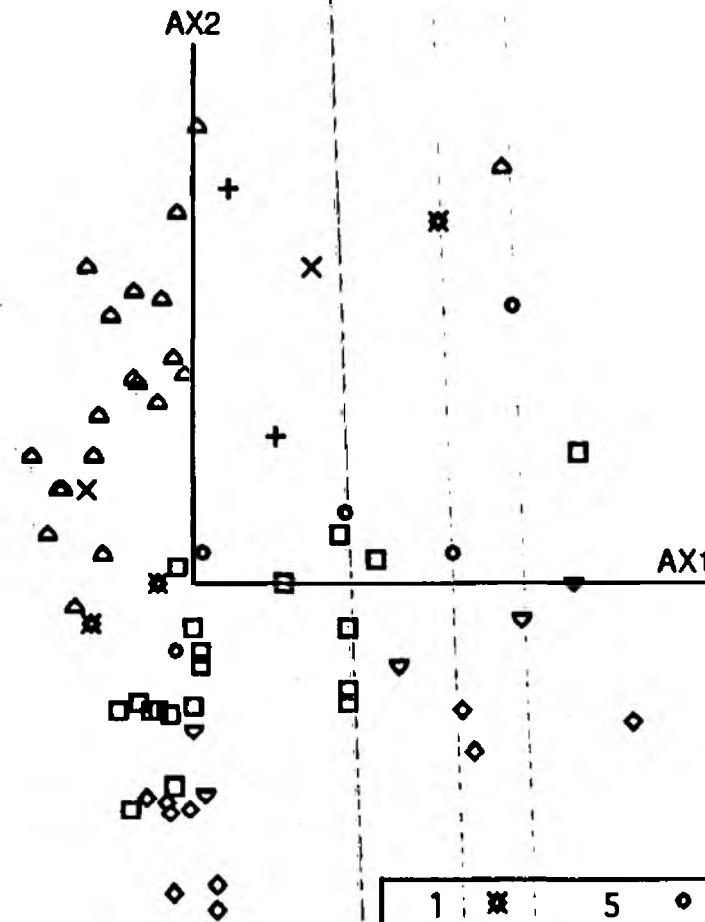
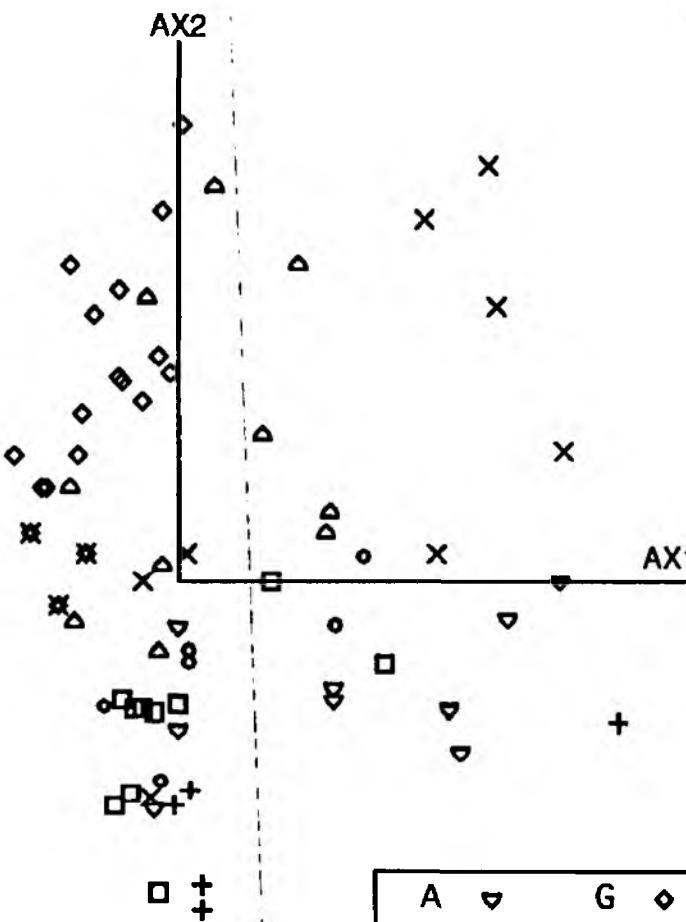
Glen LQI class A-A++ sites: TWI groupsGlen LQI class A-A++ sites: reaches

Figure B22. Ordination based on correspondence analysis of family data for samples in classes A-A++, with reaches and TWINSPAN end groups (see Figure 20) highlighted.

Hydroptilidae are no longer more frequent in the right-hand group in this classification. Thus the site-based division is generally sharper in the A class group alone.

The next division within the lower Glen samples separates a small number which are relatively taxa-rich (one GTF and three GKB samples) from the rest, while in the West Glen and Tham samples split roughly between the lower West Glen and the few East and main Glen samples on the left, and the upper West Glen and Tham samples on the right. This again mirrors the classification of the whole data set.

Ordination of the A class data (see Figure B22, which shows the same ordination with reaches and TWINSPAN groups separately highlighted) confirmed the higher groupings produced by TWINSPAN, but showed the lower divisions to be somewhat arbitrary. Highlighting the samples by reach showed again that the TWINSPAN groups broadly coincided with the different reaches. The implications of the classification and ordination are that class A - A++ samples from different reaches contain communities characteristic of those reaches, with particular characteristic taxa, not just a greater diversity.

### **Classification and Ordination of LQI B and C class samples**

Repeating the TWINSPAN classification (Figures B23 and B24) and correspondence analysis (Figure B23 - axis 3 appears to be as significant as axis 2 in this analysis, so plots using axes 1 and 2 and 1 and 3 are shown) on the samples scoring either B or C on the LQI scale produced as, if not more, positive a reach-separation as for the A class data. This was rather unexpected, as a greater diversity was predicted between the higher scoring samples. Again the lower, main river and Bourne Eau sites are separated from the rest on the basis of a rich mollusc fauna and certain other taxa characteristic of slow flowing or still-water sites with abundant macrophytes, e.g. Cladocera, Coenagriidae and Notonectidae. (Preferential species for this classification are listed in Appendix 3.3.)

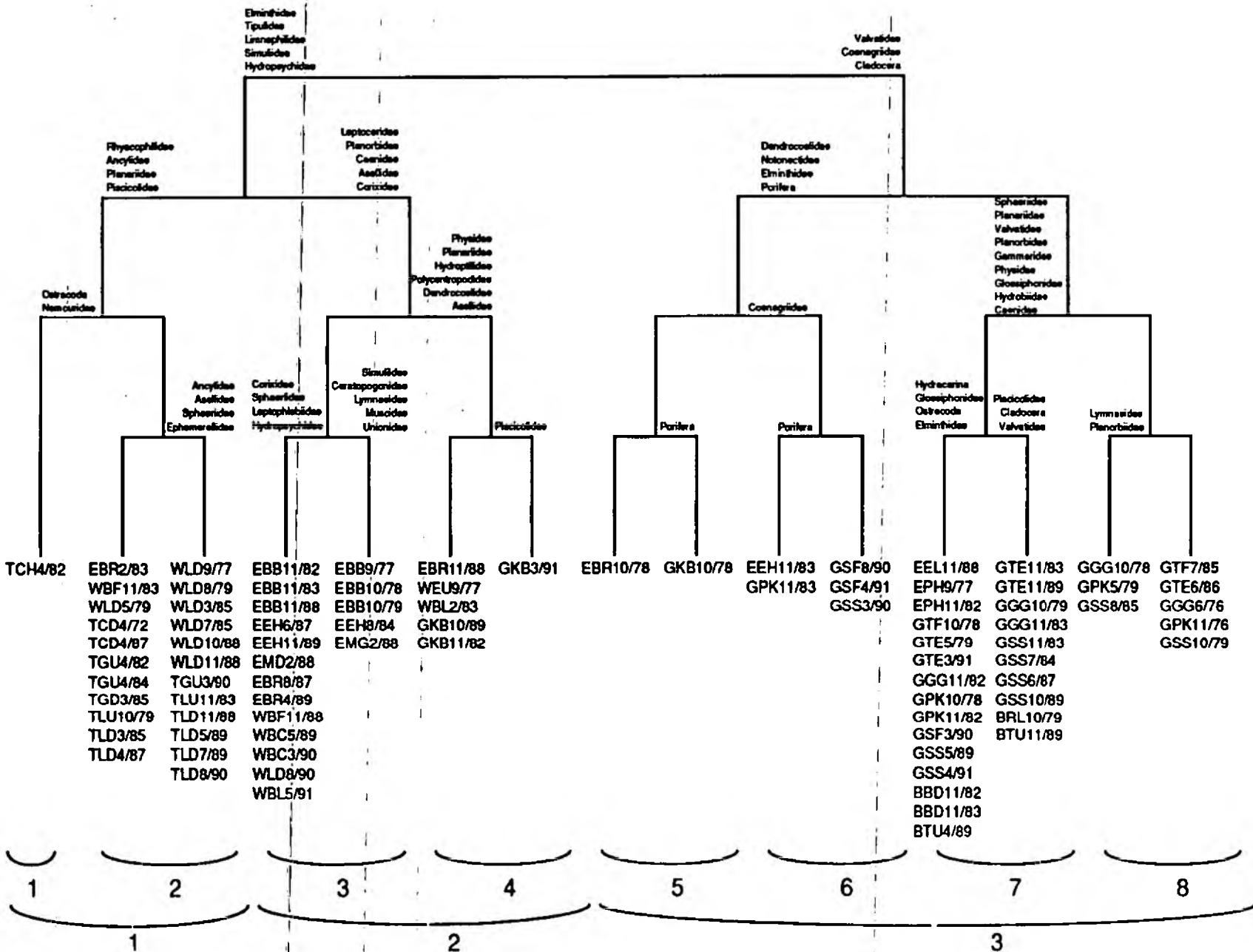


Figure B23. TWINSPLAN classification of LQI class B and C samples. Groups 1-8 are referred to in figures 23 and 24, and groups 1-3 in figure 26.



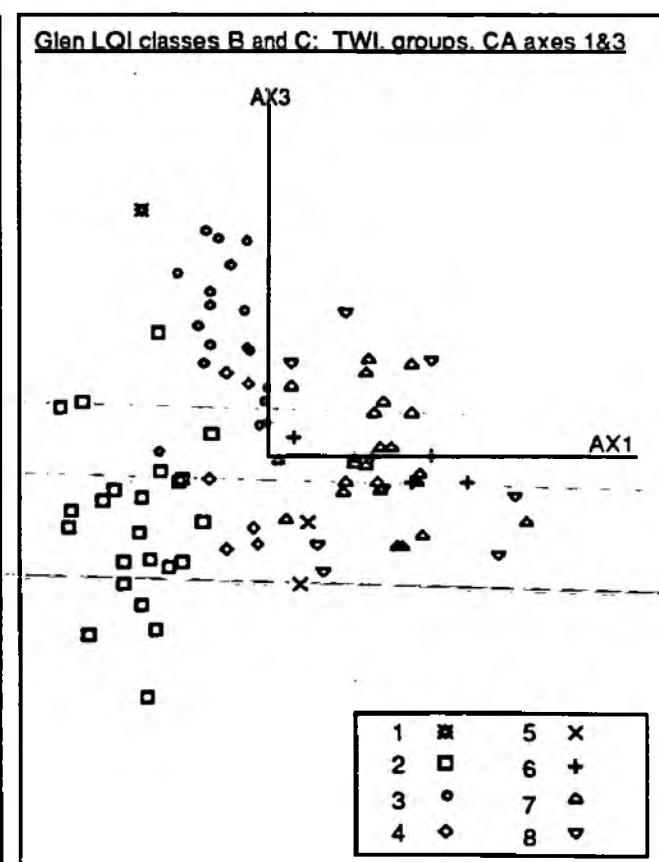
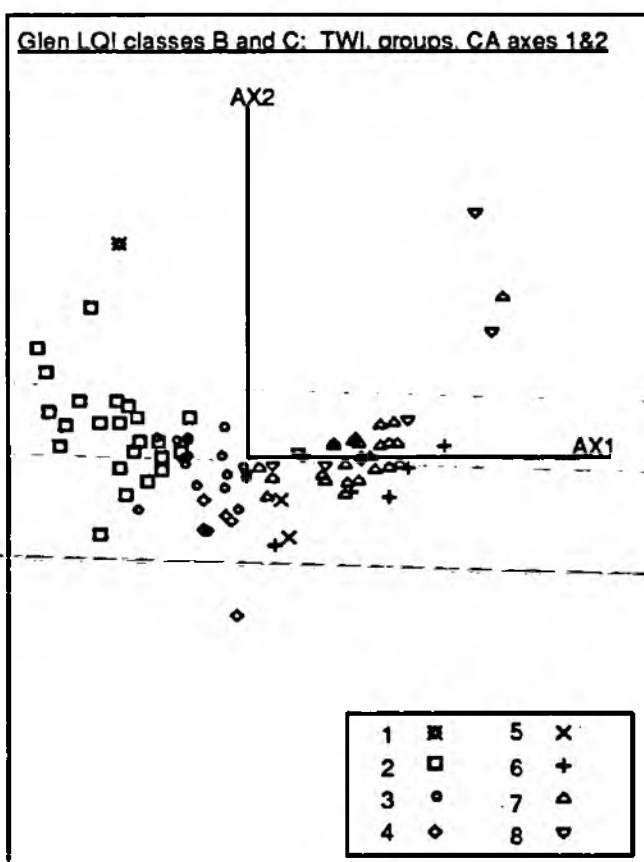
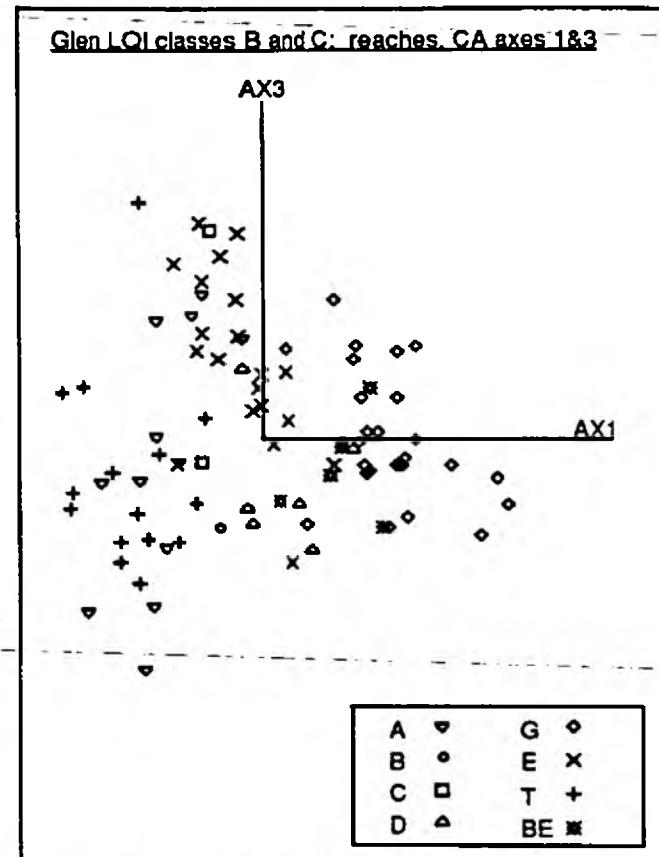
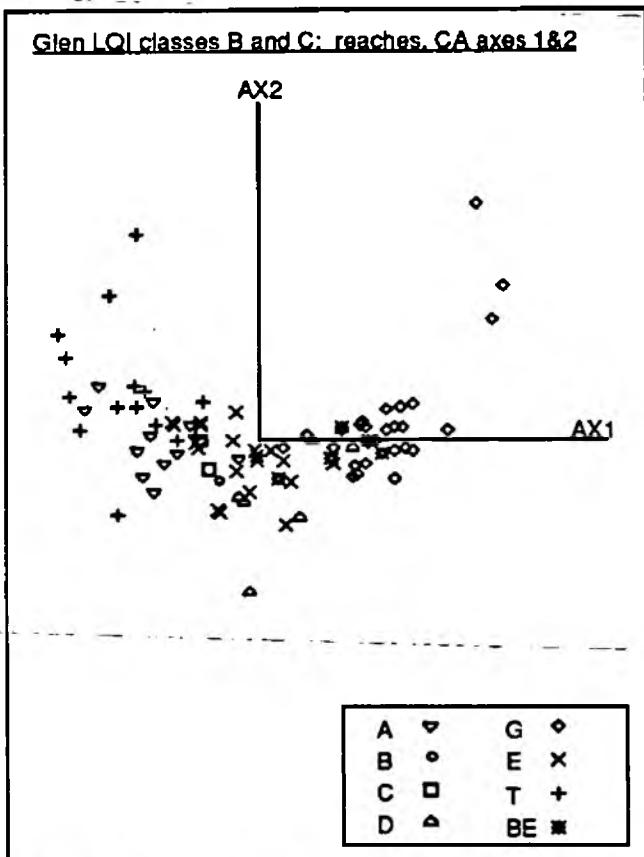


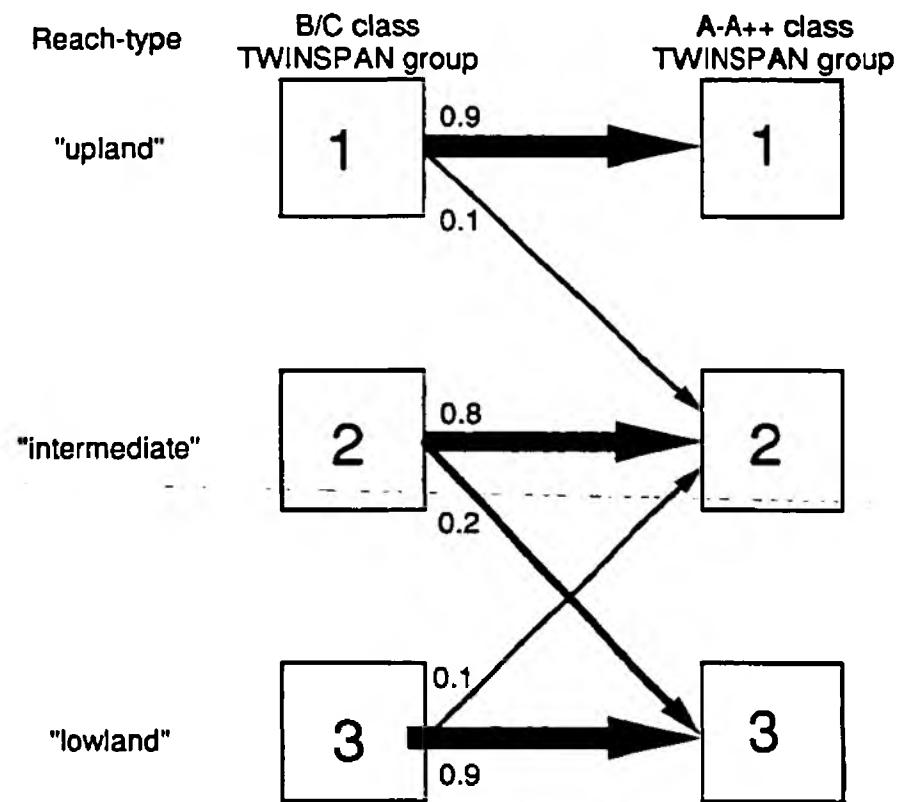
Figure B25. Ordination based on correspondence analysis of family data for samples in LQI classes B and C, axes 1&2 and 1&3, with reaches and TWINSPAN end groups (see figure 21) highlighted.

### 5.8.3 Implications of within LQI class differences

On the evidence of the within-LQI class classifications above it appears that several different types of community can be defined within each class. It may be predicted that a class B site owes its characteristic fauna to a range of environmental conditions, and that were such a site to improve to class A status, due to, for example, an improvement in water quality, the other environmental constraints at that site would determine what type of class A community would develop.

By extracting from the data those class B and C samples from sites which at the next sampling occasion were class A or above, and comparing the TWINSPAN groups into which these fell with the TWINSPAN group of the subsequent class A sample, it was possible to assess whether class B and C communities of one TWINSPAN group always developed into the same type of class A community. The eight original TWINSPAN end groups were re-grouped into three, these being based on the most significant divisions (subjectively defined from the previous classifications - see Sub-sections 5.8.1 and 5.8.2) which broadly represent the "upper", "lower" and "intermediate" type faunas. Figure B26 shows the frequency with which each transition took place. It is clear that there is a strong link between the nature of the communities from the two bands of LQI class. The taxa associated with the transitions are listed in Table B4.

The variability within LQI classes and the reach-determined nature of this variation has clear implications where conservation measures are undertaken not only to improve biological quality but also to encourage a particular kind of invertebrate community or even specific (perhaps rare) taxa. Water quality improvement alone at a class B site may ensure improvement to class A status, but does not guarantee that, for instance, damselflies, which occur frequently in class A sites, will be present. Reference to the fauna of other class A sites within that reach, however, may enable an accurate prediction of the kind of community to be expected (see section 5.4.4).



**Figure B26.** Frequency (numbers by arrows) of transition of LQI class B and C samples of TWINSPAN groups 1-3 to LQI class A-A++ samples of TWINSPAN groups 1-3.

**Table B4. Taxa associated with changes from LQI class B/C to A-A++**

| Tham and middle West Glen<br>(B/C gp.1 > A-A++ gp.3) |           | East Glen and "intermediate" sites<br>(B/C gp.2 > A-A++ gp.2) |           | lower main Glen and Bourne Eau<br>(B/C gp.3 > A-A++ gp.1) |           |
|--|-----------|---|-----------|---|-----------|
| taxa gained  | % samples | taxa gained   | % samples | taxa gained   | % samples |
| HYDRACARINA  | 50        | Piscicolidae  | 80        | Leptoceridae  | 60        |
| Helodidae  | 50        | Dendrocoelidae  | 50        | Dytiscidae  | 40        |
| Leptoceridae   | 50        | COPEPODA  | 50        | Gerridae  | 40        |
| Leptophlebiidae                                      | 50        | Leptophlebiidae   | 50        | Hydrobiidae   | 30        |
| Planariidae  | 30        | Ephemeridae   | 50        | Lymnaeidae  | 30        |
| COPEPODA   | 30        | Sericostomatidae  | 50        |   |           |
| Polycentropidae                                      | 30        | Dytiscidae  | 50        |   |           |
| Limnephilidae  | 30        | Corixidae   | 50        |   |           |
| taxa lost  |           | taxa lost   |           | taxa lost   |           |
| Sialidae   | 50        | Asellidae   | 50        |   |           |
| Hydroptilidae  | 50        | Tipulidae   | 50        |   |           |
|  |           | Hydrobiidae   | 50        |   |           |
|  |           | Planorbidae   | 50        |   |           |
|  |           | Ancylidae   | 50        |   |           |

## 6 CONCLUSIONS

### 6.1 Biological records

The data set of biological records for the River Glen system is probably a good representative of many for similar systems with a history of water quality, and quantity, problems and general conservation interest. It reflects not only changes in the invertebrate communities of the system, but the development of biological monitoring itself. The records begin in the mid-1970's when biological indices were first widely used, and when the severe drought of 1976 caused particular concerns about damage to, and recovery of, biological communities. The most widely used Biological index at the time was the Trent Biotic Index, and the level of identification used in invertebrate monitoring was that needed to derive TBI scores and little more. The later development of the BMWP system, the LQI system and later still the recognition of the importance of seasonally consistent, routine sampling and species level differences demanded improvements in the level of identification and frequency of sampling: all reflected in the detail of the data set.

These long-term changes in the nature of the invertebrate records inevitably effect their interpretation, particularly in two ways: First, where the level of identification varies, for comparisons to be valid only the highest taxonomic groups can be used, and much detail in "better" records is wasted. Secondly, as the motivation for monitoring has changed from response to particular incidents to routine quality assessment, the average quality observed and the sites targeted will have changed.

If long term assessments of biological records are considered of value, then the above problems should be considered especially carefully when either planning a new or altering an old monitoring program.

### 6.2 Site and reach level differences in Invertebrate communities

The invertebrate fauna of the Glen system is typical of lowland-type rivers, being rich in a range of taxa including many insect orders, molluscs and crustaceans. This invertebrate diversity reflects the water chemistry and diversity of substrate, flow characteristics and macrophytes. Classification and ordination of the family level data revealed distinct reach and site level differences in communities, but also great similarities. Representatives of the Trichoptera, Coleoptera, Mollusca and Crustacea were found in the majority of samples, with changes at the family level distinguishing sites as opposed to whole orders or classes as can be found in comparisons between rivers. Reach-scale differences in community were clear, with "upland", "lowland" reaches and an "intermediate" zone around the confluence of the West and East Glens being clearly distinguished. These reach-types defined on the basis of the invertebrate fauna corresponded closely to known physical differences at this scale, for examples sedimentological differences. By the nature of the faunal differences it appeared that water quality was only one of a number of physical attributes varying between sites and reaches.

The analytical tools used to investigate site and reach differences - TWINSPAN and correspondence analysis - proved successful in combination. TWINSPAN is very useful for identifying assemblages of taxa characterising groups of samples and reaches, but is less helpful in assessing the magnitude of between-group differences, especially where these are principally differences relating to presence or absence of taxa. Ordination is essential for comparing the separation of groups and identifying aberrant samples.

From the classification of the three reach-types, lists of taxa associated with each could be produced, which may be regarded as the total communities which any site within those reaches could potentially possess.

### 6.3 Biological scores

Biological scores including LQI proved efficient at identifying within site differences in fauna: In the site case-studies samples of similar score were consistently grouped together; only occasionally did temporal differences override score differences, and on these occasions long term changes in habitat were inferred. At the site and reach level, however, differences were greater between sites and reaches than between score groups, indicating that factors other than water quality were structuring the communities. Within LQI classes, for instance, site and reach differences were as clear as within the whole data set. Also, these between site and reach differences were as great within the lower LQI classes (B and C) as between the highest (A - A++). This is rather surprising as it was assumed that there would be greater uniformity in the lower-quality sites, with a limited range of common taxa. What may in fact be happening is that where the fauna is reduced it is a more specialised community which remains, being more highly adapted to the particular site-type, ie. it implies that specialism and sensitivity to water quality are not necessarily linked.

The diversity within sites of high and moderate quality (as defined by LQI) emphasises the importance of between-site differences in factors other than water quality: Hydrology, hydraulics, substrate, macrophytes and other biotic and abiotic characteristics, in determining the composition of invertebrate communities, and has clear implications for river management.

### 6.4 Long-term changes

Examination of changes in biological scores over the 1976-91 period for the whole data set indicated that biological quality rose from 1976 to the late 1980's, with a slight fall in 1991. More detailed analysis of individual sites proved that these did not all show the same pattern of changes. Whilst most sites followed the pattern described for the whole data set, some showed a continuous increase in scores, some no change and a few a progressive decrease. In addition, changes in BMWP were a result of either changes in the quality (ASPT) or number of taxa, or both. The implication of this is that changes in invertebrate diversity are occurring as a result of factors in addition to changes in water quality. Some long term changes are likely to be due to "structural" changes in the habitat, perhaps related to weed-cutting and dredging regimes, whereas other, shorter-term changes reflect more variable influences, the most significant of these is likely to be hydrology. A further examination of the biological data in conjunction with the hydrological records could therefore be very enlightening.

The five "case-studies" revealed interesting relationships between temporal changes in invertebrate community and biological score. In the case of Banthorpe Lodge on the West Glen the fauna of the 1991 sample closely resembled that of the late 1970's samples, with which it also shared a low score, suggesting that the community responded to drought in a similar manner on both occasions. However, at Surfleet Seas End on the main Glen the relatively high scoring 1979 sample showed more faunal similarities with the other 1970's samples, despite their low-scores, than with the high scoring 1980's samples; while the low scoring 1991 sample resembled the other later samples rather than low scoring 1970's ones. This could be due to saline intrusion during the 1970's period. The Glen at Kates Bridge showed a different pattern again, with the low-scoring 1991 sample resembling faunistically neither the 1980's samples it was close to temporally, nor the 1970's samples it was close to in score. It therefore appears that the fauna of different sites behave differently under stress: Either changing to a new, adapted and quite different community (if, for example, the management regime changes or there is a permanent change in discharge due to abstraction or augmentation), or to a reduced form of their normal community, with only the most sensitive taxa eliminated. Such differences may depend on the nature of the stress, its duration (whether it results in a permanent change in habitat), and also on the nature and source of recolonisation after such a stress event.

## 6.5 Seasonal differences

Evidence for seasonal changes in frequency of collection was found for a number of taxa, predictably families of univoltine insects. Although no significant seasonal changes in BMWP or number of taxa were observed in the whole data set, small but significant changes were observed in ASPT: a distinct trough was observed in September / October. This could be a result of the fact that most high-scoring taxa are insect nymphs and larvae, which do have seasonal peaks and troughs of abundance related to growth, emergence, and egg- or later-stage diapause. However, an attempt to relate numbers of the three orders of insects containing most of these taxa (Plecoptera, Ephemeroptera and Trichoptera) failed to show any significant pattern similar to that of ASPT. The trough in ASPT was later than that expected if the predominance of these taxa were indeed the cause. An alternative possibility is that annual lowest flows, which frequently coincide with this September / October period, are influencing the taxa present and / or collected.

## 7 RECOMMENDATIONS

Analysis of the existing invertebrate records for the River Glen system has revealed that the invertebrate community is responsive to:

- reach-scale habitat conditions, and
- physical and biotic, as well as chemical, variations at sites within the reach-scale constraints.

These findings suggest that there is great potential for expanding the use of biological monitoring beyond the setting and monitoring of water quality objectives (for which it was designed and appears to perform satisfactorily) to assess the potential "best" invertebrate community on a reach-specific basis and set targets if required based on this potential, and by improved and coordinated measurement of a range of habitat factors to monitor the effects of these on the invertebrate community and its ability to achieve the predicted potential. The following strategy (illustrated in Figure B27) is designed to provide the necessary information for such an expanded use of biological monitoring:

### 7.1 Preliminary reach assessment

An initial biological survey of the river system incorporating the maximum number of sites possible and at least three seasonal collections should be carried out in order to identify reach types. Existing invertebrate records could be included in this initial assessment. The choice of sampling sites should aim to include as many habitat types as possible, particularly those subjectively identified as "habitat rich". From this preliminary survey reaches would be defined on the basis of the fauna using a combination of analytical techniques including classification and ordination of sites as described in section 4. For each reach a list of taxa which could potentially be supported in that reach would be compiled.

### 7.2 Selection of routine monitoring sites

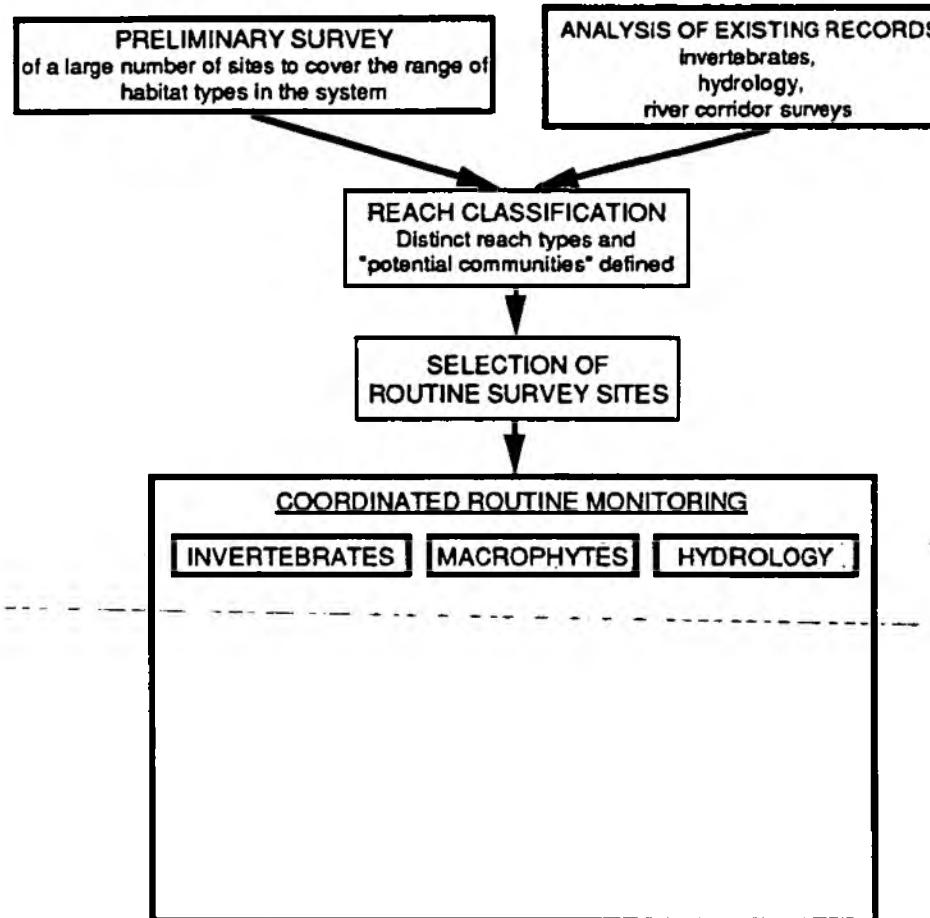
Within each reach a number of routine sites should be selected, including those which fulfil the following criteria (within the normal limits eg. accessibility):

- sites of the highest habitat quality
- sites susceptible to ecological damage (pollution, low flows, any other locally important factors)
- sites reflecting the range of management regimes in the reach

Routine monitoring of these sites should provide a continually-updated picture of the maximum and minimum biological quality of the reach, reflecting natural and anthropogenic influences on the invertebrate community.

### 7.3 Biological monitoring - frequency

Sampling carried out in spring, summer and autumn (as at present) appears to provide adequate seasonal coverage. Emergence of some insect taxa can take place within a few days, making it preferable for all sites within a reach to be sampled over the same few days each year, thus allowing accurate between-site and between-year comparisons.



**Figure B27** Strategy for reach assessment, coordinated biological and hydrological recording and integration of additional information to allow the monitoring of short- and long-term change.

## 7.4 Biological monitoring - field procedure

### **7.4.1 Invertebrate sampling**

The upstream and downstream limits of each site should be clearly defined, and sampling using the kick and sweep method carried out between these limits on each occasion, with the effect that a reduction in wet width will result in a smaller area being sampled. Sampling effort should be divided between micro-habitats in proportion to their area within the site.

### **7.4.2 Invertebrate identification**

If identification to species level is considered desirable and practicable for some taxa then this should be consistently used for these taxa. Varying identification between family and species level necessitates that the species level information be wasted when between-sample comparisons are made. For some taxa, species level identification is easy during one season but impossible during others, in which case separate recording forms might be considered for the spring, summer and autumn samples which would allow between-year if not between season comparisons for these groups.

## 7.5 Associated routine monitoring

In order to relate between-site and between sample changes to the physical and other biotic site characteristics, the assessment of a suite of variables should be performed at the same time as the invertebrate sampling:

- macrophyte cover and type - a record of the percentage cover of the site and dominant species would be more informative than a list of all species presence / absence alone.
- substrate composition - an index of dominant substrate type and amount of surface siltation.
- hydraulic characteristics - coordinating routine current metering surveys with the invertebrate monitoring would provide vital information on site depth and flow velocity distributions, which could be related to discharge and / or water level information from continuous monitoring stations.

## 8 REFERENCES

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**APPENDICES**

**APPENDIX B1. TAXA LISTS FOR ALL SAMPLES**













## Appendix 1.

| TAXON                             | B.BU. | B.BD. | B.RL. | B.TU. | B.TD. |
|-----------------------------------|-------|-------|-------|-------|-------|
| BRYOZOA                           | 11/76 | 10/78 | 10/79 | 09/77 | 10/79 |
| COELENTERATA                      |       |       |       |       |       |
| POLYZOA                           |       |       |       |       |       |
| PORIFERA                          |       |       |       |       |       |
| <i>Dugesia polychroa/lugubris</i> | 1     |       | 1     | 1     | 1     |
| <i>Dugesia tigrina</i>            |       |       |       |       |       |
| <i>Polycais sp.</i>               | 1     | 1     | 1     | 1     | 1     |
| <i>Dendrocoelum lacteum</i>       |       |       | 1     | 1     |       |
| NEMATODA                          | 1     |       |       |       |       |
| NEMATOMORPHA                      |       |       |       |       |       |
| OLIGOCHAETA                       | 1     | 1     | 1     | 1     | 1     |
| Piscicolidae                      | 1     |       |       | 1     | 1     |
| <i>Glossiphonia complanata</i>    | 1     | 1     | 1     | 1     | 1     |
| <i>Glossiphonia heteroclitia</i>  |       |       |       |       |       |
| <i>Helobdella stagnalis</i>       | 1     | 1     | 1     | 1     | 1     |
| <i>Hemiclepsis marginata</i>      |       |       |       |       |       |
| <i>Theromyzon lesselatum</i>      |       |       | 1     |       | 1     |
| Eriopodellidae                    | 1     | 1     | 1     | 1     | 1     |
| OSTRACODA                         |       |       |       | 1     | 1     |
| CLADOCERA                         | 1     |       |       |       |       |
| COPEPODA                          | 1     | 1     | 1     | 1     | 1     |
| <i>Asellus aquaticus</i>          | 1     | 1     | 1     | 1     | 1     |
| <i>Asellus mendianus</i>          |       |       |       |       |       |
| <i>Crangonyx sp.</i>              | 1     | 1     | 1     | 1     | 1     |
| <i>Gammarus sp.</i>               | 1     | 1     | 1     | 1     | 1     |
| <i>Gammarus zaddachi</i>          |       |       |       |       |       |
| Corophidae                        |       |       | 1     |       |       |
| <i>Neomysis integer</i>           |       |       |       |       |       |
| ANISOPTERA                        |       |       |       |       |       |
| Coenagnidiae                      |       |       |       | 1     | 1     |
| <i>Calopteryx splendens</i>       |       |       |       |       |       |
| Nemouridae                        |       |       |       |       |       |
| Luctidae                          |       |       |       |       |       |
| <i>Baetis sp.</i>                 | 1     | 1     | 1     | 1     | 1     |
| <i>Centropitum sp.</i>            |       | 1     |       |       |       |
| <i>Cloeon sp.</i>                 |       |       | 1     | 1     | 1     |
| Leptophlebiidae                   |       |       |       |       |       |
| <i>Ephemerella ignita</i>         |       |       |       |       |       |
| Ephemerae                         |       |       |       |       |       |
| Caenidae                          |       |       |       |       |       |
| <i>Sialis lutaria</i>             |       |       |       |       | 1     |
| Rhyacophilidae                    |       |       |       |       | 1     |
| Polycentropidae                   |       |       |       |       |       |
| Psychomyidae                      |       |       |       |       |       |
| Hydropsychidae                    |       |       |       |       |       |
| Hydroporilidae                    |       | 1     | 1     | 1     | 1     |
| Phryganidae                       |       |       |       |       | 1     |
| Limnephilidae                     |       |       |       |       | 1     |
| Molannidae                        |       |       |       |       | 1     |
| Leptoceridae                      |       |       |       | 1     | 1     |
| Goendae                           |       |       |       |       |       |
| <i>Lepidostoma hirtum</i>         |       |       |       |       |       |
| Lepidosomatidae                   |       |       |       |       |       |
| Sarcostomatidae                   |       |       |       |       |       |
| <i>Agapetus sp.</i>               |       |       |       |       |       |
| <i>Glossosoma sp.</i>             |       |       |       |       |       |
| <i>Nymphula nympheata</i>         |       |       |       |       |       |
| <i>Elmis aenea</i>                |       |       |       |       |       |
| <i>Limnius volkmani</i>           |       |       |       |       |       |
| <i>Oulimnius sp.</i>              |       |       |       |       |       |
| <i>Riokus sp.</i>                 |       |       |       |       |       |
| Halipidae                         |       | 1     | 1     | 1     | 1     |
| Dyusidae                          | 1     |       |       | 1     | 1     |
| Gyrinidae                         |       |       |       |       |       |
| Hydrophilidae                     |       |       |       |       | 1     |
| Heleidae                          |       |       |       |       | 1     |
| Dryopidae                         |       |       |       |       |       |
| Chrysomelidae                     |       |       |       |       |       |
| Mesovelidae                       |       |       |       |       |       |
| Velidae                           |       |       |       |       |       |
| Hydrometridae                     |       |       |       |       | 1     |
| Conidae                           | 1     | 1     | 1     | 1     | 1     |
| Geridae                           |       |       |       |       | 1     |
| Notonectidae                      |       | 1     | 1     |       | 1     |
| Dicranota l.                      | 1     |       |       |       |       |
| Tipulidae l.                      | 1     | 1     | 1     | 1     | 1     |
| Pedicia l.                        |       |       |       |       |       |
| Simuliidae                        |       |       |       |       |       |
| Chironomidae                      | 1     | 1     | 1     | 1     | 1     |
| Ceratopogonidae                   |       | 1     |       |       | 1     |
| <i>Culex sp.</i>                  | 1     |       |       |       |       |
| Dixidae                           |       |       |       |       |       |
| <i>Pericoma sp. l.</i>            |       |       |       |       |       |
| <i>Psychoda sp. l.</i>            |       |       |       |       |       |
| Ptychopteridae                    |       |       |       |       |       |
| Strabomyidae                      |       |       |       |       |       |
| Tabanidae                         |       |       |       |       |       |
| Muscidae                          |       |       |       |       |       |
| <i>Enstalis sp.</i>               |       |       |       |       |       |
| Empididae                         |       |       |       |       |       |
| Unionidae                         |       |       |       |       |       |
| Sphaenidae                        | 1     | 1     | 1     | 1     | 1     |
| Neridae                           |       |       |       |       |       |
| Valvatidae                        |       | 1     | 1     | 1     | 1     |
| <i>Polamopyrgus jenkinsi</i>      |       |       | 1     |       |       |
| <i>Hydrobia sp.</i>               | 1     | 1     | 1     | 1     | 1     |
| <i>Bythnea leachii</i>            |       |       |       |       |       |
| <i>Bythnea tentaculata</i>        |       |       | 1     | 1     | 1     |
| <i>Lymnea auricularia</i>         |       |       |       | 1     | 1     |
| <i>Lymnea glabra</i>              |       |       |       |       | 1     |
| <i>Lymnea palustris</i>           |       |       |       |       | 1     |
| <i>Lymnea peregra</i>             | 1     | 1     | 1     | 1     | 1     |
| <i>Lymnea stagnalis</i>           |       |       |       | 1     | 1     |
| <i>Lymnea truncatula</i>          |       |       |       |       |       |
| Physidae                          | 1     | 1     | 1     | 1     | 1     |
| Planorbidae                       | 1     | 1     | 1     | 1     | 1     |
| Vivipandae                        |       |       |       |       |       |
| <i>Ancylus lluvialis</i>          |       |       |       |       |       |
| <i>Acroloxus lacustris</i>        |       |       |       |       |       |
| Zonitidae                         |       |       |       |       |       |
| Succineidae                       |       |       |       |       |       |
| HYDRACARINA                       | 1     | 1     | 1     |       | 1     |



APPENDIX B2. BIOTIC SCORES FOR ALL SAMPLES

**Appendix 2. RIVER GLEN BIOLOGICAL SCORES**

| SAMPLE      | BMWP | NO.SCRG.TAXA | ASPT | HABITAT TYPE | LQI |
|-------------|------|--------------|------|--------------|-----|
| E.LN.11/88  | 48   | 12           | 4.0  | RR           | D   |
| E.LN.5/89   | 86   | 19           | 4.5  | RR           | A+  |
| E.LN.8/89   | 31   | 9            | 3.4  | RR           | E   |
| E.LN.3/90   | 41   | 12           | 3.4  | RR           | E   |
| E.LN.8/90   | 47   | 13           | 3.6  | RR           | D   |
| E.LN.11/90  | 35   | 10           | 3.5  | RR           | E   |
| E.LN.3/91   | 12   | 5            | 2.4  | RR           | G   |
| E.HW.11/88  | 75   | 18           | 4.2  | R            | D   |
| E.BB.9/77   | 60   | 15           | 4.0  | RR           | C   |
| E.BB.10/78  | 54   | 15           | 3.6  | RR           | C   |
| E.BB.10/79  | 54   | 14           | 3.9  | RR           | C   |
| E.BB.11/82  | 52   | 13           | 4.0  | RR           | C   |
| E.BB.6/83   | 63   | 14           | 4.5  | RR           | A   |
| E.BB.11/83  | 75   | 17           | 4.4  | RR           | B   |
| E.BB.11/88  | 64   | 15           | 4.3  | RR           | B   |
| E.EL.11/88  | 62   | 16           | 3.9  | RR           | C   |
| E.EH.11/76  | 23   | 6            | 3.8  | R            | F   |
| E.EH.10/78  | 86   | 21           | 4.1  | R            | D   |
| E.EH.10/79  | 86   | 20           | 4.3  | R            | D   |
| E.EH.11/83  | 82   | 20           | 4.6  | R            | B   |
| E.EH.8/84   | 97   | 21           | 4.6  | R            | B   |
| E.EH.7/85   | 69   | 18           | 4.3  | R            | D   |
| E.EH.6/87   | 69   | 15           | 4.6  | R            | C   |
| E.EH.11/88  | 82   | 19           | 4.3  | R            | D   |
| E.EH.4/89   | 80   | 18           | 4.4  | R            | D   |
| E.EH.11/89  | 76   | 18           | 4.8  | R            | C   |
| E.EH.2/90   | 81   | 19           | 4.3  | R            | D   |
| E.EH.7/90   | 48   | 13           | 3.7  | R            | E   |
| E.EH.10/90  | 77   | 19           | 4.1  | R            | D   |
| E.EH.3/91   | 78   | 18           | 4.3  | R            | D   |
| E.GB.11/82  | 86   | 19           | 4.5  | R            | D   |
| E.GB.11/89  | 71   | 16           | 4.4  | R            | D   |
| E.GB.7/90   | 48   | 13           | 3.7  | R            | E   |
| E.PH.9/77   | 53   | 14           | 3.8  | RR           | C   |
| E.PH.11/82  | 58   | 14           | 4.1  | RR           | B   |
| E.PH.11/88  | 82   | 19           | 4.3  | RR           | A   |
| E.MT.2/88   | 47   | 12           | 3.9  | R            | E   |
| E.MU.2/88   | 65   | 15           | 4.3  | R            | D   |
| E.MU.7/88   | 53   | 12           | 4.4  | R            | E   |
| E.MD.2/88   | 75   | 15           | 5.0  | R            | C   |
| E.MD.7/88   | 40   | 11           | 3.6  | R            | E   |
| E.MG.11/78  | 45   | 10           | 4.5  | R            | E   |
| E.MG.9/77   | 44   | 12           | 3.7  | R            | E   |
| E.MG.11/82  | 66   | 16           | 4.1  | R            | D   |
| E.MG.11/83  | 81   | 18           | 4.5  | R            | D   |
| E.MG.2/88   | 64   | 14           | 4.6  | R            | C   |
| E.MG.11/88  | 75   | 17           | 4.4  | R            | D   |
| E.BR.10/78  | 97   | 22           | 4.4  | R            | C   |
| E.BR.10/79  | 84   | 20           | 4.2  | R            | D   |
| E.BR.2/83   | 78   | 17           | 4.8  | R            | C   |
| E.BR.11/83  | 121  | 24           | 5.0  | R            | A   |
| E.BR.8/84   | 121  | 23           | 5.3  | R            | A+  |
| E.BR.7/85   | 141  | 28           | 5.0  | R            | A   |
| E.BR.8/86   | 130  | 23           | 5.7  | R            | A++ |
| E.BR.6/87   | 96   | 19           | 5.1  | R            | A   |
| E.BR.8/87   | 110  | 22           | 5.0  | R            | B   |
| E.BR.11/88  | 88   | 19           | 4.9  | R            | C   |
| E.BR.4/89   | 107  | 22           | 4.9  | R            | B   |
| E.BR.11/89  | 61   | 15           | 4.1  | R            | D   |
| E.BR.2/90   | 55   | 15           | 3.7  | R            | E   |
| E.BR.7/90   | 61   | 15           | 4.1  | R            | D   |
| E.BR.10/90  | 66   | 17           | 3.9  | R            | D   |
| E.BR.3/91   | 70   | 16           | 4.4  | R            | D   |
| ET1.1.8/83  | 18   | 3            | 6.0  | R            | B   |
| ET1.4.8/83  | 24   | 5            | 4.8  | R            | E   |
| ET1.4.11/88 | 55   | 13           | 4.2  | R            | E   |

## Appendix 2. RIVER GLEN BIOLOGICAL SCORES

| SAMPLE       | BMWP | NO.SCRG.TAXA | ASPT | HABITAT TYPE | LOI |
|--------------|------|--------------|------|--------------|-----|
| W.BP.11/88   | 62   | 14           | 4.4  | R            | D   |
| W.BF.10/78   | 64   | 20           | 4.2  | R            | D   |
| W.BF.5/79    | 71   | 17           | 4.2  | R            | D   |
| W.BF.8/79    | 68   | 17           | 4.0  | R            | D   |
| W.BF.11/83   | 84   | 18           | 4.7  | R            | C   |
| W.BF.11/88   | 74   | 18           | 4.6  | R            | C   |
| W.BC.11/88   | 76   | 17           | 4.5  | R            | D   |
| W.BC.5/89    | 101  | 22           | 4.8  | R            | B   |
| W.BC.7/89    | 88   | 21           | 4.2  | R            | D   |
| W.BC.3/90    | 82   | 18           | 4.6  | R            | C   |
| W.BC.8/90    | 74   | 18           | 4.1  | R            | D   |
| W.BC.11/90   | 32   | 9            | 3.6  | R            | E   |
| W.BC.3/91    | 56   | 13           | 4.3  | R            | E   |
| W.CG.10/78   | 34   | 10           | 3.4  | R            | F   |
| W.CG.5/79    | 63   | 14           | 4.5  | R            | D   |
| W.CG.8/79    | 39   | 11           | 3.5  | R            | F   |
| W.CG.11/82   | 9    | 4            | 2.3  | R            | I   |
| W.CG.11/83   | 63   | 14           | 4.5  | R            | D   |
| W.CG.11/88   | 6    | 3            | 2.0  | R            | I   |
| W.CG.7/89    | 30   | 8            | 3.8  | R            | F   |
| W.SW.11/88   | 67   | 15           | 4.5  | R            | D   |
| W.SW.7/89    | 46   | 11           | 4.2  | R            | E   |
| W.CR.11/88   | 50   | 12           | 4.2  | R            | E   |
| W.LU.3/85    | 78   | 18           | 4.3  | R            | D   |
| W.LD.11/76   | 73   | 18           | 4.1  | R            | D   |
| W.LD.9/77    | 87   | 19           | 4.6  | R            | C   |
| W.LD.5/79    | 102  | 21           | 4.9  | R            | B   |
| W.LD.8/79    | 84   | 17           | 4.8  | R            | C   |
| W.LD.11/83   | 161  | 30           | 5.4  | R            | A++ |
| W.LD.9/84    | 138  | 29           | 4.8  | R            | A   |
| W.LD.3/85    | 117  | 24           | 4.9  | R            | B   |
| W.LD.7/85    | 112  | 24           | 4.7  | R            | B   |
| W.LD.6/86    | 127  | 22           | 5.8  | R            | A++ |
| W.LD.6/87    | 118  | 22           | 5.4  | R            | A   |
| W.LD.10/88   | 112  | 23           | 4.9  | R            | B   |
| W.LD.11/88   | 113  | 24           | 4.7  | R            | B   |
| W.LD.5/89    | 114  | 22           | 5.2  | R            | A   |
| W.LD.7/89    | 121  | 25           | 4.8  | R            | A   |
| W.LD.3/90    | 113  | 21           | 5.4  | R            | A   |
| W.LD.8/90    | 73   | 18           | 4.6  | R            | C   |
| W.LD.11/90   | 133  | 25           | 5.3  | R            | A+  |
| W.LD.3/91    | 108  | 20           | 5.4  | R            | A   |
| W.EU.11/76   | 95   | 18           | 5.3  | R            | A   |
| W.EU.10/78   | 80   | 18           | 4.4  | R            | D   |
| W.EU.9/77    | 92   | 20           | 4.6  | R            | B   |
| W.EU.10/79   | 105  | 20           | 5.3  | R            | A   |
| W.EU.11/82   | 137  | 25           | 5.5  | R            | A++ |
| W.EU.2/11/83 | 122  | 22           | 5.5  | R            | A++ |
| W.EU.3/11/83 | 144  | 27           | 5.3  | R            | A+  |
| W.ED.11/88   | 147  | 28           | 5.3  | R            | A+  |
| W.BL.11/76   | 34   | 9            | 3.8  | R            | E   |
| W.BL.10/78   | 86   | 21           | 4.1  | R            | D   |
| W.BL.5/79    | 72   | 18           | 4.5  | R            | D   |
| W.BL.8/79    | 70   | 18           | 4.4  | R            | D   |
| W.BL.2/83    | 100  | 22           | 4.5  | R            | C   |
| W.BL.11/83   | 131  | 24           | 5.5  | R            | A++ |
| W.BL.8/84    | 130  | 25           | 5.2  | R            | A+  |
| W.BL.7/85    | 130  | 28           | 5.0  | R            | A   |
| W.BL.6/86    | 162  | 28           | 6.2  | R            | A++ |
| W.BL.8/87    | 127  | 21           | 6.0  | R            | A++ |
| W.BL.11/88   | 161  | 30           | 5.4  | R            | A++ |
| W.BL.8/89    | 112  | 22           | 5.1  | R            | A   |
| W.BL.11/89   | 150  | 25           | 8.0  | R            | A++ |
| W.BL.2/90    | 176  | 32           | 5.5  | R            | A++ |
| W.BL.7/90    | 169  | 32           | 5.3  | R            | A++ |
| W.BL.5/91    | 87   | 18           | 4.8  | R            | C   |

**Appendix 2. RIVER GLEN BIOLOGICAL SCORES**

| SAMPLE     | BMWP | NO.SCRG.TAXA | ASPT | HABITAT TYPE | LQI |
|------------|------|--------------|------|--------------|-----|
| G.KB.6/76  | 72   | 18           | 4.0  | R            | D   |
| G.KB.11/76 | 62   | 18           | 3.9  | R            | D   |
| G.KB.9/77  | 75   | 18           | 4.2  | R            | D   |
| G.KB.10/78 | 94   | 22           | 4.3  | R            | C   |
| G.KB.10/89 | 110  | 23           | 4.6  | R            | B   |
| G.KB.11/82 | 120  | 25           | 4.8  | R            | B   |
| G.KB.11/83 | 128  | 26           | 4.9  | R            | A   |
| G.KB.4/89  | 134  | 27           | 5.0  | R            | A   |
| G.KB.11/89 | 144  | 31           | 4.6  | R            | A   |
| G.KB.2/90  | 137  | 28           | 4.9  | R            | A   |
| G.KB.7/90  | 158  | 32           | 4.9  | R            | A+  |
| G.KB.10/90 | 138  | 30           | 4.6  | R            | A   |
| G.KB.3/91  | 120  | 27           | 4.4  | R            | C   |
| G.TF.11/76 | 17   | 5            | 3.4  | RR           | F   |
| G.TF.9/77  | 34   | 9            | 3.8  | RR           | D   |
| G.TF.10/78 | 75   | 20           | 3.6  | RR           | C   |
| G.TF.10/79 | 121  | 26           | 4.7  | RR           | A++ |
| G.TF.11/83 | 67   | 20           | 4.4  | RR           | A   |
| G.TF.7/85  | 45   | 10           | 4.5  | RR           | B   |
| G.TF.6/86  | 61   | 13           | 4.7  | RR           | A   |
| G.TF.8/87  | 104  | 20           | 5.2  | RR           | A++ |
| G.TE.10/78 | 98   | 22           | 4.5  | RR           | A+  |
| G.TE.5/79  | 80   | 18           | 4.4  | RR           | B   |
| G.TE.8/79  | 99   | 23           | 4.3  | RR           | A   |
| G.TE.11/82 | 82   | 20           | 4.1  | RR           | A   |
| G.TE.11/83 | 62   | 16           | 3.9  | RR           | C   |
| G.TE.7/85  | 84   | 20           | 4.2  | RR           | A   |
| G.TE.6/88  | 50   | 11           | 4.5  | RR           | B   |
| G.TE.8/87  | 85   | 19           | 4.5  | RR           | A+  |
| G.TE.4/89  | 69   | 15           | 4.6  | RR           | A   |
| G.TE.11/89 | 61   | 20           | 4.1  | RR           | B   |
| G.TE.2/90  | 94   | 23           | 4.1  | RR           | A   |
| G.TE.7/90  | 110  | 25           | 4.4  | RR           | A+  |
| G.TE.10/90 | 81   | 19           | 4.3  | RR           | A   |
| G.TE.3/91  | 82   | 16           | 3.9  | RR           | C   |
| G.GG.6/76  | 45   | 10           | 4.5  | RR           | B   |
| G.GG.11/76 | 23   | 6            | 3.8  | RR           | E   |
| G.GG.9/77  | 23   | 6            | 3.8  | RR           | E   |
| G.GG.10/78 | 54   | 13           | 4.2  | RR           | B   |
| G.GG.10/79 | 85   | 21           | 4.0  | RR           | B   |
| G.GG.11/82 | 73   | 19           | 3.8  | RR           | C   |
| G.GG.11/83 | 72   | 18           | 4.0  | RR           | C   |
| G.PK.6/76  | 18   | 5            | 3.6  | RR           | E   |
| G.PK.11/76 | 23   | 5            | 4.6  | RR           | C   |
| G.PK.9/77  | 26   | 7            | 3.7  | RR           | D   |
| G.PK.10/78 | 69   | 17           | 4.1  | RR           | B   |
| G.PK.5/79  | 69   | 17           | 4.1  | RR           | B   |
| G.PK.8/79  | 98   | 22           | 4.4  | RR           | A   |
| G.PK.11/82 | 60   | 16           | 3.8  | RR           | C   |
| G.PK.11/83 | 78   | 18           | 4.2  | RR           | B   |
| G.SF.5/89  | 82   | 19           | 4.3  | RR           | A   |
| G.SF.10/89 | 92   | 21           | 4.4  | RR           | A   |
| G.SF.3/90  | 68   | 17           | 4.0  | RR           | C   |
| G.SF.8/90  | 79   | 20           | 4.0  | RR           | C   |
| G.SF.10/90 | 115  | 25           | 4.6  | RR           | A++ |
| G.SF.4/91  | 71   | 18           | 3.9  | RR           | C   |
| G.SS.6/76  | 9    | 3            | 3.0  | RR           | H   |
| G.SS.11/76 | 8    | 2            | 4.0  | RR           | F   |
| G.SS.9/77  | 26   | 7            | 3.7  | RR           | D   |
| G.SS.10/78 | 41   | 11           | 3.7  | RR           | D   |
| G.SS.10/79 | 61   | 14           | 4.4  | RR           | B   |
| G.SS.11/83 | 50   | 12           | 4.2  | RR           | C   |
| G.SS.7/84  | 58   | 15           | 3.9  | RR           | C   |
| G.SS.8/85  | 63   | 18           | 3.9  | RR           | C   |
| G.SS.8/86  | 37   | 10           | 3.7  | RR           | D   |
| G.SS.6/87  | 66   | 15           | 4.4  | RR           | B   |
| G.SS.5/89  | 67   | 17           | 3.9  | RR           | C   |
| G.SS.10/89 | 79   | 19           | 4.2  | RR           | B   |
| G.SS.3/90  | 93   | 23           | 4.0  | RR           | B   |
| G.SS.8/90  | 97   | 23           | 4.2  | RR           | A   |
| G.SS.10/90 | 94   | 22           | 4.3  | RR           | A   |
| G.SS.4/91  | 70   | 18           | 3.9  | RR           | C   |

**Appendix 2. RIVER GLEN BIOLOGICAL SCORES**

| SAMPLE     | BMWP | NO.SCRG.TAXA | ASPT | HABITAT TYPE | LOI  |
|------------|------|--------------|------|--------------|------|
| T.CH.4/82  | 52   | 10           | 5.2  | R            | C    |
| T.CH.4/84  | 53   | 12           | 4.4  | R            | E    |
| T.CH.4/87  | 50   | 11           | 4.5  | R            | E    |
| T.CU.4/84  | 35   | 9            | 3.9  | R            | E    |
| T.CU.4/87  | 45   | 10           | 4.5  | R            | E    |
| T.CD.4/82  | 87   | 18           | 4.8  | R            | C    |
| T.CD.4/84  | 67   | 15           | 4.5  | R            | D    |
| T.CD.4/87  | 67   | 14           | 4.8  | R            | C    |
| T.GU.4/82  | 90   | 18           | 5.0  | R            | C    |
| T.GU.4/84  | 69   | 15           | 4.6  | R            | C    |
| T.GU.3/85  | 81   | 18           | 5.1  | R            | S    |
| T.GU.3/90  | 111  | 22           | 5.0  | R            | B    |
| T.GU.3/91  | 108  | 20           | 5.4  | R            | A    |
| T.GD.3/85  | 80   | 17           | 4.7  | R            | C    |
| T.LU.10/78 | 83   | 19           | 4.4  | R            | D    |
| T.LU.10/79 | 66   | 14           | 4.7  | R            | C    |
| T.LU.11/83 | 102  | 22           | 4.8  | R            | B    |
| T.LD.9/77  | 54   | 12           | 4.5  | R            | E    |
| T.LD.4/82  | 112  | 21           | 5.3  | R            | A    |
| T.LD.4/84  | 92   | 18           | 5.1  | R            | A    |
| T.LD.3/85  | 64   | 14           | 4.6  | R            | C    |
| T.LD.4/87  | 88   | 17           | 5.2  | R            | B    |
| T.LD.11/88 | 94   | 20           | 4.7  | R            | B    |
| T.LD.5/89  | 102  | 21           | 4.9  | R            | B    |
| T.LD.7/89  | 92   | 21           | 4.4  | R            | C    |
| T.LD.3/90  | 88   | 21           | 4.2  | R            | D    |
| T.LD.8/90  | 91   | 19           | 4.8  | R            | B    |
| T.LD.11/90 | 150  | 29           | 5.2  | R            | A+   |
| T.LD.3/91  | 44   | 11           | 4.0  | R            | E    |
| <br>       | <br> | <br>         | <br> | <br>         | <br> |
| B.BU.11/78 | 17   | 8            | 2.6  | RR           | G    |
| B.BU.10/78 | 52   | 15           | 3.5  | RR           | D    |
| B.BU.10/79 | 47   | 14           | 3.4  | RR           | E    |
| B.BD.9/77  | 33   | 9            | 3.7  | RR           | D    |
| B.BD.10/79 | 51   | 15           | 3.4  | RR           | D    |
| B.BD.11/82 | 63   | 17           | 3.7  | RR           | C    |
| B.BD.11/83 | 55   | 15           | 3.7  | RR           | C    |
| B.RL.10/79 | 61   | 17           | 3.6  | RR           | C    |
| B.RL.11/82 | 58   | 16           | 3.5  | RR           | D    |
| B.RL.2/83  | 36   | 11           | 3.3  | RR           | E    |
| B.RL.11/83 | 52   | 15           | 3.5  | RR           | D    |
| B.TU.4/89  | 72   | 17           | 4.2  | RR           | S    |
| B.TU.11/89 | 66   | 17           | 3.9  | RR           | C    |
| B.TD.2/90  | 101  | 22           | 4.6  | RR           | A++  |
| B.TD.7/90  | 94   | 21           | 4.5  | RR           | A+   |
| B.TD.3/91  | 107  | 24           | 4.5  | RR           | A++  |

**APPENDIX 3 TWINSPAN PREFERENTIAL TAXA**

**APPENDIX 3.1**

**TWINSPAN PREFERENTIAL TAXA FOR THE CLASSIFICATION OF ALL SAMPLES**





### Appendix 3.1

DIVISION 14 (N= 18) I.E. GROUP \*110

Eigenvalue .120 at iteration 2

INDICATORS, together with their SIGN

Lept oph1(+), COPE PODA1(+) Cera topol(-) Pisc icoll(+) -----

#### NEGATIVE PREFERENTIALS

Poly cent1( 2, 0) Cera topol( 7, 0) Psyc hodil( 3, 1) Taes nical( 2, 0) Lynn aeidl( 5, 2)

#### POSITIVE PREFERENTIALS

Pisc icoll( 0, 3) COPE PODA1( 1, 4) Lept oph1( 0, 7) Goe: idael( 0, 2) Seri cost1( 0, 2) Helo didal( 0, 3)

Cori xidal( 1, 4) -----

DIVISION 15 (N= 41) I.E. GROUP \*111

Eigenvalue .101 at iteration 2

INDICATORS, together with their SIGN

Spha erid1(-) Ephe merel(-) Caen idael(-) Dend rocol(-) Asel lidal(-) Erpa bdell(-) Ancy lidal(-)  
Plan orbil(-) Cori xidal(-) Lept ocer1(-) Simu liidl(-) Sial idael(-) Limn ephil(+) Helo didal(+) -----

Lynn aeidl(-)

#### NEGATIVE PREFERENTIALS

Dend rocol( 9, 0) Asel lidal( 10, 2) Ephe merel( 13, 4) Caen idael( 14, 5) Sial idael( 5, 0) Lept ocer1( 6, 0)

Cori xidal( 8, 2) Spha erid1( 16, 6) Plan orbil( 8, 1) -----

#### POSITIVE PREFERENTIALS

Helo didal( 2, 7) -----

## APPENDIX 3.2

### TWINSPAN PREFERENTIAL TAXA FOR THE CLASSIFICATIONS OF THE CASE-STUDY SITES

3.2.1 West Glen at Essendine

3.2.2 West Glen at Banthorpe Lodge

3.2.3 Glen at Kates Bridge

3.2.4 Glen at Surfleet Seas End

3.2.4 East Glen at Braceborough

### Appendix 3.2.1 Indicator and preferential taxa for WEU

TWINSPAN "indicator"-taxa and negative (left-hand groups) and positive (right-hand groups) preferential taxa from the classification of WEU samples (Section 5.6.1, Figure 11). N = the number of samples in the division; numbers after preferential taxa indicate the number of samples in the negative and positive groups in which these taxa occur.

DIVISION 1 (N= 7) I.E. GROUP \*  
Eigenvalue .222 at iteration 2  
INDICATORS, together with their SIGN  
Hydr opayl(-)

#### NEGATIVE PREFERENTIALS

Lept ophill( 3, 0) Ephe merell( 3, 0) Seri costil( 5, 0) Dend rocol( 3, 0) Hydr opayl( 6, 0)  
Dyti scidl( 5, 0) Cori xidal( 6, 0) Noto nectil( 2, 0) Simu liidl( 3, 0) Pisc icoll( 3, 0) Glos siphil( 5, 0)  
Erpo bdeil( 6, 0) Cera topol( 3, 0)

#### POSITIVE PREFERENTIALS

Limn ephil( 2, 1) Hydr optil( 2, 1) Hali plidl( 2, 1) Tipu lidal( 2, 1) Sial idael( 2, 1) Asel lidal( 0, 1)  
Hydr obill( 0, 1) Lynn seidl( 1, 1) Phys idael( 2, 1) COEL ENTEL( 0, 1)  
\*\*\*\*\*

DIVISION 2 (N= 6) I.E. GROUP \*0  
Eigenvalue .211 at iteration 1  
INDICATORS, together with their SIGN  
Dend rocol(+)

#### NEGATIVE PREFERENTIALS

Seri costil( 2, 1) Gyri nidal( 1, 0) Hydr ophill( 1, 0) Tipu lidal( 1, 1) Sial idael( 2, 0) COPE PODAI( 1, 0)

#### POSITIVE PREFERENTIALS

Lept ophill( 0, 3) Mola nnidl( 0, 1) Rhya cophil( 0, 1) Limn ephil( 0, 2) Hydr optil( 0, 2) Plan ariil( 1, 4)  
Dend rocol( 0, 4) Hali plidl( 0, 2) Noto nectil( 0, 2) Simu liidl( 0, 3) Pisc icoll( 0, 3) Lynn seidl( 0, 1)  
Phys idael( 0, 2) Plan'orbil( -0,-4)-PORI-FERA( -0,-1)-HYDR-ACARI( -1,-4)  
\*\*\*\*\*

DIVISION 4 (N= 2) I.E. GROUP \*00  
Eigenvalue .219 at iteration 0  
INDICATORS, together with their SIGN  
Ephe merel(-)

#### NEGATIVE PREFERENTIALS

Ephe merel( 1, 0) Plan ariil( 1, 0) Hydr ophill( 1, 0) Tipu lidal( 1, 0)

#### POSITIVE PREFERENTIALS

Lept oceril( 0, 1) Gyri nidal( 0, 1) COPE PODAI( 0, 1) Cera topol( 0, 1) HYDR ACARI( 0, 1)  
\*\*\*\*\*

DIVISION 5 (N= 4) I.E. GROUP \*01  
Eigenvalue .145 at iteration 6  
RA TROUBLE 6 ITERATIONS, AND RESIDUAL IS STILL .026 INSTEAD OF .003 (THE TOLERANCE)  
INDICATORS, together with their SIGN  
Lept ophill(-)

#### NEGATIVE PREFERENTIALS

Lept ophill( 3, 0) Ephe merel( 2, 0) Ephe merill( 3, 0) Mola nnidl( 1, 0) Lept oceril( 3, 0) Seri costil( 1, 0)  
Rhya cophil( 1, 0) Limn ephil( 2, 0) Noto nectil( 2, 0) Tipu lidal( 1, 0) Simu liidl( 3, 0) Pisc icoll( 3, 0)  
Lynn seidl( 1, 0) Phys idael( 2, 0) PORI FERA( 1, 0) Cera topol( 2, 0)

#### POSITIVE PREFERENTIALS

Hydr optil( 1, 1) Hali plidl( 1, 1)  
\*\*\*\*\*

DIVISION - 10- (N= 3) I.E. GROUP \*010-  
Eigenvalue .178 at iteration 6

RA TROUBLE 6 ITERATIONS, AND RESIDUAL IS STILL .045 INSTEAD OF .003 (THE TOLERANCE)

INDICATORS, together with their SIGN

Ephe merel(+)

#### NEGATIVE PREFERENTIALS

Mola nnidl( 1, 0) Seri costil( 1, 0) Noto nectil( 1, 1) Glos siphil( 1, 1) Phys idael( 1, 1) Cera topol( 1, 1)

#### POSITIVE PREFERENTIALS

Ephe merel( 0, 2) Rhya cophil( 0, 1) Poly centil( 0, 2) Limn ephil( 0, 2) Hydr optil( 0, 1) Hali plidl( 0, 1)  
Dyti scidl( 0, 2) Tipu lidal( 0, 1) Lynn seidl( 0, 1) Chir onoml( 0, 2) PORI FERA( 0, 1)  
\*\*\*\*\*



## Appendix 3.2.2

DIVISION 9 (N= 5) I.E. GROUP \*001

Eigenvalue .198 at iteration 1  
INDICATORS, together with their SIGN  
Simu liidl(-)

### NEGATIVE PREFERENTIALS

Plan ariil(- 2, 1) HYDR ACARI( 2, 1) Simu liidl( 2, 0) Cera topo( 2, 0) Hydr optil( 1, 0) Psyc hodil( 1, 0)  
Mole nnidl( 1, 0) Helo didal( 1, 0) Coen agril( 1, 0) Noto nectil( 1, 0)

### POSITIVE PREFERENTIALS

Ephe merel( 0, 2) Ephe meril( 1, 3) Seri costil( 1, 3) Poly centil( 0, 1) Lynn seidil( 0, 3) Rhys cophil( 0, 1)  
Pisc icoll( 0, 2) Lepi dosil( 0, 2) OSTR ACODil( 0, 1) Gyri nidal( 0, 1)  
\*\*\*\*\*

DIVISION 10 (N= 3) I.E. GROUP \*010

Eigenvalue .269 at iteration 2  
INDICATORS, together with their SIGN  
Plan orbil(-)

### NEGATIVE PREFERENTIALS

Plan orbil( 1, 0) Dyti scidil( 1, 1) Glos siphil( 1, 0) Spha eridil( 1, 1) Hydr obili( 1, 0) Hali plidl( 1, 1)  
Hydr opayl( 1, 1) Cori xidal( 1, 0) Seri costil( 1, 1) Lynn seidil( 1, 0) Hydr optil( 1, 1) Gyri nidal( 1, 0)

### POSITIVE PREFERENTIALS

Plan ariil( 0, 2) HYDR ACARI( 0, 1) Goer idael( 0, 2) Ephe meril( 0, 1) Lept oceril( 0, 2) Lept ophil( 0, 1)  
Sial idael( 0, 1) Ancy lidel( 0, 1) Lepi dosil( 0, 1) Leuc tridil( 0, 1) Glos mosol( 0, 1)  
\*\*\*\*\*

DIVISION 11 (N= 2) I.E. GROUP \*011

Eigenvalue .236 at iteration 0  
INDICATORS, together with their SIGN  
Plan orbil(-)

### NEGATIVE PREFERENTIALS

Plan orbil( 1, 0) Hydr obili( 1, 0) Simu liidl( 1, 0) Ephe merel( 1, 0) Sial idael( 1, 0) Veli idael( 1, 0)

### POSITIVE PREFERENTIALS

Lynn ephil( 0, 1) Cori xidal( 0, 1) Phys idael( 0, 1) Hydr optil( 0, 1) Lept ophil( 0, 1) COPE PODAl( 0, 1)  
PORI FERAL( 0, 1)  
\*\*\*\*\*

DIVISION 12 (N= 3) I.E. GROUP \*100

Eigenvalue .362 at iteration 2  
INDICATORS, together with their SIGN  
Caen idael(+)

### NEGATIVE PREFERENTIALS

Tipu lidal(- 1, 1) Ephe merel( 1, 1) Hydr opayl( 1, 0) Dend rocoll( 1, 0) Rhys cophil( 1, 0) Asel lidal( 1, 0)  
Hydr ophil( 1, 0)

### POSITIVE PREFERENTIALS

Caen idael( 0, 2) Dyti scidil( 0, 1) Spha eridil( 0, 1) Plan ariil( 0, 2) Hali plidl( 0, 1) Simu liidl( 0, 2)  
Lynn ephil( 0, 1) Lept oceril( 0, 1) Hydr optil( 0, 1) COPE PODAl( 0, 1) Psyc hodil( 0, 1) Stra tiomil( 0, 2)  
Musc idael( 0, 1) Glos mosol( 0, 1) Taba nidal( 0, 1) Nemo uridil( 0, 1)  
\*\*\*\*\*

### Appendix 3.2.3 Indicator and preferential taxa for GKB

TWINSPAN "indicator" taxa and negative (left-hand groups) and positive (right-hand groups) preferential taxa from the classification of GKB samples (Section 5.6.3, Figure 13). N = the number of samples in the division; numbers after preferential taxa indicate the number of samples in the negative and positive groups in which these taxa occur.

DIVISION 1 (N= 13) I.E. GROUP \*

Eigenvalue .118 at iteration 2

INDICATORS, together with their SIGN

Simu liid1(-) Gyri nidal(+)

NEGATIVE PREFERENTIALS

Lept ophili( 2, 0) Goer idael( 2, 0) Psyc homyl( 2, 0) Simu liid1( 8, 0) PORI FERAL( 3, 0)

POSITIVE PREFERENTIALS

Ephe merel( 1, 1) Ephe meril( 2, 2) Mola nnid1( 0, 1) Coen agril( 2, 2) Gyri nidal( 0, 3) Veli idael( 0, 1)  
Hydr omet1( 0, 2) Gerr idael( 0, 2) Noto nect1( 1, 3) Sial lidael( 4, 3) Valv atid1( 3, 2) Musc idael( 3, 2)

\*\*\*\*\*

DIVISION 2 (N= 10) I.E. GROUP \*0

Eigenvalue .129 at iteration 1

INDICATORS, together with their SIGN

Sial idael(-)

NEGATIVE PREFERENTIALS

Lept ophili( 2, 0) Ephe meril( 2, 0) Lept oceri( 4, 3) Agri idael( 1, 0) Limn ephil( 3, 2) Dend rocol( 4, 2)  
Elmi nthili( 4, 3) Helo didal( 1, 0) Tipu lidal( 4, 3) Sial lidael( 4, 0) Valv atid1( 2, 1) NEMA TOMOL( 1, 0)  
Psyc hodil( 1, 0) Stra tioml( 1, 0) Musc idael( 2, 1)

POSITIVE PREFERENTIALS

Goer idael( 0, 2) Pisc icoll( 1, 3) Lynn meidl( 2, 6) PORI FERAL( 0, 3) CLAD OCER1( 1, 4) COPE PODAL( 1, 3)

\*\*\*\*\*

DIVISION 3 (N= 3) I.E. GROUP \*1

Eigenvalue .076 at iteration 0

INDICATORS, together with their SIGN

Ephe meril(-)

NEGATIVE PREFERENTIALS

Ephe merel( 1, 0) Ephe meril( 2, 0) Poly cent1( 2, 0) Limn ephil( 1, 0) Hydr opayl( 2, 0) Veli idael( 1, 0)  
Pisc icoll( 1, 0) COPE PODAL( 1, 0)

POSITIVE PREFERENTIALS

Mola nnid1( 0, 1) Coen agril( 1, 1) Hydr omet1( 1, 1) Gerr idael( 1, 1) Tipu lidal( 1, 1) Valv atid1( 1, 1)  
OSTR ACOD1( 1, 1) CLAD OCER1( 1, 1) Cera topol( 1, 1) Musc idael( 1, 1) HYDR ACARI( 1, 1)

\*\*\*\*\*

DIVISION 4 (N= 4) I.E. GROUP \*00

Eigenvalue .093 at iteration 6

RA TROUBLE 6 ITERATIONS, AND RESIDUAL IS STILL .026 INSTEAD OF .003 (THE TOLERANCE)

INDICATORS, together with their SIGN

Hydr optil(+)

NEGATIVE PREFERENTIALS

Lept ophili( 1, 1) Ephe meril( 1, 1) Poly cent1( 1, 1) Hydr opayl( 1, 1) COPE PODAL( 1, 0)

POSITIVE PREFERENTIALS

Agri idael( 0, 1) Psyc homyl( 0, 1) Coen agril( 0, 1) Hydr optil( 0, 3) Helo didal( 0, 1) Pisc icoll( 0, 1)  
Valv atid1( 0, 2) Lynn meidl( 0, 2) NEMA TOMOL( 0, 1) CLAD OCER1( 0, 1) Psyc hodil( 0, 1) Stra tioml( 0, 1)  
Musc idael( 0, 2)

\*\*\*\*\*

DIVISION 5 (N= 6) I.E. GROUP \*01

Eigenvalue .191 at iteration 1

INDICATORS, together with their SIGN

Simu liid1(-)

NEGATIVE PREFERENTIALS

Ephe merel( 1, 0) Lept oceri( 3, 0) Goer idael( 2, 0) Psyc homyl( 1, 0) Poly cent1( 3, 0) Gamm arid1( 4, 1)  
Hydr optil( 3, 0) Hydr opayl( -2, 0) Elmi nthili( -3, 0) Simu liid1( -4, 0) Pisc icoll( 3, 0) Glos siphil( 4, 1)  
Phys idael( 4, 0) Plan orbil( 4, 1) Musc idael( 1, 0)

POSITIVE PREFERENTIALS

Limn ephil( 1, 1) Coen agril( 0, 1) Dend rocol( 1, 1) Hall plid1( 2, 2) Noto nect1( 0, 1) Valv atid1( 0, 1)  
OSTR ACOD1( 2, 2) CLAD OCER1( 2, 2) Cera topol( 2, 2)

\*\*\*\*\*

DIVISION 6 (N= 2) I.E. GROUP \*10

Eigenvalue .213 at iteration 0

INDICATORS, together with their SIGN

Ephe merel(+)

NEGATIVE PREFERENTIALS

Coen agril( 1, 0) Tipu lidal( 1, 0) Pisc icoll( 1, 0) Valv atid1( 1, 0) OSTR ACOD1( 1, 0) CLAD OCER1( 1, 0)  
COPE PODAL( 1, 0) Cera topol( 1, 0) Musc idael( 1, 0) HYDR ACARI( 1, 0)

POSITIVE PREFERENTIALS

Ephe merel( 0, 1) Limn ephil( 0, 1) Veli idael( 0, 1) Hydr omet1( 0, 1) Gerr idael( 0, 1)

\*\*\*\*\*

### Appendix 3.2.3

DIVISION 9 (N= 3) I.E. GROUP \*001

Eigenvalue .212 at iteration 6  
RA TROUBLE 6 ITERATIONS, AND RESIDUAL IS STILL .016 INSTEAD OF .003 (THE TOLERANCE)  
INDICATORS, together with their SIGN  
Coen agril(-)

NEGATIVE PREFERENTIALS

Limn ephili( 1, 1) Coen agril( 1, 0) Hydr opayli( 1, 0) Dytii acidli( 1, 1) Helo didal( 1, 0) Pisc icoll( 1, 0)  
Spha eridli( 1, 1) Valv stidi( 1, 1) Hydr obili( 1, 1) Lynn aeidl( 1, 1) Psyc hodli( 1, 0) Stra tiomi( 1, 0)  
Musc idael( 1, 1)

POSITIVE PREFERENTIALS

Lept ophili( 0, 1) Ephe merili( 0, 1) Agri idael( 0, 1) Psyc homyl( 0, 1) Poly centli( 0, 1) Hali plidli( 0, 2)  
Cori xidal( 0, 2) Baet idael( 0, 2) NEMA TOMOL( 0, 1) CLAD OCERI( 0, 1)  
\*\*\*\*\*

DIVISION 10 (N= 4) I.E. GROUP \*010

Eigenvalue .092 at iteration 6  
RA TROUBLE 6 ITERATIONS, AND RESIDUAL IS STILL .019 INSTEAD OF .003 (THE TOLERANCE)  
INDICATORS, together with their SIGN  
PORI FERAL(-)

NEGATIVE PREFERENTIALS

Lept ocerli( 2, 1) Psyc homyl( 1, 0) Poly centli( 2, 1) Myar optili( 2, 1) Dend rocol( 1, 0) Erpo bdeli( 2, 1)  
Spha eridli( 2, 1) Chir onomi( 2, 1) PORI FERAL( 2, 0) Musc idael( 1, 0)

POSITIVE PREFERENTIALS

Ephe mereli( 0, 1) Caen idael( 1, 2) Limn ephili( 0, 1) Elmi nthili( 1, 2) Pisc icoll( 1, 2)  
\*\*\*\*\*

DIVISION 11 (N= 2) I.E. GROUP \*011

Eigenvalue .267 at iteration 0  
INDICATORS, together with their SIGN  
Caen idael(-)

NEGATIVE PREFERENTIALS

Caen idael( 1, 0) Limn ephili( 1, 0) Dend rocol( 1, 0) Noto necili( 1, 0) Glos siphli( 1, 0) Plan orbili( 1, 0)  
PORI FERAL( 1, 0) COPE PODAL( 1, 0)

POSITIVE PREFERENTIALS

Gamm aridli( 0, 1) Coen agril( 0, 1) Tipu lidal( 0, 1) Valv stidi( 0, 1)  
\*\*\*\*\*



### Appendix 3.2.4

**DIVISION 6 (N= 5) I.E. GROUP \*10**  
 Eigenvalue .285 at iteration 1  
 INDICATORS, together with their SIGN  
 Pisc icoll(-)  
  
**NEGATIVE PREFERENTIALS**  
 Hali plid1( 1, 1) Dyti scidl( 1, 1) Cori xidal( 1, 2) Pisc icoll( 1, 0) Baet idael( 1, 2) Hydr obili( 1, 1)  
 Lynn seidl( 1, 1) Plan orbil( 1, 2) OSTR ACOD1( 1, 0) CLAD OCER1( 1, 2) Hali plid2( 1, 1) Dyti scid2( 1, 1)  
 Cori xida2( 1, 2) Pisc icoll2( 1, 0) Baet idae2( 1, 2) Hydr obili2( 1, 1) Lynn seid2( 1, 0) Plan orbil( 1, 2)  
 OSTR ACOD2( 1, 0) CLAD OCER2( 1, 2)

**POSITIVE PREFERENTIALS**  
 Gerr idael( 0, 1) Glos siphil( 0, 1) COPE PODA1( 0, 2) HYDR ACAR1( 0, 3) Gerr idae2( 0, 1) Glos siph2( 0, 1)  
 COPE PODA2( 0, 2) HYDR ACAR2( 0, 3)  
\*\*\*\*\*

**DIVISION 7 (N= 2) I.E. GROUP \*11**  
 Eigenvalue .318 at iteration 0  
 INDICATORS, together with their SIGN  
 Poly cent1(+)  
  
**NEGATIVE PREFERENTIALS**  
 Plan ariil( 1, 0) Gerr idael( 1, 0) Valv stid1( 1, 0) Plan ariil2( 1, 0) Gerr idae2( 1, 0) Valv stid2( 1, 0)

**POSITIVE PREFERENTIALS**  
 Poly cent1( 0, 1) Hydr optil( 0, 1) Dyti scidl( 0, 1) Erpo bdel( 0, 1) Asel lidal( 0, 1) OSTR ACOD1( 0, 1)  
 Cera topol( 0, 1) HYDR ACAR1( 0, 1) Poly cent2( 0, 1) Hydr opti2( 0, 1) Dyti scid2( 0, 1) Erpo bdel2( 0, 1)  
 Asel lida2( 0, 1) OSTR ACOD2( 0, 1) Cera topo2( 0, 1) HYDR ACAR2( 0, 1)  
\*\*\*\*\*

**DIVISION 8 (N= 3) I.E. GROUP \*001**  
 Eigenvalue .255 at iteration 4  
 INDICATORS, together with their SIGN  
 Caen idael(-)  
  
**NEGATIVE PREFERENTIALS**  
 Caen idael( 1, 0) Dend rocol( 1, 0) Elmi nthil( 1, 1) Dyti scidl( 1, 1) Helo didal( 1, 0) Noto nect1( 1, 0)  
 Lynn seidl( 1, 1) Coro phidl( 1, 0) Caen idae2( 1, 0) Dend roco2( 1, 0) Elmi nthi2( 1, 1) Dyti scid2( 1, 1)  
 Helo dida2( 1, 0) Noto nect2( 1, 0) Lynn seid2( 1, 1) Coro phid2( 1, 0)

**POSITIVE PREFERENTIALS**  
 Psyc homyl( 0, 1) Rhya cophl( 0, 2) Hydr optil( 0, 1) Unio nidal( 0, 1) Pisc icoll( 0, 1) OSTR ACOD1( 0, 2)  
 COPE PODA1( 0, 1) Mysl dae1( 0, 1) Pyra lidal( 0, 1) Cera topol( 0, 1) HYDR ACAR1( 0, 2) Psyc homy2( 0, 1)  
 Rhya coph2( 0, 2) Hydr opti2( 0, 1) Unio nida2( 0, 1) Pisc icoll2( 0, 1) OSTR ACOD2( 0, 2) COPE PODA2( 0, 1)  
 Mysl dae2( 0, 1) Pyra lidai2( 0, 1) Cera topo2( 0, 1) HYDR ACAR2( 0, 2)  
\*\*\*\*\*

**DIVISION 10 (N= 3) I.E. GROUP \*010**  
 Eigenvalue .254 at iteration 4  
 INDICATORS, together with their SIGN  
 Hydr ophill(-)  
  
**NEGATIVE PREFERENTIALS**  
 Lept ocerl( 1, 1) Gamm aridl( 1, 1) Plan ariil( 1, 1) Elmi nthil( 1, 1) Hali plid1( 1, 1) Hydr ophill( 1, 0)  
 Gerr idael( 1, 1) Noto nect1( 1, 1) Tipu lidal( 1, 0) Pisc icoll( 1, 1) Glos siphil( 1, 1) Erpo bdel1( 1, 0)  
 Asel lidal( 1, 0) OSTR ACOD1( 1, 0) Cera topol( 1, 0) Lept ocer2( 1, 1) Gamm arid2( 1, 1) Plan ariil2( 1, 1)  
 Elmi nthi2( 1, 1) Hali plid2( 1, 1) Hydr ophii2( 1, 0) Gerr idae2( 1, 1) Noto nect2( 1, 1) Tipu lida2( 1, 0)  
 Pisc icoll2( 1, 1) Glos siph2( 1, 1) Erpo bdel2( 1, 0) Asel lida2( 1, 0) OSTR ACOD2( 1, 0) Cera topo2( 1, 0)

**POSITIVE PREFERENTIALS**  
 Unio nidal( 0, 1) Meao velil( 0, 1) Cori xidal( 0, 2) CLAD OCER1( 0, 1) HYDR ACAR1( 0, 2) Unio nida2( 0, 1)  
 Meao veli2( 0, 1) Cori xida2( 0, 2) CLAD OCER2( 0, 1) HYDR ACAR2( 0, 2)  
\*\*\*\*\*

**DIVISION 11 (N= 2) I.E. GROUP \*011**  
 Eigenvalue .148 at iteration 0  
 INDICATORS, together with their SIGN  
 Caen idael(+)  
  
**NEGATIVE PREFERENTIALS**  
 Cera topol( 1, 0) Cera topo2( 1, 0)

**POSITIVE PREFERENTIALS**  
 Caen idael( 0, 1) Hydr optil( 0, 1) Tipu lidal( 0, 1) Asel lidal( 0, 1) Mysl dae1( 0, 1) Caen idae2( 0, 1)  
 Hydr opti2( 0, 1) Tipu lida2( 0, 1) Asel lida2( 0, 1) Mysl dae2( 0, 1)

**DIVISION 13 (N= 4) I.E. GROUP \*101**  
 Eigenvalue .334 at iteration 6  
 RA TROUBLE 6 ITERATIONS, AND RESIDUAL IS STILL .036 INSTEAD OF .003 (THE TOLERANCE)  
 INDICATORS, together with their SIGN  
 Cori xidal(+)  
**NEGATIVE PREFERENTIALS**  
  
**POSITIVE PREFERENTIALS**  
 Hali plid1( 0, 1) Dyti scidl( 0, 1) Cori xidal( 0, 2) Gerr idael( 0, 1) Baet idael( 0, 2) Glos siphil( 0, 1)  
 Hydr obili( 0, 1) Plan orbil( 0, 2) Olig ochal( 1, 2) COPE PODA1( 0, 2) HYDR ACAR1( 1, 2) Hali plid2( 0, 1)  
 Dyti acid2( 0, 1) Cori xida2( 0, 2) Gerr idae2( 0, 1) Baet idae2( 0, 2) Glos siph2( 0, 1) Hydr obili2( 0, 1)  
 Plan orbil2( 0, 2) Olig ochal2( 1, 2) COPE PODA2( 0, 2) HYDR ACAR2( 1, 2)  
\*\*\*\*\*





### Appendix 3.2.5

DIVISION 12 (N= 2) I.E. GROUP \*100

Eigenvalue .168 at iteration 0  
INDICATORS, together with their SIGN  
Plan ariil(+)

#### NEGATIVE PREFERENTIALS

OSTR ACOD1( 1, 0) CLAD OCER1( 1, 0) Goer idae1( 1, 0) OSTR ACOD2( 1, 0) CLAD OCER2( 1, 0) Goer idae2( 1, 0)

#### POSITIVE PREFERENTIALS

Plan ariil( 0, 1) COPE PODA1( 0, 1) Asel lidal( 0, 1) Limn ephil( 0, 1) Hydr ophi1( 0, 1) Gerr idae1( 0, 1)  
Simu liid1( 0, 1) Plan ariil2( 0, 1) COPE PODA2( 0, 1) Asel lidal2( 0, 1) Limn ephi2( 0, 1) Hydr ophi2( 0, 1)  
Gerr idae2( 0, 1) Simu liid2( 0, 1)

DIVISION 13 (N= 2) I.E. GROUP \*101

Eigenvalue .316 at iteration 0  
INDICATORS, together with their SIGN  
Glos siph1(+)

#### NEGATIVE PREFERENTIALS

Ephe merel( 1, 0) Ephe meril( 1, 0) Caen idae1( 1, 0) Poly cent1( 1, 0) Limn ephil( 1, 0) Goer idae1( 1, 0)  
Elmi nthil( 1, 0) Tipu lidal( 1, 0) Simu liid1( 1, 0) Hydr obili1( 1, 0) Ephe mere2( 1, 0) Ephe meriz( 1, 0)  
Caen idae2( 1, 0) Poly cent2( 1, 0) Limn ephi2( 1, 0) Goer idae2( 1, 0) Elmi nthil2( 1, 0) Tipu lidaz( 1, 0)  
Simu liid2( 1, 0) Hydr obili2( 1, 0)

#### POSITIVE PREFERENTIALS

Glos siph1( 0, 1) OSTR ACOD1( 0, 1) CLAD OCER1( 0, 1) Coen agril( 0, 1) Hydr ophi1( 0, 1) Ptyc hopt1( 0, 1)  
Glos siph2( 0, 1) OSTR ACOD2( 0, 1) CLAD OCER2( 0, 1) Coen agril2( 0, 1) Hydr ophi2( 0, 1) Ptyc hopt2( 0, 1)

## **APPENDIX 3.3**

### **TWINSPAN PREFERENTIAL TAXA FOR THE CLASSIFICATION OF SAMPLES WITHIN LOI CLASSES A - A++ AND B &C.**

**3.3.1 LQI classes A - A++**

**3.3.2 LQI classes B & C**



### Appendix 3.3.1

DIVISION 6 (N= 24) I.E. GROUP \*10  
Eigenvalue .097 at iteration 2  
INDICATORS, together with their SIGN  
Seri cost1(+) Ephe meril(+)

NEGATIVE PREFERENTIALS  
CLAD OCER1( 2, 0) Leuc trid1( 2, 1) Ephe meril( 5, 9) Hydr optil( 3, 5) Gerr idael( 2, 2) Noto nect1( 3, 5)  
CLAD OCER2( 2, 0) Leuc trid2( 2, 1) Ephe mere2( 5, 9) Hydr opt2( 3, 5) Gerr idae2( 2, 2) Noto nect2( 3, 5)

POSITIVE PREFERENTIALS  
Dend roco1( 0, 11) Pisc icoll( 1, 11) Ephe meril( 0, 17) Sial idael( 0, 10) Hydr opay1( 2, 16) Lept ocer1( 1, 15)  
Seri cost1( 0, 12) Gyri nidal( 0, 5) Phys idael( 1, 10) Dend roco2( 0, 11) Pisc icoll2( 1, 11) Ephe meril2( 0, 17)  
Sial idae2( 0, 10) Hydr opsy2( 2, 16) Lept ocer2( 1, 15) Seri cost2( 0, 12) Gyri nida2( 0, 5) Phys idae2( 1, 10)  
\*\*\*\*\*

DIVISION 7 (N= 15) I.E. GROUP \*11  
Eigenvalue .119 at iteration 2  
INDICATORS, together with their SIGN  
Cori xidal(-)

NEGATIVE PREFERENTIALS  
Dend roco1( 3, 3) Pisc icoll( 4, 4) COPE PODA1( 3, 1) Asel lida1( 4, 2) Ephe meril( 2, 1) Sial idael( 2, 1)  
Seri cost1( 3, 3) Hali plid1( 5, 2) Gyri nidal( 2, 0) Cori xidal( 5, 0) Spha erid1( 5, 5) Lynm aeidl( 5, 2)  
Plan orb1( 5, 2) Dend roco2( 3, 3) Pisc icoll2( 4, 4) COPE PODA2( 3, 1) Asel lida2( 4, 2) Ephe meril2( 2, 1)  
Sial idae2( 2, 1) Seri cost2( 3, 3) Hali plid2( 5, 2) Gyri nida2( 2, 0) Cori xida2( 5, 0) Spha erid2( 5, 5)  
Lynn aeidl( 5, 2) Plan orb2( 5, 2)

POSITIVE PREFERENTIALS  
Poly cent1( 1, 4) Lept ocer1( 1, 6) Glos soso1( 0, 4) Poly cent2( 1, 4) Lept ocer2( 1, 6) Glos soso2( 0, 4)  
\*\*\*\*\*

DIVISION 8 (N= 3) I.E. GROUP \*000  
Eigenvalue .262 at iteration 2  
INDICATORS, together with their SIGN  
Plan arili(+)

NEGATIVE PREFERENTIALS  
Pisc icoll( 1, 0) COPE PODA1( 1, 1) Asel lida1( 1, 1) Gamm aridl( 1, 1) Ephe mere1( 1, 0) Limn eph1( 1, 1)  
Lept ocer1( 1, 0) Elmi nth1( 1, 1) Cori xidal( 1, 1) Noto nect1( 1, 0) Chir oncm1( 1, 1) Phys idae1( 1, 0)  
HYDR ACAR1( 1, 1) Pisc icoll2( 1, 0) COPE PODA2( 1, 1) Asel lida2( 1, 1) Gamm arid2( 1, 1) Ephe mere2( 1, 0)  
Limn eph1( 1, 1) Lept ocer2( 1, 0) Elmi nth2( 1, 1) Cori xida2( 1, 1) Noto nect2( 1, 0) Chir oncm2( 1, 1)  
Phys idae2( 1, 0) HYDR ACAR2( 1, 1)

POSITIVE PREFERENTIALS  
Plan arili1( 0, 2) Glos siph1( 0, 2) Erpo bdel1( 0, 2) Halo didal( 0, 1) Spha erid1( 0, 2) Plan arili2( 0, 2)  
Glos siph2( 0, 2) Erpo bdel2( 0, 2) Halo didal2( 0, 1) Spha erid2( 0, 2)  
\*\*\*\*\*

DIVISION 9 (N= 20) I.E. GROUP \*001  
Eigenvalue .093 at iteration 2  
INDICATORS, together with their SIGN  
Elmi nth1(+) COPE PODA1(+)

NEGATIVE PREFERENTIALS  
Sial idael( 6, 1) Limn eph1( 3, 0) Unio nidal( 5, 0) Sial idae2( 6, 1) Limn eph1( 3, 0) Unio nida2( 5, 0)

POSITIVE PREFERENTIALS  
COPE PODA1( 0, 5) Elmi nth1( 0, 7) Hydr ophil( 1, 5) Hydr omet1( 1, 2) Neri tidal( 0, 2) COPE PODA2( 0, 5)  
Elmi nth2( 0, 7) Hydr ophil2( 1, 5) Hydr omet2( 1, 2) Neri tida2( 0, 2)  
\*\*\*\*\*

DIVISION 10 (N= 2) I.E. GROUP \*010  
Eigenvalue .137 at iteration 0  
INDICATORS, together with their SIGN  
COPE PODA1(+)

NEGATIVE PREFERENTIALS  
Phry gane1( 1, 0) Succ ineif1( 1, 0) Phry gane2( 1, 0) Succ ine12( 1, 0)

POSITIVE PREFERENTIALS  
COPE PODA1( 0, 1) Ephe meril1( 0, 1) Hydr opay1( 0, 1) Gyri nidal( 0, 1) Noto nect1( 0, 1) Tipu lida1( 0, 1)  
Musc idae1( 0, 1) Spha erid1( 0, 1) COPE PODA2( 0, 1) Ephe meril2( 0, 1) Hydr opsy2( 0, 1) Gyri nida2( 0, 1)  
Noto nect2( 0, 1) Tipu lida2( 0, 1) Musc idae2( 0, 1) Spha erid2( 0, 1)  
\*\*\*\*\*

DIVISION 11 (N= 2) I.E. GROUP \*011  
Eigenvalue .203 at iteration 0  
INDICATORS, together with their SIGN  
Coen agril(-)

NEGATIVE PREFERENTIALS  
Coen agril1( 1, 0) Mola nnid1( 1, 0) Dytli scid1( 1, 0) Gyri nidal( 1, 0) Hydr omet1( 1, 0) Gerr idae1( 1, 0)  
Noto nect1( 1, 0) Cera topo1( 1, 0) Musc idae1( 1, 0) Coen agril2( 1, 0) Mola nnid2( 1, 0) Dytli scid2( 1, 0)  
Gyri nida2( 1, 0) Hydr omet2( 1, 0) Gerr idae2( 1, 0) Noto nect2( 1, 0) Cera topo2( 1, 0) Musc idae2( 1, 0)

POSITIVE PREFERENTIALS  
Agri idae1( 0, 1) Lept oph1( 0, 1) Ephe meril1( 0, 1) Limn eph1( 0, 1) Simu liid1( 0, 1) Agri idae2( 0, 1)  
Lept oph2( 0, 1) Ephe meril2( 0, 1) Limn eph2( 0, 1) Simu liid2( 0, 1)

### Appendix 3.3.1

DIVISION 12 (N= 5) I.E. GROUP \*100

Eigenvalue .212 at iteration 1  
 INDICATORS, together with their SIGN  
 Caen idael(+)

#### NEGATIVE PREFERENTIALS

Aasel lidal( 1, 2) Dryo pida( 1, 0) Gerr idae( 1, 1) Aasel lidas2( 1, 2) Dryo pida2( 1, 0) Gerr idae2( 1, 1)

#### POSITIVE PREFERENTIALS

Plan ariil( 0, 3) Pisc icoll( 0, 1) Glos siph1( 0, 3) Erpo bde1( 0, 3) OSTR ACOD1( 0, 1) CLAD OCER1( 0, 2)  
 COPE PODA1( 0, 1) Leuc trid1( 0, 2) Lept ophill( 0, 3) Caen idael( 0, 4) Poly cent1( 0, 3) Hydr opsay1( 0, 2)  
 Hydr optil( 0, 3) Lept ocer1( 0, 1) Goer idae( 0, 2) Elmi nth1( 0, 4) Hali plid1( 0, 3) Hydr ophill( 0, 1)  
 Noto nect1( 0, 3) Tipu lidal( 0, 2) Cera topo1( 0, 3) Spha erid1( 0, 4) Hydr obili1( 0, 4) Lynn aeid1( 0, 4)  
 Phys idael( 0, 1) Plan ariil2( 0, 3) Pisc icoll2( 0, 1) Glos siph2( 0, 3) Erpo bde12( 0, 3) OSTR ACOD2( 0, 1)  
 CLAD OCER2( 0, 2) COPE PODA2( 0, 1) Leuc trid2( 0, 2) Lept ophill2( 0, 3) Caen idae2( 0, 4) Poly cent2( 0, 3)  
 Hydr opsay2( 0, 2) Hydr optil2( 0, 3) Lept ocer2( 0, 1) Goer idae2( 0, 2) Elmi nth12( 0, 4) Hali plid2( 0, 3)  
 Hydr ophi2( 0, 1) Noto nect2( 0, 3) Tipu lidas2( 0, 2) Cera topo2( 0, 3) Spha erid2( 0, 4) Hydr obili2( 0, 4)  
\*\*\*\*\*

DIVISION 13 (N= 19) I.E. GROUP \*101

Eigenvalue .104 at iteration 2  
 INDICATORS, together with their SIGN  
 Plan ariil(-) Pisc idael(-) Hydr optil(-) Limn ephil(+) COPE PODA1(+)

#### NEGATIVE PREFERENTIALS

Plan ariil( 11, 2) Ephe merel( 7, 2) Hydr optil( 5, 0) Mola nnid1( 3, 0) Veli idae( 3, 0) Noto nect1( 4, 1)  
 Simu lid1( 6, 2) Pisc idael( 6, 2) Plan ariil2( 11, 2) Ephe merel2( 7, 2) Hydr optil2( 5, 0) Mola nnid2( 3, 0)  
 Veli idae2( 3, 0) Noto nect2( 4, 1) Simu lid2( 6, 2) Pisc idae2( 8, 2)

#### POSITIVE PREFERENTIALS

OSTR ACOD1( 0, 3) COPE PODA1( 1, 4) Aasel lidal( 4, 6) Psyc homyl( 0, 3) Lepi dosta1( 0, 2) OSTR ACOD2( 0, 3)  
 COPE PODA2( 1, 4) Aasel lidas2( 4, 6) Psyc homyl2( 0, 3) Lepi dosta2( 0, 2)

\*\*\*\*\*

DIVISION 14 (N= 5) I.E. GROUP \*110

Eigenvalue .194 at iteration 1  
 INDICATORS, together with their SIGN  
 Dend rocol(-)

#### NEGATIVE PREFERENTIALS

Dend rocol( 3, 0) Ephe merel( 3, 1) Poly cent1( 1, 0) Psyc homyl( 1, 0) Hydr opsayl( 3, 1) Hydr optil( 2, 0)  
 Lept ocer1( 1, 0) Lepi dosta1( 1, 0) Gyri nida( 2, 0) Chry some1( 1, 0) Noto nect1( 1, 0) Valv stidi( 1, 0)  
 Dend roco2( 3, 0) Ephe merel2( 3, 1) Poly cent2( 1, 0) Psyc homyl2( 1, 0) Hydr opsay2( 3, 1) Hydr optil2( 2, 0)  
 Lept ocer2( 1, 0) Lepi dosta2( 1, 0) Gyri nida2( 2, 0) Chry some2( 1, 0) Noto nect2( 1, 0) Valv stidi2( 1, 0)

#### POSITIVE PREFERENTIALS

NEMA TODA1( 0, 1) NEMA TOMO1( 0, 1) Lept ophill( 0, 2) Seri cost1( 1, 2) Helo didal( 0, 2) Cera topo1( 0, 1)  
 Psyc hodil( 0, 1) Taba nida1( 0, 1) Musc idae1( 0, 1) NEMA TODA2( 0, 1) NEMA TOMO2( 0, 1) Lept ophill2( 0, 2)  
 Seri cost2( 1, 2) Helo dida2( 0, 2) Cera topo2( 0, 1) Psyc hodil2( 0, 1) Taba nida2( 0, 1) Musc idae2( 0, 1)  
\*\*\*\*\*

DIVISION 15 (N= 10) I.E. GROUP \*111

Eigenvalue .142 at iteration 1  
 INDICATORS, together with their SIGN  
 Glos siph1(-)

#### NEGATIVE PREFERENTIALS

Pisc icoll( 4, 0) Glos siph1( 7, 0) Aasel lidal( 2, 0) Lept opn1( 6, 1) Cera topo1( 3, 0) Lynn aeid1( 2, 0)  
 Plan orbil( 2, 0) Pisc icoll2( 4, 0) Glos siph2( 7, 0) Aasel lidas2( 2, 0) Lept ophill( 6, 1) Cera topo2( 3, 0)  
 Lynn aeid2( 2, 0) Plan orbil2( 2, 0)

#### POSITIVE PREFERENTIALS

Dend rocol( 1, 2) Leuc trid1( 0, 1) Ephe meril( 0, 1) Sial idael( 0, 1) Poly cent1( 1, 3) Lepi dosta( 0, 1)  
 Seri cost1( 1, 2) Hali plid1( 1, 1) Psyc hodil( 0, 1) Dend roco2( 1, 2) Leuc trid2( 0, 1) Ephe meril2( 0, 1)  
 Sial idae2( 0, 1) Poly cent2( 1, 3) Lepi dosta2( 0, 1) Seri cost2( 1, 2) Hali plid2( 1, 1) Psyc hodil2( 0, 1)

### Appendix 3.3.2 LOI classes B and C

TWINSPAN "indicator" taxa and negative (left-hand groups) and positive (right-hand groups) preferential taxa from the classification of B and C class samples (Section 5.8.2, Figures 23 and 24). N = the number of samples in the division; numbers after preferential taxa indicate the number of samples in the negative and positive groups in which these taxa occur.

**DIVISION 1 (N= 88) I.E. GROUP \***  
Eigenvalue .197 at iteration 2  
INDICATORS, together with their SIGN  
Elmi nthil(-) Tipu lidal(-) Valv stid1(+) Coen agrill(+) Limn ephill(-) Simu liidl(-) Hydr opsy1(-)  
CLAD OCER1(+)

**NEGATIVE PREFERENTIALS**  
Lept ophill( 10, 0) Ephe merel( 12, 3) Sial idael( 12, 3) Rhya coph1( 15, 1) Hydr opsy1( 25, 1) Limn ephill( 28, 0)  
Lept ocer1( 17, 6) Goer idael( 13, 0) Elmi nthil( 46, 13) Tipu lidal( 39, 8) Simu liidl( 31, 2) Ancy lidal( 15, 0)

**POSITIVE PREFERENTIALS**  
OSTR ACOD1( 4, 14) CLAD OCER1( 1, 19) COPE PODA1( 7, 17) Coen agrill( 3, 27) Valv stid1( 1, 25) Phys idael( 5, 18)  
\*\*\*\*\*

**DIVISION 2 (N= 48) I.E. GROUP \*0**  
Eigenvalue .129 at iteration 1  
INDICATORS, together with their SIGN  
Rhya coph1(-) Lept ocer1(+) Plan orbil(+) Ancy lidal(-) Plan ariil(-) Pisc icoll(-) Caen idael(+)  
Limn ephill(-) Asel lidal(+) Cori xidal(+) Limn ephill( 19, 9) Goer idael( 9, 4)  
Glos sosol( 5, 0) Ancy lidal( 13, 2) Plan ariil2( 21, 10) Pisc icoll2( 11, 1) Ephe mer2( 9, 3) Rhya coph2( 15, 0)  
Limn ephill2( 19, 9) Goer idae2( 9, 4) Glos sosol2( 5, 0) Ancy lidae2( 13, 2)

**POSITIVE PREFERENTIALS**  
Asel lidal( 6, 16) Lept ophill( 3, 7) Caen idael( 10, 20) Sial idael( 4, 8) Lept ocer1( 2, 15) Cori xidal( 7, 15)  
Cera topo1( 2, 9) Phys idael( 0, 5) Plan orbil( 6, 19) Asel lidae2( 6, 16) Lept ophill2( 3, 7) Caen idae2( 10, 20)  
Sial idae2( 4, 8) Lept ocer2( 2, 15) Cori xida2( 7, 15) Cera topo2( 2, 9) Phys idae2( 0, 5) Plan orbil2( 6, 19)  
\*\*\*\*\*

**DIVISION 3 (N= 40) I.E. GROUP \*1**  
Eigenvalue .097 at iteration 3  
INDICATORS, together with their SIGN  
Dend rocol(-) Noto nect1(-) Elmi nthil(-) PORI FERA1(-)

**NEGATIVE PREFERENTIALS**  
PORI FERA1( 3, 0) Dend rocol( 5, 3) Poly cent1( 2, 2) Lept ocer1( 2, 4) Pyra lidal( 2, 0) Elmi nthil( 6, 7)  
Noto nect1( 5, 2) Simu liidl( 2, 0) Musc idael( 3, 2) PORI FERA2( 3, 0) Dend roco2( 5, 3) Poly cent2( 2, 2)  
Lept ocer2( 2, 4) Pyra lidae2( 2, 0) Elmi nthil2( 6, 7) Noto nect2( 5, 2) Simu liid2( 2, 0) Musc idae2( 3, 2)

**POSITIVE PREFERENTIALS**  
Pisc icoll( 1, 13) Pisc icoll2( 1, 13)  
\*\*\*\*\*

**DIVISION 4 (N= 24) I.E. GROUP \*00**  
Eigenvalue .140 at iteration 2  
INDICATORS, together with their SIGN  
OSTR ACOD1(-) Nemo uridl(-)

**NEGATIVE PREFERENTIALS**  
OSTR ACOD1( 1, 1) COPE PODA1( 1, 3) Nemo uridl( 1, 1) Lept ophill( 1, 2) OSTR ACOD2( 1, 1) COPE PODA2( 1, 3)  
Nemo urid2( 1, 1) Lept ophill2( 1, 2)

**POSITIVE PREFERENTIALS**  
Plan ariil( 0, 21) Pisc icoll( 0, 11) Glos siphil( 0, 20) Erpo bdeil( 0, 16) Asel lidal( 0, 6) Ephe merel( 0, 9)  
Caen idael( 0, 10) Rhya coph1( 0, 15) Hydr opsy1( 0, 14) Hydr optil( 0, 11) Goer idael( 0, 9) Glos sosol( 0, 5)  
Elmi nthil( 0, 22) Hali plid1( 0, 14) Cori xidal( 0, 7) Spha erid1( 0, 13) Hydr obil1( 0, 19) Lymn aeidl( 0, 15)  
Plan orbil( 0, 6) Ancy lidal( 0, 13) HYDR ACARI( 0, 18) Plan ariil2( 0, 21) Pisc icoll2( 0, 11) Glos siph2( 0, 20)  
Erpo bdeil2( 0, 16) Asel lidae2( 0, 6) Ephe mer2( 0, 9) Caen idae2( 0, 10) Rhya coph2( 0, 15) Hydr opsy2( 0, 14)  
Hydr optil2( 0, 11) Goer idae2( 0, 9) Glos sosol2( 0, 5) Elmi nthil2( 0, 22) Hali plid2( 0, 22) Cori xida2( 0, 7)  
Spha erid2( 0, 13) Hydr obil2( 0, 19) Lymn aeid2( 0, 15) Plan orbil2( 0, 6) Ancy lidae2( 0, 13) HYDR ACAR2( 0, 18)  
\*\*\*\*\*

**DIVISION 5 (N= 24) I.E. GROUP \*01**  
Eigenvalue .144 at iteration 1  
INDICATORS, together with their SIGN  
Phys idael(+) Plan ariil(+) Hydr optil(+) Poly centil(+) Dend rocol(+) Asel lidal(+) Limn ephill(-) Cori xidal(-)  
Glos sosol(-) Elmi nthil(-) Hali plid(-) Spha erid(-) Hydr obil(-) Lymn aeidl(-) Hydr opsy(-) Erpo bdeil(-)

**NEGATIVE PREFERENTIALS**  
Lept ophill( 7, 0) Taba nidal( 4, 0) Lept ophill2( 7, 0) Taba nida2( 4, 0)

**POSITIVE PREFERENTIALS**  
Plan ariil( 4, 6) Dend rocol( 1, 3) OSTR ACOD1( 0, 2) Sial idael( 4, 4) Poly centil( 0, 3) Psyc homyl( 0, 2)  
Hydr optil( 1, 5) Goer idael( 1, 3) Musc idael( 2, 2) Phys idael( 0, 5) Plan ariil2( 4, 6) Dend roco2( 1, 3)  
OSTR ACOD2( 0, 2) Sial idae2( 4, 4) Poly cent2( 0, 3) Psyc homy2( 0, 2) Hydr optil2( 1, 5) Goer idae2( 1, 3)  
Musc idae2( 2, 2) Phys idae2( 0, 5)  
\*\*\*\*\*

### Appendix 3.3.2

DIVISION 6 (N= 7) I.E. GROUP \*10  
Eigenvalue .197 at iteration 1  
INDICATORS, together with their SIGN  
Coen agril(+)

NEGATIVE PREFERENTIALS  
Pisc icoll( 1, 0) Erpo bdell( 2, 2) OSTR ACOD1( 2, 1) COPE PODA1( 1, 1) Ephe mere1( 1, 0) Poly cent1( 2, 0)  
Hydr optil( 2, 2) Lept ocerl( 1, 1) Tipu lidal( 1, 0) Simu liid1( 2, 0) Cera topol( 1, 1) Musc idae1( 2, 1)  
Lynn aeidl( 2, 2) Pisc icoll2( 1, 0) Erpo bdell2( 2, 2) OSTR ACOD2( 2, 1) COPE PODA2( 1, 1) Ephe mere2( 1, 0)  
Poly cent2( 2, 0) Hydr optil2( 2, 2) Lept ocer2( 1, 1) Tipu lida2( 1, 0) Simu liid2( 2, 0) Cera top2( 1, 1)  
Musc idae2( 2, 1) Lynn aeidl2( 2, 2)

POSITIVE PREFERENTIALS  
Coen agril( 0, 5) Caen idae1( 0, 3) Pyra lida1( 0, 2) Elmi nthil( 1, 5) Noto nect1( 0, 5) Spha erid1( 1, 5)  
Hydr obili( 1, 5) Coen agril2( 0, 5) Caen idae2( 0, 3) Pyra lida2( 0, 2) Elmi nthil2( 1, 5) Noto nect2( 0, 5)  
Spha erid2( 1, 5) Hydr obil2( 1, 5)  
\*\*\*\*\*

DIVISION 7 (N= 33) I.E. GROUP \*11  
Eigenvalue .105 at iteration 2  
INDICATORS, together with their SIGN  
Spha erid1(-) Plan aril(-) Valv atidl(-) Plan orbil(-) Gamm erid1(-) Phys idae1(-) Glos siph1(-)  
Hydr obili(-) Caen idae1(-)

NEGATIVE PREFERENTIALS  
Plan aril1( 21, 2) Pisc icoll( 12, 1) Glos siph1( 19, 3) Caen idae1( 19, 3) Hydr optil( 9, 1) Spha erid1( 17, 0)  
Valv atidl( 19, 1) Phys idae1( 12, 1) Plan orbil( 25, 4) Plan aril2( 21, 2) Pisc icoll2( 12, 1) Glos siph2( 19, 3)  
Caen idae2( 19, 3) Hydr optil2( 9, 1) Spha erid2( 17, 0) Valv atidl2( 19, 1) Phys idae2( 12, 1) Plan orbil2( 25, 4)

POSITIVE PREFERENTIALS  
Ephe-mare1(-0,-2)-Poly-cent1(-0,-2)-Gerr-idael(-1,-2)-Cera-topol(-7,-5)-Ephe-mere2(-0,-2)-Poly-cent2(-0,-2)-  
Gerr idae2( 1, 2) Cera top2( 7, 5)  
\*\*\*\*\*

DIVISION 8 (N= 1) I.E. GROUP \*000  
DIVISION FAILS - There are too few items  
\*\*\*\*\*

DIVISION 9 (N= 23) I.E. GROUP \*001  
Eigenvalue .125 at iteration 3  
INDICATORS, together with their SIGN  
Ancy lidal(+) Ase1 lidal(+) Spha erid1(+) Ephe mere1(+)

NEGATIVE PREFERENTIALS  
Glos sosol( 4, 1) Glos soso2( 4, 1)

POSITIVE PREFERENTIALS  
Dend rocol( 0, 4) Pisc icoll( 3, 8) Erpo bdell( 5, 11) Ase1 lidal( 0, 6) Ephe mere1( 0, 9) Sial idae1( 0, 4)  
Cori xida1( 1, 6) Spha erid1( 3, 10) Plan orbil( 1, 5) Ancy lidal( 2, 11) Dend roco2( 0, 4) Pisc icoll2( 3, 6)  
Erpo bdell2( 5, 11) Ase1 lida2( 0, 6) Ephe mere2( 0, 9) Sial idae2( 0, 4) Cori xida2( 1, 6) Spha erid2( 3, 10)  
Plan orbil2( 1, 5) Ancy lida2( 2, 11)  
\*\*\*\*\*

DIVISION 10 (N= 18) I.E. GROUP \*010  
Eigenvalue .126 at iteration 2  
INDICATORS, together with their SIGN  
Simu liid1(+) Cori xidal(-) Cera topol(+) Spha erid1(-) Lept ophi1(-) Hydr opay1(-) Lynn aeidl(+)  
Musc idae1(+) Unio nida1(+)

NEGATIVE PREFERENTIALS  
Plan aril1( 4, 0) Lept ophi1( 7, 0) Ephe mere1( 3, 0) Sial idae1( 4, 0) Hydr opay1( 8, 1) Cori xida1( 10, 1)  
Psyche hodil( 3, 0) Spha erid1( 12, 2) Plan aril2( 4, 0) Lept ophi2( 7, 0) Ephe mere2( 3, 0) Sial idae2( 4, 0)  
Hydr opay2( 8, 1) Cori xida2( 10, 1) Psyche hodil2( 3, 0) Spha erid2( 12, 2)  
\*\*\*\*\*

POSITIVE PREFERENTIALS  
Simu liid1( 5, 5) Cera topol( 3, 4) Musc idae1( 0, 2) Unio nida1( 0, 2) Lynn aeidl( 5, 4) Simu liid2( 5, 5)  
Cera top2( 3, 4) Musc idae2( 0, 2) Unio nida2( 0, 2) Lynn aeidl2( 5, 4)  
\*\*\*\*\*

DIVISION 11 (N= 6) I.E. GROUP \*011  
Eigenvalue .211 at iteration 1  
INDICATORS, together with their SIGN  
Pisc icoll(+)

NEGATIVE PREFERENTIALS  
Baet idae1( 4, 0) Poly cent1( 3, 0) Psyche homyl( 2, 0) Goer idae1( 3, 0) Hali plidi( 3, 0) Cori xida1( 4, 0)  
Cera top1( 2, 0) Baet idae2( 4, 0) Poly cent2( 3, 0) Psyche homyl2( 2, 0) Goer idae2( 3, 0) Hali plidi2( 3, 0)  
Cori xida2( 4, 0) Cera topo2( 2, 0)

POSITIVE PREFERENTIALS  
Dend rocol( 2, 1) Pisc icoll( 0, 1) OSTR ACOD1( 1, 1) Coen agril( 0, 1) Hydr opay1( 1, 1) Lynn ephil( 2, 1)  
Helo didal( 0, 1) Psyche hodil( 0, 1) Stra tiom1( 0, 1) Musc idae1( 1, 1) Valv atidl( 0, 1) Lynn aeidl( 2, 1)  
Dend roco2( 2, 1) Pisc icoll2( 0, 1) OSTR ACOD2( 1, 1) Coen agril2( 0, 1) Hydr opay2( 1, 1) Lynn ephil2( 2, 1)  
Helo dida2( 0, 1) Psyche hodil2( 0, 1) Stra tiom2( 0, 1) Musc idae2( 1, 1) Valv atid2( 0, 1) Lynn aeidl2( 2, 1)  
\*\*\*\*\*

Appendix 3.3.2

DIVISION 12 (N= 2) I.E. GROUP \*100

Eigenvalue .221 at iteration 0  
 INDICATORS, together with their SIGN  
 PORI FERA1(+)

NEGATIVE PREFERENTIALS

CLAD OCER1( 1, 0) Ephe mere1( 1, 0) Elmi nthil( 1, 0) Tipu lidal( 1, 0) Valv atid1( 1, 0) CLAD OCER2( 1, 0)  
 Ephe mere2( 1, 0) Elmi nthil2( 1, 0) Tipu lidal2( 1, 0) Valv atid2( 1, 0)

POSITIVE PREFERENTIALS

PORI FERA1( 0, 1) Pisc icoll( 0, 1) COPE PODA1( 0, 1) Lept ocer1( 0, 1) Cera topo1( 0, 1) Spha erid1( 0, 1)  
 Hydr obil1( 0, 1) PORI FERA2( 0, 1) Pisc icoll2( 0, 1) COPE PODA2( 0, 1) Lept ocer2( 0, 1) Cera topo2( 0, 1)  
 Spha erid2( 0, 1) Hydr obil12( 0, 1)

DIVISION 13 (N= 5) I.E. GROUP \*101

Eigenvalue .226 at iteration 1  
 INDICATORS, together with their SIGN  
 PORI FERA1(-)

NEGATIVE PREFERENTIALS

PORI FERA1( 2, 0) Dend rocol1( 2, 1) Caen idae1( 2, 1) Hydr opsy1( 1, 0) Lept ocer1( 1, 0) Musc idae1( 1, 0)  
 PORI FERA2( 2, 0) Dend roco2( 2, 1) Caen idae2( 2, 1) Hydr opsy2( 1, 0) Lept ocer2( 1, 0) Musc idae2( 1, 0)

POSITIVE PREFERENTIALS

Plan arill( 1, 3) Glos siph1( 1, 3) OSTR ACOD1( 0, 1) COPE PODA1( 0, 1) Asel lidal1( 1, 3) Coro phid1( 0, 1)  
 Pyra lidal1( 0, 2) Helo didal1( 0, 1) Cera topo1( 0, 1) Neri tida1( 0, 1) Valv atid1( 1, 3) Lynn aeidl1( 0, 2)  
 Phys idael1( 0, 3) Plan arill2( 1, 3) Glos siph2( 1, 3) OSTR ACOD2( 0, 1) COPE PODA2( 0, 1) Asel lidal2( 1, 3)  
 Coro phid2( 0, 1) Pyra lidal2( 0, 2) Helo didal2( 0, 1) Cera topo2( 0, 1) Neri tida2( 0, 1) Valv atid2( 1, 3)  
 Lynn aeidl2( 0, 2) Phys idae2( 0, 3)

DIVISION 14 (N= 25) I.E.-GROUP- \*110

Eigenvalue .096 at iteration 2  
 INDICATORS, together with their SIGN  
 Pisc icoll(+) HYDR ACAR1(-) Glos siph1(-) CLAD OCER1(+) Valv atid1(+) OSTR ACOD1(-) Elmi nthil(-)

NEGATIVE PREFERENTIALS

OSTR ACOD1( 8, 1) Elmi nthil( 5, 0) HYDR ACAR1( 12, 2) OSTR ACOD2( 8, 1) Elmi nthil2( 5, 0) HYDR ACAR2( 12, 2)

POSITIVE PREFERENTIALS

Pisc icoll( 3, 9) CLAD OCER1( 4, 7) Tipu lidal( 2, 4) Unio nidal( 0, 3) Pisc icoll2( 3, 9) CLAD OCER2( 4, 7)  
 Tipu lidal2( 2, 4) Unio nida2( 0, 3)

DIVISION 15 (N= 6) I.E. GROUP \*111

Eigenvalue .298 at iteration 2  
 INDICATORS, together with their SIGN  
 Lynn aeidl(-) Plan orbil(-)

NEGATIVE PREFERENTIALS

Plan arill( 2, 0) Glos siph1( 2, 1) Elmi nthil( 2, 0) Hydr ophil( 1, 0) Meso velil( 1, 0) Hydr cmetl( 1, 0)  
 Gerr idae1( 2, 0) Tipu lidal( 1, 0) Taba nidal( 1, 0) Musc idae1( 1, 0) Valv atid1( 1, 0) Lynn aeidl( 3, 1)  
 Phys idae1( 1, 0) Plan orbil( 3, 1) Plan arill2( 2, 0) Glos siph2( 2, 1) Elmi nthil2( 2, 0) Hydr cphi2( 1, 0)  
 Meso velil2( 1, 0) Hydr cmet2( 1, 0) Gerr idae2( 2, 0) Tipu lidal2( 1, 0) Taba nida2( 1, 0) Musc idae2( 1, 0)  
 Valv atid2( 1, 0) Lynn aeidl2( 3, 1) Phys idae2( 1, 0) Plan orbil2( 3, 1)

POSITIVE PREFERENTIALS

Erpo bdel1( 0, 3) OSTR ACOD1( 0, 2) Ephe mere1( 0, 2) Caen idae1( 0, 3) Poly cent1( 0, 2) Erpo bdel2( 0, 3)  
 OSTR ACOD2( 0, 2) Ephe mere2( 0, 2) Caen idae2( 0, 3) Poly cent2( 0, 2)